Supplemental Report to the Bioregional Assessment of Northwest Forests
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Supplemental Report to the Bioregional Assessment of Northwest Forests

U.S. Department of Agriculture, Forest Service
Pacific Northwest Region
Portland, Oregon
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Vallejo, California
2021
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Introduction

This Supplemental Report to the Bioregional Assessment of Northwest Forests (report) is a companion document to the Bioregional Assessment of Northwest Forests (BioA). The USDA Forest Service Pacific Southwest Region and the Pacific Northwest Region collectively prepared a bioregional assessment of the expanded Northwest Forest Plan (NWFP) amendment area. The BioA was developed to look and feel different from standard Forest Service planning assessments. It uses plain language and an abundance of graphics to convey current conditions and trends in a relatively brief format. It also highlights the most important findings of the underlying assessment that produced the BioA, but not all findings.

This report was developed to provide additional detail about the BioA and the wealth of information considered by its authors. It includes issues that did not rise to the level of urgency to be included in the BioA but are important for land managers to consider when national forests, grasslands, and scenic areas begin preassessment and assessment tasks. It also provides additional details related to the issues included in the BioA.

As with the BioA, the focus of this report is on the 17 units that are wholly or partially covered by the NWFP amendment and other planning frameworks, as well as two additional adjacent units (the BioA area). This effort evaluates successes, challenges, trends, and potential management changes in the BioA area. Neither this report nor the BioA makes any land management decisions.

Call-out box 1: Northwest Forest Plan amendment

The 1994 Northwest Forest Plan amended the land management plans of the national forests in the range of the northern spotted owl in the Pacific Northwest region of the United States. The amendment was developed in response to mounting public concern and legal battles that halted timber harvesting in old forests throughout the owl’s range. Approval of the amendment allowed timber management to continue with new operating restrictions, while providing for the management of habitat for northern spotted owls, marbled murrelets, and other species associated with old forests and protection of aquatic habitats. However, the goal of maintaining a viable timber industry to sustain rural communities and economies was not fully realized under the NWFP.

The Forest Service’s Pacific Northwest and Pacific Southwest Regions manage the national forests and grasslands in the BioA area across Washington, Oregon, and California (map 1). A regional approach to modernizing the land management plans across the BioA area’s broad landscape provides an opportunity to understand the unique contributions from, as well as challenges and opportunities within, each individual national forest and grassland. The modernization effort is focused primarily on national forests and grasslands; to assess ecological and social connections across the landscape, we also considered some other federal and non-federal lands.
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Map 1. Bioregional Assessment boundary

This map includes the BioA boundary, Northwest Forest Plan boundary, and the related 19 national forests and grasslands. The BioA area is larger than the Northwest Forest Plan area; it includes the entire national forests and two additional adjacent units.
Introduction

As natural resource values and ecological, social, and economic conditions have shifted across time, so have the policies that guide Forest Service land management. The U.S. Forest Service 2012 Planning Rule emphasizes an adaptive planning process and resilience to climate change impacts. However, existing land management plans that have not yet been revised are not always consistent with current policies or evolving science and often do not reflect current and anticipated conditions.

Since 1994, when most land management plans in the BioA area were adopted and amended (figure 1), national forests and grasslands have experienced an increase in uncharacteristic wildfires, in the number of species listed under the U.S. Endangered Species Act, and in the spread of invasive species. Climate change, likely the biggest challenge in modern land management, is expected to result in broad-scale ecosystem changes throughout the BioA area. In the past 25 years, the Forest Service has gained more knowledge about the watersheds in the area as well as the terrestrial and aquatic species that depend on the habitats in the BioA area. The agency continues to build relationships with, and to better understand, the communities and American Indian tribes (tribes) served by the national forests and grasslands in the BioA area.

Figure 1. Bioregional Assessment area Land Management Plans

Land management plans in the BioA area are more than 25 years old. There have been changes in social, ecological, and economic conditions as well as in resource demands; and new scientific information and policy are available.
Introduction

This report is grounded in science and was written by an interdisciplinary team of specialists from Regions 5 and 6. It is a snapshot of the challenges and opportunities associated with today’s resource management but does not include the level of detail required for formal land management plan modernization. The team relied on existing data gathered from several sources, including the 2018 Synthesis of Science to Inform Land Management Within the Northwest Forest Plan Area, the 2014 Science Synthesis to Support Socioecological Resilience in the Sierra Nevada and Southern Cascade Range, the Northeastern California Plateaus Bioregion Science Synthesis (draft, anticipated March 2020), and more than 20 years of NWFP monitoring and implementation experience, two science reviews, as well as information gathered through ongoing Sierra Nevada Framework, PacFish¹ and InFish² effectiveness monitoring programs. In addition, the team used scientific literature from professional journals and technical bulletins, and information gathered from Forest Service personnel who have intimate knowledge of the area. This report is based on the best science and information available at the time it was written; no new analyses or studies were completed.

This report focuses on issues that can be resolved through the land management planning process and does not consider other management constraints, such as budget and staffing concerns. This report does not include all important planning issues, such as lands and realty, cultural resources, grazing and minerals. It does not include specific planning components and is intentionally nonprescriptive. Forest-, project-, and site-specific topics are not discussed in this document; these will be collaboratively developed with partners and stakeholders as the planning process continues. Upcoming land management planning efforts will include new analysis; development or revision of plan components; engagement with stakeholders, tribes, and local governments; and environmental review as required by the National Environmental Policy Act and other laws, regulations, and policies.

The assessment process identifies information gaps that will be addressed in future planning efforts to help consider uncertainties and knowledge gaps related to fire, climate change, invasive species, tradeoffs between ecosystem and species goals, and tradeoffs between ecological and social components. There were some essential geographic information system datasets that contained knowledge gaps. For example, critical habitat maps were not available for southern Oregon and northern California coho salmon, a species listed as threatened under the U.S. Endangered Species Act.

Organization of the Supplemental Report

This report has four chapters: (1) Serving People, (2) Tribal Rights and Interests, (3) Caring for the Land, and (4) Northwest Forest Plan Land Use Allocations and Management Direction. Each chapter includes background and context along with information under the following section and subsection headings:

What is Working Well
This section highlights what has been working well under the existing plans and suggests that some guidance and direction should be retained as we move through the modernization process.

Key Change Issues
This section identifies landscape management challenges that cannot effectively be addressed under our current plan direction for a variety of reasons.

Planning Considerations
This subsection includes references to the recommendations from the BioA that reflect the key change issue. Because the BioA recommendations are integrated, multiple recommendations may be connected to one key change issue. In addition, this section may include planning considerations that were not explored in the BioA but could be considered as plan modernization moves forward.

Geographic Considerations
This subsection generally provides more detail than the BioA on geographic considerations and may inform upcoming plan modernization efforts by listing groups of forests or individual forests where key change issues occur.
Chapter 1: Serving People

“To sustain the health, diversity, and productivity of the Nation’s forests and grasslands to meet the needs of present and future generations.” – U.S. Forest Service mission statement.

Introduction

The Forest Service is dedicated to achieving quality land management while providing for sustainable multiple-use management to meet the diverse needs of people. In this chapter we examine key forest benefits and highlight their social and economic values and contributions to the communities we serve. We will further update information related to direct economic benefits to communities with more detailed and site-specific local data during future planning phases. Here, we look back at what’s been working well for local communities and where communities have struggled to adapt under current land management plans on the national forests and grasslands in the BioA area. We acknowledge limitations and consider what we can do to improve how we serve communities into the future. More information about the benefits that national forests and grasslands provide to local communities is detailed throughout each chapter of this report.

Benefits to People and Communities

National forests and grasslands provide clean air and water, and habitat for plants and animals; they also preserve cultural resources, conserve natural settings and provide essential commodities and recreational opportunities for the benefit of present and future generations. These benefits (also referred to as ecosystem services) deliver significant value for all Americans. Some benefits, including cultural heritage and biodiversity, may be more difficult to connect directly to monetary values than others, such as timber and water. However, nonmonetary benefits greatly contribute to improving the quality of people’s lives. Recognition of the immense value national forests and grasslands generate through this wide range of benefits was included in the original vision of the Northwest Forest Plan Amendment (NWFP), which was to provide “a balanced and comprehensive strategy for the conservation and management of forest ecosystems, while maximizing economic and social benefits from forests” (USDA FS and USDI BLM 1993: E-5).

In 2016, activities on federal lands in the NWFP area supported almost 25,000 jobs in local communities (Grinspoon et al., n.d.). Specifically, recreational opportunities, timber activities, and agency employment on National Forest System and Bureau of Land Management lands supported the most jobs as a result of forest activities in these communities (refer to figure 2). Over time, there have been reductions in agency employment and recent increases in employment related to timber harvest on federal lands. Direct jobs are those supported by forest activities (for example, timber harvesting) and

3 Some data and information presented in this chapter are taken from the Northwest Forest Plan—the first 25 years (1994–2018): socioeconomic monitoring results report. The 25-year monitoring report is currently in development and is not finalized as of the writing of the Bioregional Assessment of Northwest Forests and this report. Therefore, the findings presented here are the best available at this time, but the information may change as the 25-year monitoring report is reviewed and finalized.
indirect jobs are supported by subsequent business-to-business transactions that support these forest activities (for example, spending on materials, equipment, and fuel for forest work). There is a relatively high number of jobs supported by recreation on public lands. (Grinspoon et al., n.d.).

Job opportunities on national forests and grasslands outside the NWFP area are important as well. For example, in 2016, the Ochoco National Forest supported almost 800 jobs with the majority resulting from agency employment, grazing and forest products. The Lassen National Forest supported around 1,400 jobs in local communities, mostly in agency employment, forest products, and recreation. Employment opportunities resulting from activities on the Modoc National Forest totaled around 1,000 jobs focused on grazing and agency employment.

![Figure 2. Jobs supported by Forest Service and Bureau of Land Management activities](image)

**Figure 2. Jobs supported by Forest Service and Bureau of Land Management activities**

*Employment supported by Forest Service and Bureau of Land Management programs from 2001, 2012, and 2016. Adapted from Grinspoon et al., n.d.*

**Sustainable Recreation**

The recreational opportunities available in the BioA area are numerous and include biking, hiking, off-road vehicles, hunting, fishing, wildlife viewing and water sports to name a few. These activities provide enjoyment for millions of visitors, improve physical and mental health and allow people to establish a connection to the outdoors. Recreational activities are not evenly distributed throughout the NWFP area. Some national forests and grasslands provide a wider variety of opportunities and are closer to population centers than others. As a result, some communities have transitioned to a more recreational- and amenity-based economy, while others have not (Charnley et al. 2018). In addition to these benefits, visitors to the area also contribute to local economies by spending around $600 million annually on lodging, restaurants, souvenirs, and other trip-related expenses (Charnley et al. 2018). In 2016, this spending by recreation visitors supported around 7,800 jobs in these areas.

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4 No estimates were available for the Crooked River National Grassland.
communities. These recreation-based jobs are critical to supporting the social and economic conditions in many communities throughout the BioA area (Grinspoon et al., n.d.). Therefore, maintaining and enhancing the quality of recreational experiences on national forests and grasslands is important to supporting visitation and sustaining these socioeconomic contributions to communities.

Forest Products
Land management on national forests and grasslands contributes to supporting forest product jobs in local communities (Charnley et al. 2018, White et al. 2015). The Forest Service hires and trains local contractors to perform forest restoration activities, which provides employment opportunities in many rural areas where such opportunities might otherwise be limited (Nielsen-Pincus and Moseley 2010). Total employment in forest products industries in the NWFP area, including logging and primary and secondary wood manufacturing, has declined over the past 25 years. However, even as harvest on all land ownerships in the area has increased slightly since 2009, there has not been a similar increase in jobs as a result of this increased activity across all ownerships (figure 3).

Overall, timber harvest declined dramatically from 2006 through 2009 (leading up to and through the great recession) with most of the reduction in harvest occurring on non-federal lands. The resulting decline in employment from this period has not experienced an increase because of recent increases in harvest, which may have to do with changes in industry structure and improvements in mill efficiencies (Grinspoon et al., n.d.). These job losses are likely to have a more pronounced effect in small rural communities where up to 10 percent of employment in the community can be in the forest products manufacturing sector (Grinspoon et al. 2016).

Figure 3. Northwest Forest Plan (NWFP) timber harvest and employment from 1995 to 2016
Timber-related employment (logging and mill processing) and timber harvest on all ownerships in the NWFP area, 1995–2016. MMBF = million board feet. Adapted from Grinspoon et al., n.d.
Despite the overall reduction in traditional timber sector jobs in rural communities in the NWFP area, there is potential for restoration activities to generate employment opportunities (Charnley et al. 2018, Nechodom et al. 2008, White et al. 2015). Contractors from local communities can be hired to perform restoration activities, which is important because employment opportunities in these areas are often limited. However, restoration contracting in some areas has transitioned to lower skilled jobs and can favor mobile businesses that employ a high proportion of temporary and migrant laborers, and therefore may not provide wages that replace former timber industry jobs (Charnley et al. 2018).

A study examining forest and watershed restoration work found that about 16 to 24 jobs are supported for each $1 million invested in forest restoration activities (Nielsen-Pincus and Moseley 2010). This range is dependent on the type of restoration activities. Another study estimated that upland forest work supports about 10 jobs for each $1 million invested (White et al. 2015). These job estimates are only for the noncommercial work, where timber is not extracted for sale or use, and any jobs that follow from commercial timber harvest and mill processing in support of these forest restoration projects would be in addition to those stated.

Investments in labor-intensive activities (such as site preparation, tree and shrub planting, and cutting small trees and brush by hand) support a greater number of jobs, whereas investments in highly technical and equipment-intensive activities (such as forest thinning, small-diameter and selective logging, masticating ground fuels, constructing stream habitat features, and excavating floodplain and wetland features) support fewer jobs. Increased pace and scale of restoration leads to employment and economic activity beyond the effects of employment generated by the activities themselves, including the need for materials and equipment purchased from suppliers and restoration workers spending their paychecks for goods and services.

National forests and grasslands in the BioA area also provide a wide variety of nontimber forest products (non-timber forest products), such as moss, mushrooms, cones, grasses, and firewood. These products provide valuable economic and cultural benefits to rural and urban households through their use, harvest, and processing.

Due to the diverse range of products harvested, estimating the true economic contribution of all forest products is difficult. In addition, collection of many special forest products is for subsistence or personal consumption and therefore market transactions do not capture the economic value of these forest products. The retail value of non-timber forest products in the United States is estimated to be at least $1.4 billion, with much of that coming from the NWFP region (Charnley et al. 2018). In 2012, 99 percent of the value of special forest product permits in the NWFP area included seven categories: foliage, fruits and berries, fuelwood, grass, limbs/boughs, mushrooms, and Christmas trees (Grinspoon et al., n.d.).

**Water**

Water provides tremendous value to people and communities for municipal, industrial, and agricultural uses. National forest system lands supply more water in the Western United States than any other landowner (Brown and et al. 2008, Luce et al. 2017) In fact, about 49 percent of the water in the Western United States comes from national forests and grasslands (Brown et al. 2016). Water has a monetary value of more than $3.7 billion
nationally, with the highest values in Oregon, Washington, and California (Sedell et al. 2000). As highlighted in the Bull Run Watershed example, the actual value of this water is immense as communities depend on a reliable water supply for their subsistence and economic growth. National forest and grassland water also provides valuable ecological benefit supporting terrestrial and aquatic species as well as opportunities for recreational activities.

Call-out box 2. National forests and grasslands as a water supplier—Bull Run example

Providing a sustainable flow of clean water is a foundational role of national forests and grasslands and increasingly important given growing populations and climate change projections that reflect changing precipitation patterns. One example in the Bioregional Assessment of Northwest Forests (BioA) is the Bull Run Watershed, the primary drinking water supply for the City of Portland and its 20 wholesale customers. Water from the Bull Run serves more than 950,000 residents in the Portland metropolitan region, which is a quarter of Oregon’s population.

Located 26 miles from downtown Portland in the Sandy River basin on the Mount Hood National Forest, the 102-square-mile watershed collects rainwater and snowmelt that then flows to the Bull Run River and its tributaries. The river drains into two reservoirs that store more than 17 billion gallons of water. The watershed has been managed under increasing levels of protection since it was established as a forest reserve in 1892. Ninety six percent of the lands within the Bull Run Watershed Management Unit are under federal management; the rest are owned by the City of Portland.

National Forest Road System

Roads and motorized transportation on roads in and around the national forests and grasslands provide an essential public service through access for recreation, fire suppression, timber sales, and other contracted work. Local communities also use some forest roads for daily commuting. Construction of Forest Service roads generally is paid for by the revenue from timber sales and maintenance occurs through timber and stewardship sale funds. Revenue from these sources has been in decline over the past three decades as timber sale volumes and cut values have declined. This threatens the sustainability of forest road systems and the benefits they provide.

Livestock Grazing

Livestock grazing provides important public benefits and supports employment and the ranching culture in local communities, particularly in the eastern portion of the BioA area on the Modoc, Ochoco and Fremont-Winema National Forests and the Crooked River National Grassland. Livestock grazing opportunities in these areas are critical to ranchers for summer range because of the quality of forage and cooler summer temperatures at higher elevations. Current and projected climate change and related increased frequency, intensity, and scale of wildfires pose challenges for sustaining grazing practices on uplands and threaten the benefits that livestock grazing on National Forest System lands delivers to ranching communities.
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Special Use Services
The Forest Service special uses program authorizes land uses that benefit the public and protect public and natural resource values. Each year, the Forest Service receives individual and business applications for activities and facilities associated with water transmission, agriculture, outfitting and guiding, recreation, telecommunications and other utilities, research, photography, and video productions. The Forest Service reviews each application to determine how the request could affect public use of national forests and grasslands.

In the BioA area, the special use activities that are currently generating the most revenues for national forests and grasslands are recreational residences, winter recreation resorts, powerlines, outfitting and guide services, and marinas. In 2018, special uses in the bioregion generated $19 million for national forests and grasslands (USDA FS 2019). Revenue from special uses has been increasing over time, thus showing the increasing importance of these activities for both providing public services and generating forest revenues.

Air Quality
National forests and grasslands contribute to improved air quality and better visibility from reduced ozone loss and less particulate matter. People view fresh air and good visibility as being healthy, so these factors contribute to an improved sense of wellbeing in the people who encounter them. Good visibility and fresh clean air further enhance peoples’ desire to engage in spiritual and cultural activities and recreate on national forests and grasslands, which in turn positively affects recreation-associated jobs and local economies.

Sustaining These Benefits to Influence Community Wellbeing
Management actions that maintain forest health, diversity, and productivity are critical to ensuring sustainable forest benefits to people and communities. It is important for planning to effectively support these management actions by setting a vision for how the forest should be managed and establishing guidance for how these activities should be undertaken. Some forest management actions and their benefits to people and communities include the following:

- implementing restoration projects that improve ecological integrity and resilience to fire and drought, accelerate the development of late successional forest characteristics, and reduce wildfire risk to communities.
- providing forest products to local mills, which supports the local workforce and contributes to the local and regional economies.
- offering access to recreational opportunities and facilities that attract visitors from the local area as well as around the region, country, and world.
- providing clean water and air to promote the good health and wellbeing of residents and visitors.
- restoring habitat for sensitive species, which contributes to biodiversity and supports activities, such as plant gathering, wildlife viewing, hunting, and fishing.
• restoring functional watersheds, riparian areas, and water bodies to provide sustainable water, vegetation, and recreational opportunities.
• restoring populations of threatened and endangered anadromous fish.
• establishing fire-resilient forest conditions along transmission corridors and around towers to meet community energy and communications needs.
• offering livestock grazing, which contributes to local economies through jobs and income, as well as to local cultures, by supporting a community’s history, its sense of place, and traditional and cultural uses of the land.

Achieving large-scale success in these types of forest management activities is enhanced by communities that actively support the efforts. Many of the jobs supported by national forests and grasslands are in rural areas and these jobs often represent a significant contribution to local economic development and social sustainability.

Communities, Wood Processing Infrastructure, and Restoration
Over the past 20 years, timber-related infrastructure and workforce have been declining across the BioA area, and wood processing infrastructure east of the Cascade Range has become especially sparse. Limited infrastructure results in higher transportation costs because timber needs to be hauled farther for processing. The lower concentration of sawmills and a lack of local workforce make implementing forest restoration activities particularly challenging in the eastern part of the BioA area.

Planning at the land management plan level provides the Forest Service an opportunity to determine broad strategies to encourage investment in local mill facilities and the workforce that would be needed to successfully implement timber sales and restoration to improve the ecological integrity of forests and support the socioeconomic health of communities.

Communities are all different and as a result, there has been no one common community experience over the past few decades. Some communities have been able to adapt to changes; these include amenity communities that have taken advantage of natural forest settings to benefit from an influx of residents and an increase in visitors interested in recreation opportunities on the national forests and grasslands. Some communities have also successfully developed economic opportunities outside of the traditional forest timber sector by pursuing energy, agriculture, and tourism (Charnley et al. 2018). In contrast, other communities have faced significant challenges because of a variety of economic, social, geographical, and institutional factors as well as changes in federal land management. Restructuring in the forest products industry, improved mill efficiencies, competition for products nationally and internationally, and fluctuations of harvesting on non-federal lands have all affected the timber-based economy (Charnley et al. 2018).

Environmental Justice and Vulnerable Communities
Many communities throughout the BioA area have experienced shifts in population demography. Some communities have seen an increase in minority populations while others experienced an increase in low-income residents. These shifts bring new social challenges that interact with our land management decisions. The 1994 Executive Order on Environmental Justice and the U.S. Forest Service 2012 Planning Rule direct land managers to pay attention to how policy changes or program implementation may affect
vulnerable sectors of communities, such as minority, low-income, elderly, and disabled populations, who may experience the effects of land management changes differently than less vulnerable community members. The planning rule requires the agency to consider that communities are diverse and some community subgroups may require unique outreach to ensure effective communication.

Vulnerable populations are increasing in rural counties in the BioA area, and poverty is increasing or remaining high in these same areas. These rural communities have historic links to federal land management and may be subject to the greatest socioeconomic effects if changes are made to the existing land management plans. The values and views of these vulnerable populations are often underrepresented in planning processes; therefore, outreach efforts must account for the needs of these communities.

Smaller communities in remote locations with a heavy dependence on the timber economy and on a federal supply of timber to support their workforce and infrastructure are directly affected by reductions in federal harvests (Charnley et al. 2006, 2018). These types of communities are found throughout the BioA area.

Figure 4 illustrates the strength of county economic and social ties to federal forest activities in 1990, prior to the NWFP era. The year 1990 is a mid-point between a period of very high harvest activity, as measured by total volume on federal forests in the 1980s, and the adoption of the NWFP in 1994. Counties in southern Oregon and northern California were more likely to be strongly linked to federal forest land management in 1990 than counties in other areas of the NWFP region. Factors determining strength of links to federal forest land management in 1990 include employment in land management agencies (Forest Service and Bureau of Land Management), revenue sharing payments from the federal government to counties linked to federal timber harvests, share of county’s total national forest and grassland land area subject to management under the NWFP, share of timber milled within the county from forests and grassland, and private sector jobs in forestry and wood products manufacturing relative to all jobs in 1990.

Individual communities can also be described in terms of strength of links to national forests and grasslands, but community-scale data are not available for this kind of analysis. Communities may not have had the same strength of links to federal forest lands management in 1990 as did the county in which they were located, though the proportion of communities within a county that were strongly linked to federal forest lands management is probably higher in counties with strong links as shown by darker green shades on the map (figure 4). Counties or communities that were extremely strongly linked to federal forest land management may have been more likely to experience challenging economic conditions in the early 1990s related to reduced timber harvests on federal lands immediately before and during the NWFP era (Adams, n.d.).
Figure 4. County ties to federal forest activities in 1990

This figure displays the strength of county economic and social ties to federal forest activities in 1990, before the Northwest Forest Plan era. Adapted from Adams, n.d.
The following community stories provide examples of experiences in forest communities in the BioA area. These communities were all historically linked to national forests and grasslands and had to adapt to management changes that were introduced by the NWFP. These brief stories have been adapted from the comprehensive formal analysis of communities in the forthcoming 25-year monitoring report (Coughlan et al., n.d.).

**Happy Camp, California**
Happy Camp is a small mountain community surrounded by the Klamath National Forest in northern California. The town’s timber economy, based almost exclusively on national forest harvests, boomed between the 1950s and 1980s, after which, national forest harvests were sharply curtailed. The last sawmill in the area shut down in 1994. Happy Camp and the surrounding area lost 22 percent of their population during the 1990s. This dramatic change created a void in the community as younger working-class families left to pursue other opportunities. A way of life (working in the woods) that had defined the community was lost and has not returned.

Today, the town is home to the Klamath-Siskiyou Art Center, Karuk Tribe administrative offices that provide important services to the community, and the Forest Service’s Happy Camp/Oak Knoll Ranger District offices that have remained open but at lower staffing levels. The community has also worked with the Forest Service to try to expand sustainable recreation visitation in the two wilderness areas adjacent to the community as well as on the Klamath River, which offers excellent opportunities for rafting and fishing.

The community persists, but its population, economy, and social structure have changed dramatically over the past 30 years.

**Leavenworth, Washington**
Leavenworth is a small mountain town surrounded by the Okanogan-Wenatchee National Forest in central Washington. Nearly 6,000 people lived in Leavenworth by 1920, and the town once supported a large sawmill. However, the timber boom ended when the railroad was rerouted in 1926, and soon afterward the mill closed. Although many locals continued in the timber industry at nearby mills or served in jobs at the Leavenworth Ranger District, the town population steadily declined until the 1960s.

In 1963, community leaders decided to create a Bavarian-themed tourist town to boost the economy. Leavenworth’s economy steadily improved during the latter 20th century. Tourism continued to grow as mountain biking, rock climbing, and rafting on the Okanogan-Wenatchee National Forest became popular. In addition, the forest’s beautiful scenery and amenities attracted new permanent and part-time residents that spurred burgeoning real estate and vacation rental markets. Though Leavenworth is now prosperous, neither it nor nearby communities in the Wenatchee Valley are associated with timber sector work anymore, a fact that some residents lament. The lack of forestry sector infrastructure in the area makes needed restoration work on the national forest challenging.
**Mill City, Oregon**

Mill City is at the mouth of the North Santiam Canyon, 30 miles east of the City of Salem. Between the late 1950s and the late 1980s, Mill City thrived. High-wage jobs with the town’s lumber mills and the Forest Service sustained a variety of local businesses, a community theatre, and a bowling alley.

Unlike Happy Camp and Leavenworth, the timber industry did not disappear from Mill City. At least one wood products mill has operated in the North Santiam Canyon area throughout the NWFP era. A few small outdoor recreation businesses are in operation, and the area has attracted some retirees. Yet the town still experienced significant social and economic change starting in the 1990s. Because of the elimination of Forest Service positions funded by timber receipts many employees and their families left town. Many local logging contractors either folded or moved away. Houses were left vacant and the number of absentee property owners grew. Highly transitory residents, often with few job skills or prospects, moved to the area to take advantage of cheap housing, bringing additional needs for social services with them. School enrollment declined substantially, reflecting far fewer families with children in the community.

Mill City’s population did not decline sharply, unlike many other rural forest towns in the Northwest. In 2017, it had roughly the same number of residents as in 1990, but it is a much different community with fewer services and jobs.

**Land Management Plans Can Help Sustain Community Benefits**

"Participants would like plan revision to balance local social values and economic considerations (including tourism, recreation, and timber) with environmental concerns and forest health." — From 2015 Forest Listening Session for NWFP modernization page 7.

It is important for planning to support new approaches to forest management that can better sustain the benefits provided by national forests and grasslands. These forest benefits are important in influencing economic conditions and the quality of life in communities across the BioA area. Smaller more rural forest communities can be more susceptible to changes in these benefits, leading to impacts on day-to-day life. Therefore, in order to reinforce needed social and economic sustainability throughout the BioA area, planning can accomplish the following:

- **Emphasize the value of social and economic benefits of national forests and grasslands**—Planning can highlight the potential for national forests and grasslands to provide a multitude of social and economic benefits to people and local communities and include language that supports forest management decisions that sustain these important benefits.

- **Improve relationships**—Planning should foster outreach activities to develop mutual understanding on management needs and local community concerns and should support improved information sharing with stakeholders.

- **Increase community partnerships and collaboration**—Planning should encourage collaboration with states, counties, and local communities to achieve compatible planning goals and facilitate synergy between federal, state, and local fiscal and staff resources.
Recognize forest contributions to local community plans—Planning can contribute to effective stewardship of recreational opportunities and sustainable, healthy, and fire-resilient forest landscapes that are, to the extent possible, compatible with local community needs and plans.

Studies suggest that improving public involvement can reduce conflict, improve public buy-in, increase compliance with agency regulations, and remove barriers to project implementation (Cerveny et al. 2018, Koontz 1999, Stern 2008, Whitall 2007). Forest plans should emphasize the need for these types of actions and establish the planning process as a key phase for building consensus among stakeholders with different sets of values on how landscapes should be managed. Planning endeavors to meet as many concerns of stakeholders as possible, but given the diversity of social and economic values, not all needs can be met. To that end, planning can recognize the necessity for an ongoing conversation on public values in order to enhance the understanding of the forest benefits that are important to people and communities in the BioA area.

While peoples’ values and social constraints transcend planning, they can hamper planning efforts and strengthen the conflicts that restrain management options. For example, values around old trees and old-growth forest range widely and create challenges to creating broadly accepted ecologically appropriate definitions of old forest that are needed for planning and implementation purposes. Therefore, improving working relationships, building trust and engaging stakeholders and partners is critically important to the success of planning. Improving the understanding of how people connect to the national forests and grasslands strengthens the relationship between the agency and communities and contributes toward building trust. Forest planning efforts that incorporate these connections and acknowledge the diversity of peoples’ values, attitudes, and beliefs can be better equipped to understand the interconnections of social and ecological systems and better anticipate future needs for change (Cerveny et al. 2018).

Sustainable Recreation
Most people experience their national forests and grasslands through recreational activities. Recreational opportunities and settings in the BioA area provide people with multiple benefits, including improved physical health and fitness; conservation of open space and the environment; educational opportunities about the values of conservation, land stewardship, and responsible recreation; and making nearby communities more desirable places to live. Recreation contributes to social and economic sustainability and provides opportunities to connect people with nature. Diverse topographies, landscapes, water features, vegetation, fish, wildlife, and histories make national forests and grasslands a valued outdoor playground.
Rivers provide highly valued settings and opportunities for healthy, active outdoor activities; and they support or diversify local economies and provide needed access to open space and the outdoors.

To provide and enhance recreation benefits, national forests and grasslands establish partnerships with private entities and volunteer-based and nonprofit organizations; these partnerships are mutually beneficial agreements that build capacity to complete essential resource work and benefit communities. As part of land management planning, the Forest Service conducts recreation assessments that include information about existing conditions, trends in visitor uses and preferences, sustainability of recreation settings and opportunities, access, and scenic character. To complete these assessments, land managers use several resources, including a standardized protocol for inventorying the location, condition, and use type, and monitoring resource conditions, as well as planning tools such as the Recreation Opportunity Spectrum, Limits of Acceptable Change, the Scenery Management System, and Visitor Impact Management frameworks. Newer efforts to understand recreation use on public lands include participatory mapping, place-based planning, human ecology mapping, and developing approaches to better understand user values, attitudes, and beliefs. Recreation assessments help managers determine the extent to which a planning area meets the demand for recreation opportunities and the ability of a planning area to sustain recreation settings, opportunities, access, and scenic character, which contributes to better informed management decisions.
What is Working Well

**What is Working Well 1—Designated Areas**

Designated areas are public lands that receive special protections and management provisions, in recognition of their significant and often unique qualities or values. There are numerous designated areas in the BioA area, including wilderness, scenic and free-flowing rivers, volcanic landscapes, and popular recreation areas. Designated areas can have a wide range of nationally significant qualities that are scenic, geologic, cultural, historic, or ecological. They also may contain outstanding and often unique recreation opportunities, delivering important benefits to the American public, including clean water, biodiversity, and opportunities for adventure or solitude.

Congressional legislation establishes some designations and others occur through agency administrative processes. Some designations have been the source of controversy, and public sentiment varies about the best way to manage designated areas. Regardless, federal land management agencies are charged with the stewardship of these special places, so they may be enjoyed by current and future generations.

Highlights of the Forest Service’s designated areas network include the following:

- Three national monuments (two of which are volcanic monuments)
- Six national recreation areas
- Three scenic areas, each with a unique emphasis
- Portions of two national scenic trails and three national historic trails
- More than 1,000 miles of All-American Road
- 5.5 million acres of wilderness across 74 designated units
- Approximately 1,700 miles of wild and scenic river corridor along 204 designated river segments
- Nearly 100 research natural areas
- Three national natural landmarks and approximately 60 special interest areas
Photo 2: Berryessa Snow Mountain National Monument, Mendocino National Forest

The monument’s terrain and topography comprise a strong diversity of habitat types that support a variety of plant and wildlife species. In the higher elevation Snow Mountain area, the biological diversity is among the richest in California.

Photo 3. Historic Mott Bridge on the Umpqua National Forest

The bridge represents many combined special area features; it is located on the designated Wild and Scenic North Umpqua River, was built by the Civilian Conservation Corps in 1935, and is a stop on the Rogue-Umpqua National Scenic Byway on the North Umpqua Highway.
Map 2. Nationally Designated Areas within the Bioregional Assessment area

These areas can have a wide range of nationally significant qualities, including scenic, geologic, cultural, historic, or ecological. They also may contain outstanding and often unique recreation opportunities, delivering important economic and health benefits to the American public.
Map 3. Nationally designated trails and roads in the Bioregional Assessment area

These travel routes recognize a variety of scenic, historic, and recreation values and often cross multiple local, state, and federal jurisdictions, requiring collaborative management across boundaries. Volunteer organizations, user groups, and chambers of commerce also play a key role in stewarding these iconic routes that celebrate our past and America’s natural beauty.
Map 4. Designated wilderness in the Bioregional Assessment area

These areas provide social, cultural, economic, scientific, and ecological benefits for present and future generations. Many of America’s iconic landscapes include wilderness areas that provide outstanding opportunities for solitude and primitive and unconfined recreation. BLM = Bureau of Land Management, FWS = U.S. Fish and Wildlife Service, NPS = National Parks Service, USFS = U.S. Forest Service.
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Map 5. Designated wild and scenic rivers in the Bioregional Assessment area

These select rivers preserve outstanding natural, cultural, and recreational values in a free-flowing condition. They also provide ecological services and values such as clean water, flood irrigation, and fish and wildlife habitat, and they serve an important role in the global water cycle. Many rivers cross national forest boundaries and can include multiple ownerships requiring consistent management and coordination.
What is Working Well 2—Major Supplier of Recreation
The Forest Service is one of the largest suppliers of outdoor recreation in the BioA area. The agency’s role in recreation management includes providing for recreation experiences and activities for a broad range of users, as well as administration of special use permits that provide additional experiences for the recreating public and expand capacity for recreation on National Forest System lands. Recreation use is expected to increase, and recreation user demographics and the types of uses are changing, although deferred maintenance for recreation facilities and the transportation system may affect the quality of recreation experiences that the Forest Service provides. National Visitor Use Monitoring results indicate that satisfaction with the recreation experience on each forest is quite high.

What is Working Well 3—National Forest Road System
The national forests and grasslands have an extensive road and trail network that provides opportunities to connect people with nature. Forest transportation systems are essential infrastructure, providing access to the national forests and grasslands for public benefit, permitted uses, agency administrative activity, and traditional and tribal harvesting of forest products. The National Forest Road System also provides key access for fire suppression and search and rescue operations. Impacts and tradeoffs associated with the road system are discussed in the “Key Change Issues” sections in “Recreation,” “Sustainable Timber and Forest Products,” and “Aquatics, Fish and Water.”

Call-out box 3. National scenic trails and associations

The Pacific Crest Trail (PCT) was designated as a national scenic trail in the National Trails System Act of 1968, and stretches from the Mexican border to the Canadian border, covering 2,650 miles and crossing through several national forests within the BioA area. Each year hundreds of thousands of visitors set foot on at least one section of the PCT, whether it is for a day, weekend, or a weeks-long excursion.

The Pacific Crest Trail Association (PCTA) partners with the Forest Service and other land management agencies on a variety of trail stewardship activities along the trail corridor, including maintenance, monitoring, and visitor use management. The PCTA also engages in advocacy and visitor education. In 2014, 1,600 volunteers and staff completed maintenance work on about 1,700 miles of the PCT.

The Pacific Northwest National Scenic Trail (PNT) was designated as a National Scenic Trail in 2009, though the concept of the trail has existed since the late 1970s. The PNT forms a continuous, 1,200-mile path from the Pacific Ocean near Cape Alava to the Continental Divide in Glacier National Park, and many of the PNT segments and rights-of-way existed before the trail’s formal designation. Congress tasked the Forest Service with management of the PNT, and the trail crosses through three national forests within the Bioregional Assessment area: the Okanogan-Wenatchee, Mt. Baker-Snoqualmie, and Olympic.

The Pacific Northwest Trail Association (PNTA) works with the Forest Service and other land management agencies on trail construction, maintenance, monitoring, and volunteer coordination. In 2017, trail crews provided 11,800 hours of volunteer labor along the length of the trail.
What is Working Well 4—Sustainable Recreation, Special Uses, and Economic Growth

Recreation visitor spending is an important source of economic activity associated with Bureau of Land Management and Forest Service management in the BioA area. Visitors spend money on lodging, restaurants, souvenirs, and other trip-related expenses. Recent estimates indicate that visitors to the NWFP area spend about $612 million each year (Charnley et al. 2018). Recreational visitor spending can provide critical economic contributions in forest communities that were adversely affected by changes in the forest products industry. Recreational opportunities on the forest also provide tremendous benefits to the visitors themselves by way of the enjoyment and experiences they have on the forests.

Special uses provide important opportunities and benefits for recreation, energy, communications, and infrastructure. These activities allow access and opportunities on public lands in a way that provides a vast array of social and economic benefits to people and communities. Special uses generate critical forest revenues, $19 million across the bioregion in 2018, and this revenue has nearly doubled since 2010. Special uses are increasingly important in providing services that are valuable to the public.

Key Change Issues

Key Change Issue 1—Aquatic Strategies and Late-Successional Reserves

The aquatic management strategies implemented within the BioA area provide for clean water, healthy vegetation, and improved fishery resources that support recreational and economic benefits. This contributes to the broad suite of public benefits found within the BioA area and demonstrates the relationship between people, nature, and the continued need to maintain and restore healthy ecosystems.

Within the NWFP area, forest recreation managers have observed conflicts between maintaining existing recreation facilities (including roads and trails) and meeting the NWFP’s Reserve and Aquatic Conservation Strategy objectives. The NWFP requires adjusting recreation areas where they conflict with Aquatic Conservation Strategy objectives; if adjustments are not feasible, then NWFP objectives require eliminating or removing the conflicting use.5

Forest recreation managers encounter challenges when addressing “saturated” use areas along river corridors because proposed additional facilities are generally not considered “neutral/beneficial” for meeting Aquatic Conservation Strategy objectives. Similar conflicts occur in late-successional reserves, where a primary emphasis is the protection of old-growth-dependent species, such as northern spotted owls and marbled murrelets; existing recreation facilities and uses often come into conflict with species protection. Mitigation measures that would make facilities consistent with the NWFP can be challenging to implement and consulting regulatory agencies would prefer to see conflicting uses eliminated.

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5 See RM-2, pg. C-34 in Standards and guidelines for management of habitat for late-successional and old-growth forest related species within the range of the northern spotted owl (USDA and USDI 1994).
Sustainable recreation challenges and needs include (1) maintaining existing recreation facilities in NWFP reserves, (2) expanding existing sites, (3) creating new recreation sites in areas near water/stream corridors where needed to relieve existing/increasing pressure at existing, similarly sited recreation areas, and (4) achieving Aquatic Conservation Strategy objectives through appropriately designed and managed recreation facilities that meet both recreation and aquatic objectives.

Planning Considerations

Refer to BioA Chapter 2 Management Recommendation

Recommendation 10: Recognize the social and economic benefits to communities and people from sustainable recreation opportunities.

- Develop forest plan components to achieve sustainable recreation management objectives in riparian settings without compromising meeting Aquatic Conservation Strategy and late-successional reserve objectives at the appropriate scale.

Geographic Considerations

National forests and grasslands located near metropolitan areas are most likely to face challenges related to increasing recreational use, deteriorating natural resource conditions within riparian reserve corridors as a result of overcrowding/overuse, and the need to develop new recreational opportunities for growing and increasingly diverse populations. BioA national forests and grasslands can be grouped into three tiers: tier 1 (high visitation) = Columbia River Gorge National Scenic Area and Deschutes, Gifford Pinchot, Mt. Baker-Snoqualmie, and Mt. Hood National Forests; tier 2 (moderate visitation) = Okanogan-Wenatchee, Shasta-Trinity, Siuslaw, and Willamette National Forests; and tier 3 (low visitation) = Fremont-Winema, Klamath, Lassen, Mendocino, Modoc, Ochoco, Olympic, Rogue River-Siskiyou, Six Rivers, and Umpqua National Forests.

Key Change Issue 2—Need to Create Single, Cohesive Planning Document

Existing land management plans are specific to individual forests and grasslands and provide varying levels of plan direction for sustainable recreation-related resources. In contrast to the NWFP’s uniform guidance for natural resource management, there is no overall cohesion and consistency to recreation management across national forests and grasslands within the BioA area. Additionally, national forests and grasslands in the BioA area must conform to layers of “plan” direction, including the NWFP and other large-scale plan amendments, species recovery plans, and critical habitat designations, which contributes to a confusing and chaotic planning framework. Congressionally designated areas (including wild and scenic rivers, national monuments, national recreation areas, and national scenic areas) often require additional comprehensive management plans. Many of these plans were approved in the 1990s, and similar to forest land management plans, they may need to be updated, revised or amended. An example is existing recreation direction that relies on outdated recreation opportunity spectrum and scenery management system inventories that do not reflect current recreation uses on the national forests and grasslands and are not responsive to the agency’s vegetation restoration challenges and strategies. Updated agency protocols for recreation opportunity spectrums and scenery management systems are available.
Planning Considerations

Refer to BioA Chapter 2 Management Recommendation

Recommendation 10: Recognize the social and economic benefits to communities and people from sustainable recreation opportunities.

- Using information gathered through public engagement, identify and develop recreation emphasis management/geographic area(s) that provide flexibility to address future development needs in terms of increased use and new recreational uses and activities.
- Review the existing trail network and its past, present, and expected use; and evaluate how this use relates to land use allocations and Aquatic Conservation Strategy objectives.
- Develop varying recreation trail and recreation site density for riparian areas with specialized habitat. Use new mapping protocols and integrate inventory findings into updated recreation opportunity spectrums and scenery management system to modernize these inventories and streamline future project implementation.
- Evaluate designated area comprehensive management plans to ensure they are consistent with the new plan components.

Geographic Considerations

The need to create a cohesive recreation strategy is applicable throughout the entire BioA area.

Key Change Issue 3—Sustainable Recreation Opportunities

Regional trends for recreation activity are similar to national trends. Nationally, participation in outdoor recreation activities on public and private lands is level to slightly declining. However, a projected increase in population will overcome the slightly decreasing or level use trends, and overall usage will increase (Bowker et al. 2012). Increasing population within the BioA area exceeds national averages, specifically the Portland and Seattle metro areas and other Northwest population centers. Higher visitation rates resulting from increased populations are expected primarily at developed and interpretive sites. Within the NWFP area, hiking, downhill skiing, and nature-related pursuits are the most common recreation activities as reported by the Forest Service’s National Visitor Use Monitoring (NVUM) program. See figure 5 and photo 4.
Figure 5. Top recreation activities on National Forest System lands within the Bioregional Assessment area

While many Forest Service programs, including timber, grazing, and energy production, generate revenue for the U.S. Treasury, ski areas regularly outperform them all.
Planning Considerations

Refer to BioA Chapter 2 Management Recommendation

Recommendation 10: Recognize the social and economic benefits to communities and people from sustainable recreation opportunities.

Expand existing and develop new partnerships to continue delivering the sustainable recreation mission of the agency; and acknowledge existing partnerships for all the critical services they provide. Develop more unified management direction for recreation-related resources to increase agency efficiency given the continued downward trends for agency capacity. Develop sustainable recreation-themed plan components that can adapt to new recreational user activities and opportunities.

Geographic Considerations

National forests and grasslands located near metropolitan areas are most likely to face challenges related to increasing recreational use, changing user demographics, diverse populations, and recreation activity related to new/emerging technologies. See figure 6.

![Figure 6. Annual visitor estimates by national forest and recreation setting](image)

*Visitation varies widely from forest to forest, and the type of visits vary. This data indicates much higher visitation on national forests and grasslands closer to metropolitan areas and the urgent need to ensure land management plans are adapted to this high level of use (data reported by the Forest Service’s National Visitor Use Monitoring program).*

**Key Change Issue 4—Access**

The Forest Service faces an extensive maintenance backlog for its road system. Recent national reporting indicates an estimated $3.2 billion maintenance backlog for fiscal year 2016 (USDA 2017). Roads naturally deteriorate over time, and when combined with an excessive maintenance backlog this reduces the overall sustainability of recreation use on national forests and grasslands throughout the BioA area. The forest trail system faces similar maintenance backlogs, which reduces recreational trail access, activities, and opportunities.
Planning Considerations

Refer to BioA Chapter 2 Management Recommendation

Recommendation 10: Recognize the social and economic benefits to communities and people from sustainable recreation opportunities.

- Forest plan components, including desired conditions, objectives, and goals, can be developed with public and stakeholder participation and input to help address the National Forest Road System while considering the tradeoffs associated with access, capacity, and broader natural resource considerations.

- Planning and revision efforts should incorporate partnerships for road and trail system maintenance such as those described in the National Forest Trail System Stewardship Act of 2016—Public Law 114-245 and should use improved analysis tools.

- National forests and grasslands should continue to evaluate motor vehicle use maps annually, which provide better consistency between travel analysis report findings and future land management planning/revision efforts.

Geographic Considerations

Figure 7 displays open road miles by forest by maintenance level\(^6\), which represents the geographic distribution of the road network in terms of mileage. Trend information presented in the NWFP 20-year monitoring report does not show “open road” mileage per forest, but rather displays mileage per maintenance level from years 1999 to 2012. The total mileage for the National Forest Road System has declined during the past 20-plus years. The 10-year monitoring report for socioeconomic monitoring results provided a summary regarding the overall reduction in the road system:

In general, the Forest Service is adding very few new miles to its road system. Road mileage on national forests and grasslands is generally decreasing; more miles of road are decommissioned than are built. Road decommissioning is ongoing and proceeds as funds become available. Level 1 and 2 road miles have increased with an associated decrease in level 3 through 5 road miles, so fewer miles are accessible to passenger cars. The increase in level 1 and 2 miles has occurred because the loss of funding from appropriated sources. The reduction in work done by timber sale operators means the agency does not have the budget to maintain as many of its roads to higher standards (Charnley 2006: 42).

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\(^6\) Maintenance Levels: ML1 – Roads that have been placed in storage between intermittent uses; ML2 - Roads open for use by high-clearance vehicles. Passenger car traffic is not a consideration; ML3 - Roads open and maintained for travel by prudent drivers in a standard passenger car. User comfort and convenience are low priorities; ML4 - Roads that provide a moderate degree of user comfort and convenience at moderate travel speeds; ML5 - Roads that have a high degree of user comfort and convenience. These roads are normally double-lane, paved facilities.
Figure 7. Open roads reported on motor vehicle use maps

The amount of open road miles by maintenance level vary widely across the Bioregional Assessment area and all are generally affected by an extensive maintenance backlog. Dedicating the necessary resources to successfully address and reduce the deferred maintenance backlog will require strategic and long-term actions.

**Key Change Issue 5—Sustainable Recreation and Economics**

National forests and grasslands face challenges in sustaining current levels of recreation services, developing partnerships to help deliver the agency’s sustainable recreation mission, and administering special use permits to provide essential community services and provide recreational opportunities. The deferred maintenance backlog for recreation related infrastructure (including recreation facilities and road and trail infrastructure) is significant and inhibits sustainability, potentially affecting recreation user experiences. The special use program also faces administration challenges, such as the capacity to review and administer applications and enforce permit restrictions to protect resources. This results in a backlog of permit proposals and increased resource damage and limits the benefits that are provided to the public.

**Planning Considerations**

*Refer to BioA Chapter 2 Management Recommendation*

Recommendation 10: Recognize the social and economic benefits to communities and people from sustainable recreation opportunities.

Changes to forest plans that address increasing forest visits would be beneficial to the sustainability of recreational opportunities and the economic and social benefits they provide. Forest plan components, including desired conditions, objectives, and goals, can be developed with public and stakeholder participation and input to support sustainable recreation objectives, to improve special use permit administration and delivery, and to balance these with associated natural resource conservation objectives.
Examples include the following:

- Provide more facilities for overnight stay (as well as multiday activities) as these types of facilities draw visitors from farther away that spend five to eight times more money in local communities than do day users.

- Expand partnerships and encourage engagement with outside agencies, partners, and volunteers to overcome financial challenges associated with deferred maintenance and developing new opportunities. Because current plans do not speak to this topic, revised plans could better highlight the need for these types of expanded partnerships.

Changes to forest plans that help address the sustainability of special use activities can contribute to the economic and social benefits these activities provide to people and communities and include the following:

- Improved plan components that align the need for special use permits with desired forest conditions (for instance, criteria 2 FSH 2709.11, sec. 12.2, which requires that permits be reviewed for consistency with plan elements).

- Design consistent approaches and plan components to address the controversy associated with existing and proposed energy utility corridors. Such approaches may include developing management areas or modifying existing NWFP land use allocations for these types of permitted uses and infrastructure, all of which will be fully explored during the formal planning process.

- Stabilize road and recreation infrastructure to provide continued access, sustainable recreation opportunities, special uses, and associated economic activity.

**Geographic Considerations**

National forests and grasslands located nearest to metropolitan areas are most likely to face challenges related to increasing recreational use and the need to develop new recreational opportunities for growing and increasingly diverse populations.

Visitation and visitor spending may be disproportionately affected in communities near areas with recreational opportunities most vulnerable to climate change and its impacts to recreational opportunities and access (see next section).

**Key Change Issue 6—Sustainable Recreation and Climate Change**

The effects of climate change to sustainable recreation activities can be broadly characterized; warmer temperatures may create more opportunities for warm-weather activities (for example, hiking, camping) and present fewer opportunities for snow-based activities (for example, skiing, snowmobiling). The recreating public often adapts to changing conditions by substituting other activities or adjusting the timing or location of activities. Federal agencies administering these affected resources are less nimble in adjusting to changing conditions, which include the following:

- Longer warm-weather recreation and a shortened winter recreation season.
- Reduced opportunities for winter activities in lower elevation zones.
- Increased susceptibility of facilities being underused or unusable due to decreased snow amounts.
- Increase crowding at sites that remain viable for winter recreation.
• Access to high-elevation zones both earlier and later compared to traditional seasons.
• Increased impacts to recreational road and trail systems, such as rutting.
• Wildfire impacts, including reduced access due to road and area closures, smoke impacts, and reduced aesthetics that may impair recreation experiences.
• Hunting/fishing opportunities affected through changes to vegetation, instream flows, and water temperatures.
• Increased temperatures may not align with regulated seasons and might reduce opportunities for cold-water species fishing.
• Reduced streamflow, reservoir levels, and hazardous algal blooms may adversely affect water-based recreation opportunities.
• Increased flooding will affect recreation infrastructure, including riparian trails, trailheads, roads, campgrounds, dispersed campsites, boat launches, bridges, and other infrastructure that supports recreation-related services and benefits.

Planning Considerations

Refer to BioA Chapter 2 Management Recommendation

Recommendation 10: Recognize the social and economic benefits to communities and people from sustainable recreation opportunities.

• Applying adaptation strategies based on key vulnerabilities caused by climate change to inform plan revision assessments that describe potential climate conditions and effects to key resources will be critical to success.
• Plan revisions may also identify and prioritize resource vulnerabilities from climate change. Plan components should include specific emphasis on desired conditions, objectives, and monitoring programs.

Geographic Considerations

Areas that are more vulnerable to climate change occur in the southern, drier portions of the BioA area and are characterized by mid-elevation sites where snow levels are projected to fluctuate and increase in elevation over time, riparian areas affected by an increased frequency of high flow events, and areas more susceptible to increasing size and severity of wildland fire that may affect forest access and reduce scenic and aesthetic values.

Sustainable Timber

Introduction

During the 1980s and 1990s, Forest Service land management plans generally emphasized timber outputs rather than desired landscape conditions. When the NWFP was adopted, the separation of timber-producing lands from the rest of the landscape was aimed at creating a compromise between stopping the loss of old-forest habitats and protecting their dependent

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7 This document does not make assumptions, draw conclusions, or make recommendations related to future timber sale viability and connected timber markets, value of wood products, or stumpage values. The factors that contribute to the value of vegetation management projects, connected treatments, and timber sale viability are highly variable or unpredictable.
species from timber harvest, while still providing predictable timber outputs from federal lands (USDA FS and USDI BLM 1994: 26). As a result, the amount of old forest on federal lands has stabilized (Davis et al. 2015, Davis et al. in progress [20- and 25-year monitoring reports]), but timber output, while relatively stable for about the past decade, has remained below anticipated levels. In addition, conflicting management direction related to northern spotted owl designated critical habitat has affected anticipated timber harvest. Harvest levels are unlikely to increase under current plans because the objectives for timber production and restoration often conflict with habitat protection objectives.

In recent years, the key challenges have changed, and Forest Service paradigms have shifted, but land management plans have not kept up. See “By Design: Isolation of Timber Production Lands; a Paradigm Shift” section in this document.

What is Working Well

**What is Working Well 1—Stable Timber Production**

Since 2005, timber production levels have remained relatively stable, producing an average of 450 million board feet (MMBF) per year from Forest Service lands within the NWFP area, although the predicted harvest of about 600 MMBF per year has not been realized. Recent harvest levels have also been regionally stable at about 72 percent of Forest Service anticipated timber production (figure 8). While we view our stable timber production rates as a success, the fact that production levels are consistently below what was anticipated in our land management plans is a concern. Although commercial harvest of timber is planned to continue, conflicting plan direction and restrictions on tree size or stand age, as well as a lack of social acceptance of planned harvest methods like regeneration, will likely limit future harvest.

Predicted Forest Service harvest levels within the NWFP is the approximate Forest Service proportional contributions to probable sale quantity (PSQ) as defined under the NWFP (see “NWFP Probable Sale Quantity Context” section below). Total PSQ estimates in the NWFP Record of Decision and Environmental Impact Statement were 1.1 billion board feet per year and included all federal lands, including Bureau of Land Management timber production. In 1993 it was estimated that although late-successional reserve volumes are not included in PSQ calculations, an additional volume of 100 to 170 MMBF might be obtained from reserve areas (Johnson et al. 1993).

Today, it is important to consider the context of the cumulative changes in what land base is available for timber production emphasis, changes in management direction, and changes in the social acceptability of planned harvest methods when analyzing how the Forest Service has managed timber outputs both within the NWFP area and programmatically within its Pacific Southwest and Pacific Northwest Regions (regions 5 and 6, respectively). These conditions are outlined in more detail in the following sections on key change issues. Over a 5-year period (2014–2018), NWFP national forests and grasslands have reach about 72 percent of Forest Service PSQ (USDA 2019b). National forests and grasslands outside the NWFP area have contributed a disproportionately large amount (given intrinsic productivity and original forest plan estimates) to timber outputs for the Northwest area. Peak timber production was in 2013 at 84 percent of Forest Service PSQ.
Figure 8. Northwest Forest Plan (NWFP) harvest levels from 1994 to 2018

Harvest levels have remained stable in the NWFP area for the past decade, although they have been less than predicted. The graph shows the amount of timber sold from national forests and grasslands in the NWFP area between 1994 and 2018 compared to the Forest Service estimated probable sale quantity. Forest Service anticipated timber volume is about 600 MMBF and the agency has produced an average of about 450 MMBF within the NWFP area annually since 2005.

Timber harvest area between 1994 and 2017 primarily has been within matrix lands (about 65 percent), and secondarily within late-successional reserves (about 19 percent). About 10 percent of harvest was in adaptive management areas (figure 9). Matrix lands are those federal lands within the NWFP that are outside the six categories of NWFP designations (congressionally reserved areas, late-successional reserves, adaptive management areas, managed late successional areas, administratively withdrawn areas, and riparian reserves). All other land use allocations contributed 1 percent or less by land area to timber production (totaling about 5 percent). Commercial thinning in historic plantations in late-successional reserves and thinning in matrix continues to contribute to timber output. Commercial thinning harvest methods continue across the BioA area, often in concert with restoration and resilience projects.
Figure 9. Recent proportion of harvest by Northwest Forest Plan (NWFP) land use allocation

*Trends in the proportion of acres of timber harvest activities by primary NWFP land use allocation. About 76 percent of all harvest acres since 1994 has been in matrix or adaptive management area lands. About 19 percent of timber harvest acres have been in late-successional reserve. All other land use allocations contributed 1 percent or less by land area to timber production (totaling about 5 percent not displayed in graphic).*

What is Working Well 2—Nontimber Forest Products

In addition to timber products, national forests and grasslands provide a variety of nontimber forest products such as moss, mushrooms, cones, grasses, and firewood. These products support community and household well-being by providing income and economic opportunities, strengthening community networks and relationships, facilitating intergenerational ecological knowledge transfer, and enabling nontimber forest product gatherers to develop stronger connections with nature and improve their mental and physical health (figure 10, photo 5 and photo 6) (Spies et al. 2018). The nominal value is just one indicator of the real value of these products harvested from public lands (Grinspoon et al., n.d.).
Chapter 1: Serving People

Figure 10. Nontimber forest product values within the Northwest Forest Plan (NWFP) area

Value of special forest products sold from National Forest System lands in the NWFP area 2002–2017. Adapted from Grinspoon et al. n.d.

Photo 5. A young harvester picking mushrooms

Finding a connection to the land by picking mushrooms.
A family outing to harvest a Christmas tree from Forest Service lands can be a cultural tradition.

Call-out box 4. Nontimber forest products

Forests in the Bioregional Assessment area also provide a wide variety of nontimber forest products (NFTP), such as berries, moss, mushrooms, cones, grasses, and firewood. These products are collected both for personal use by individuals and families, and for sale by commercial collectors. Non-timber forest products provide supplemental income and an important safety net for many households as well as providing people with a strong connection to nature. A considerable portion of the workers in the Northwest Forest Plan area are members of minority groups, and environmental justice issues related to these groups’ access to and use of nontimber forest products is important to consider in planning.

The retail value of non-timber forest products in the United States is estimated to be at least $1.4 billion, with much of that coming from the NWFP region (Charnley et al. 2018). In 2012, 99 percent of the value of special forest product permits for national forests in the BioA area was from seven categories: foliage, fruits and berries, fuelwood, grass, limbs/boughs, mushrooms, and Christmas trees (figure 11) (Grinspoon et al., n.d.). In the Pacific Northwest, roughly 20,000 people participated directly in the floral greens/bough sector or the wild mushroom sector. The most important impacts of the NWFP on nontimber forest products are likely to be landscape level changes in forest structure and composition under the plan’s provisions. This means that generally late-successional products such as matsutake mushrooms will do well, but early-seral products such as salal and berries may do less well.
What is Working Well 3—Shift in Harvest Methods

Forest Service harvest methods shifted from primarily clear-cutting in the 1980s and early 1990s to mainly commercial thinning after 1994 as we implemented more intermediate harvest treatments with multiple objectives (Spies et al. 2018). Harvest methods that retain significant structural elements of the preharvest stand largely have replaced clear-cutting (Franklin et al. 2018: 108). These harvest methods continue to create timber outputs that contribute to local economies, often in concert with restoration and resilience projects. Forest Service commercial thinning includes traditional, evenly spaced thinning and has, over the past several decades, evolved into more dynamic intermediate thinning methods, such as variable density thinning and variable retention harvest. These more variable thinning methods have been used broadly and studied regularly in the BioA area since the early 1990s (figure 11).

![Figure 11. Harvest methods used in the Bioregional Assessment area](image)

**Figure 11. Harvest methods used in the Bioregional Assessment area**

*Acres of primary harvest types (clear-cutting, commercial thinning, salvage, sanitation, and single-tree selection of trees) within the BioA area from 1944 to 2017. There were more than 75,000 acres clear-cut in 1988, while 2007 saw a peak in commercial thinning at just under 82,000 acres. Timber harvest practices have moved to mostly commercial thinning with almost no stand clear-cutting.*

Key Change Issues

**Key Change Issue 1—Timber Processing Infrastructure and Forestry Workforce**

Timber processing infrastructure and workforce have declined in recent decades (Charnley et al. 2018). Many rural communities in and around the BioA area that rely on federal timber and landscape restoration have been socially and economically affected by declines since the 1990s. Economically feasible restoration efforts, timber processing infrastructure, and having a skilled workforce are needed to support ecological integrity and benefits to communities in the region.

The ability to implement restoration treatments on federal lands is influenced by several factors, including the presence of markets, sawmill facilities, and a capable workforce (map 6). Conversely, the types of restoration treatments implemented, and products generated influence the viability of the infrastructure and the presence of a capable...
workforce. For timber production to function as a tool to meet ecological objectives, purchasers must be able to meet the requirements for minimum acceptable bid prices, use the offered wood material, and perform the required work. Therefore, it is difficult to rely on timber harvest to fund forest management and restoration on all but the most valuable timberlands.

Timber processing infrastructure and a skilled workforce are sparse in some locations, especially east of the Cascade Range. The number of mills has decreased regionally, and the average distance between large mills has grown to about 100 miles, resulting in an increased cost to transport logs to mills throughout much of the BioA area. However, remaining mills have modernized, and total wood processing capacity has increased (Charnley et al. 2018).

Call-out box 5. Mills and workforces

Maintaining a sustainable local wood products infrastructure and workforce is key to Forest Service restoration goals and its mission to provide renewable sources of timber. Local wood processing facilities and the skilled workers they employ transform restoration harvests into valuable products such as lumber, which can in turn financially support more restoration activities.

In the past 20 years, this infrastructure and workforce have declined across the BioA area; processing infrastructure east of the Cascade Range has become especially sparse. Fewer sawmills make restoration activities more of a challenge. This limited infrastructure can result in higher transportation costs and fewer buyers for the timber from these lands.

Modernizing forest plans allows the Forest Service to determine effective strategies to encourage investment in the local mill facilities and workforces needed to increase the pace and scale of restoration to improve the ecological health of forests and support the socioeconomic health of communities.
Map 6. Sawmills within the Bioregional Assessment area and restoration needs as of 2016

Vegetation management can be limited by the ability to use harvested trees and pay for landscape treatments.
Updates to land management plan direction, while unable to resolve all issues, can help to improve social and economic sustainability and better reflect the needs of local communities throughout the BioA area, especially those in hard-hit rural communities across southern Oregon and northern California. While the Forest Service strives to balance the social, economic, and ecological needs of communities and landscapes, there are instances where one resource objective might require more emphasis than others. When updating land management plans, the Forest Service collaborates with American Indian tribes, states, counties, and communities to develop goals and discover potential management approaches that seek the right emphasis on community and ecosystem objectives. For example, plan direction can facilitate increased timber generated from national forests, which can increase timber available to mills, which would increase the workforce needed to maintain a stable pace and scale of restoration. We acknowledge that there is often controversy when the best available science indicates that active management restoration is needed. We understand that conflicting values surround timber harvest and active management, and that we will need to address such issues in upcoming planning efforts.

Geographic Considerations
The number of timber processing facilities has decreased in Washington, Oregon, and northern California. Although total processing capacity has increased in Washington and remains constant in Oregon, the cost of transporting harvested timber to mills has increased in areas with limited remaining infrastructure (Charnley et al. 2018). This is particularly true east of the Cascade Range and in the southern Coast Range (map 6) where restoration needs in frequent-fire dependent ecosystems are urgent.

Key Change Issue 2—Anticipated Timber Volume Output
Similar to limited timber processing infrastructure, timber harvest below projected levels restricts our ability to achieve restoration objectives and support communities and infrastructure. Harvest levels are unlikely to increase under current plans because the objectives for timber production and restoration often conflict with habitat protection objectives. For example, timber production is no longer emphasized on much of the NWFP matrix land because large areas of matrix have been designated as critical habitat for the northern spotted owl. See the “Conflicting Management Direction” section below for more discussion.

Restoration needs in frequent-fire dependent ecosystems usually require the application of mechanical treatments, including timber harvest, often in combination with fuels reduction treatments. Restoration needs exist across the BioA area; this includes restoration needs in NWFP late-successional reserves, matrix, and other land use allocations. Often forest plan direction in these areas conflicts with the application of restoration treatments and the subsequent coproduction of timber.

One example includes commercial thinning in historic plantations within late-successional reserves. Thinning from below is generally only commercially viable every 30 to 50 years on the most productive sites and may not be commercially viable on drier sites for much longer as growth is often much slower in drier areas. Restrictions on harvesting trees that are more than 80 years old in late-successional reserves of the NWFP means that restoration that coproduces timber will be more difficult for the next couple of decades, especially on forests and districts that have a high proportion of late-successional reserves. The challenge arises
because areas available for restoration and commercial timber production will be thinned or have already been thinned within the past 30 years. This progression through time results in stands that are too old and trees that are too large to be commercially harvested under the restrictions in the land management plans. Therefore, needed restoration would likely not be accomplished due to the conflicts between forest plan direction and the application of restoration treatments.

Existing land management plans are not consistent with modern harvest methods and technology. Generally, projects that produce timber currently do not use some harvest methods included in the original plans, like regeneration harvest of old forest in matrix lands. More modern methods and concepts (developed since the 1980s), such as variable density thinning, modern logging methods, and the role of timber harvest in working toward ecological resilience and integrity, have not been incorporated into existing plans. Our current direction focuses more on setting standards and guidelines for timber harvest than on working toward desired conditions. Such a focus can prevent leveraging new technology; an example is technology that allows harvest on steep slopes with little impact to soils. Timber outputs from our highly productive fire infrequent and fire diverse (mixed severity) lands have been particularly curtailed. However, social values related to land management have begun to shift toward recognition of the broad benefits associated with our natural resources and the importance of balancing resource protection with timber production (Charnley et al. 2018).

NWFP late-successional reserves and matrix are similar in their current need for disturbance restoration, which involves mechanical treatments, including timber harvest, in combination with fire treatments (figure 12).

![Figure 12. Fire and mechanical treatment needs in the Northwest Forest Plan area](image)

**Figure 12. Fire and mechanical treatment needs in the Northwest Forest Plan area**

*Millions of acres need some combination of mechanical or fire treatment to reduce density and restore forests (disturbance restoration need) in land use allocations, including matrix, late-successional reserve, congressionally reserved and administratively withdrawn, and adaptive management areas. Mechanical treatments to achieve restoration goals can often produce commercial timber as a coproduct of improving ecological resilience.*
Planning Considerations

The main BioA recommendation for providing sustainable timber and forest products involves better integration of ecological integrity, landscape restoration, and resilient landscape management with the coproduction of timber. The need is to create 21\textsuperscript{st} century desired conditions with forest plans that emphasize ecological integrity and resilient landscapes. This includes integration of desired conditions related to ecological forestry restoration and resilience in the face of climate change.

Refer to BioA Chapter 2 Management Recommendations

Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.

Recommendation 2: Address the dynamic nature of ecosystems to be better positioned to respond to future environmental uncertainties.

Recommendation 6: Recognize that fire is a natural process and plays an important role in reducing the risk of uncharacteristic fire and in promoting ecosystem health.

Recommendation 7: Expand the use of timber harvest as a restoration tool to provide economic and social benefits to communities.

Recommendation 8: Shift from single-species management to maintaining and restoring habitat for multiple species to help ensure we are managing our ecosystem to be resilient in the face of future change.

Recommendation 9: Promote active management in plant and animal habitats to restore and promote ecological resilience.

Geographic Considerations

The land management plans on the Gifford Pinchot, Mt. Hood, Willamette, and Umpqua National Forests projected the highest timber outputs under the NWFP, and these national forests have experienced significant gaps between projected production and actual output (table 1); these are highly productive national forests and the timber output does not reflect that productivity. This is primarily due to interactions between (1) conflict between underlying forest plan management direction, inventoried roadless area designation and NWFP assumptions; (2) unforeseen complexity and unpredictability for timber production given unmapped riparian reserves, the Survey and Manage program, and northern spotted owl critical habitat; and (3) lack of social acceptability of harvest in old forest and regeneration harvest on matrix lands. Both social acceptability of timber harvest in old forest and regeneration harvest were assumed when the original PSQ calculations were made.

The Willamette, Okanogan-Wenatchee, Rogue River-Siskiyou, Umpqua, Shasta-Trinity, Gifford Pinchot, and Klamath National Forests, and to some extent the Mt. Baker-Snoqualmie and Mt. Hood National Forests, are prime examples of where the Forest Service could be simultaneously producing timber volume and meeting ecological needs but are hampered from implementing sound projects. All these national forests have high or moderate restoration needs (Ringo et al. 2019) in addition to the standing volumes available outside of congressionally reserved areas where active management could both implement
ecologically sound projects and produce timber volumes (Franklin et al. 2018, Spies et al. 2018). We acknowledge that there is often controversy when the best available science indicates that active management restoration is needed. We understand that conflicting values surround timber harvest and active management, and that we will need to address such issues in upcoming planning efforts.

Table 1. Timber volume gap by national forest
*Four national forests have the highest planned timber volume output gap: Gifford Pinchot, Mt. Hood, Willamette, and Umpqua National Forests.*

<table>
<thead>
<tr>
<th>National Forests</th>
<th>Projected Timber Volume Output Gap</th>
</tr>
</thead>
<tbody>
<tr>
<td>Okanogan-Wenatchee</td>
<td>Low</td>
</tr>
<tr>
<td>Mount Baker-Snoqualmie</td>
<td>Medium</td>
</tr>
<tr>
<td>Olympic</td>
<td>Medium</td>
</tr>
<tr>
<td>Gifford Pinchot</td>
<td>High</td>
</tr>
<tr>
<td>Columbia River Gorge</td>
<td>Low</td>
</tr>
<tr>
<td>Mt. Hood</td>
<td>High</td>
</tr>
<tr>
<td>Siuslaw</td>
<td>Medium</td>
</tr>
<tr>
<td>Willamette</td>
<td>High</td>
</tr>
<tr>
<td>Deschutes</td>
<td>Low</td>
</tr>
<tr>
<td>Ochoco</td>
<td>Low</td>
</tr>
<tr>
<td>Umpqua</td>
<td>High</td>
</tr>
<tr>
<td>Fremont-Winema</td>
<td>Low</td>
</tr>
<tr>
<td>Rogue River-Siskiyou</td>
<td>Medium</td>
</tr>
<tr>
<td>Six Rivers</td>
<td>Medium</td>
</tr>
<tr>
<td>Klamath</td>
<td>Medium</td>
</tr>
<tr>
<td>Modoc</td>
<td>Low</td>
</tr>
<tr>
<td>Lassen</td>
<td>Low</td>
</tr>
<tr>
<td>Shasta-Trinity</td>
<td>Medium</td>
</tr>
<tr>
<td>Mendocino</td>
<td>Low</td>
</tr>
</tbody>
</table>

Table 1 ranks each forest with a qualitative ranking based on a comparison of planned timber output and actual timber output. The land management plans signed in the 1980s and amended by the NWFP, PacFish, InFish, and Sierra Nevada Framework include projected timber volume outputs. Some national forests have met the projected outputs while others have not. While these volume projections are now more than 25 years old, it is important to show the gap between projected and actual outputs.

Planned timber output was either NWFP PSQ or initial forest plan estimates for national forests or portions of national forests found outside the NWFP area. National forests ranked as “high” generally have the largest volume gap between planned timber output and actual output. These national forests tend to be most naturally high in primary
production capabilities and have more timber production lands (for instance, matrix). National forests ranked as “medium” have a moderate gap between planned and actual timber output. These mid-range national forests are relatively either high or mid-level in intrinsic primary production capabilities and have a moderate level of potential for additional timber production and a moderate level of timber production lands. National forests ranked as “low” generally have a lower gap between planned and actual timber output. Often this is because these national forests have relatively low primary production capabilities, limited potential for additional levels of timber production, and little to no timber production lands.

**NWFP Probable Sale Quantity Context**

Historically, timber production that was projected has not been met (figure 13). The PSQ was defined under the NWFP as a "rough approximation" for the first decade after the decision (1995–2005), and it was acknowledged that it was unlikely that the annual PSQ estimates would be achieved during the first several years. The total PSQ for all federal lands for option 9 (selected) was 1.1 billion board feet (plus or minus 10 percent) for the first decade. “It will take time for the land management agencies to develop new timber sales that conform to the planning amendments effected by our decision. In addition, the decision contains requirements to perform various levels of analysis or survey work before awarding timber sales in certain areas.” (USDA FS and USDI BLM 1994:19). PSQ was not defined by the agency in the record of decision, but projected calculations were provided by the Forest Ecosystem Management Assessment Team (FEMAT and USDA FS 1993: table VI-2 page VI-5).

![Forest Service Timber Volume Sold within the NWFP area between 1994 and 2018 as compared with Forest Service Propable Sale Quantity](image)

**Figure 13.** Historic timber volume output in the Bioregional Assessment area

Timber volume output history measured in millions of board feet for the BioA area from 1994 to 2018. Individual forest plans calculated an allowable sale quantity and sale schedule for each forest under the 1982 Forest Service planning rule. These have never been updated.
Anticipated PSQ within the NWFP area was based on the assumption that most of the volume was to come from four of the most productive national forests, the Gifford-Pinchot, Mount Hood, Umpqua, and Willamette (Johnson et al. 1993). The assumed harvest methods on these national forests also included the use of regeneration harvest (15 percent retention) for primarily timber volume production. These harvests with 85 percent removal of live trees tend to look nearly as visually striking as clear-cutting. This level of regeneration harvest was not implemented in large part due to the lack of social acceptance of regeneration harvest and timber production-focused projects. Volume awarded from these four national forests from 2014 to 2018 composed only 28 percent of the total volume awarded from NWFP forests.

In 1993, Johnson and others acknowledged that most of the projected volume of PSQ was anticipated from older forest (50 percent of volume from forest more than 200 years old). This has not been implemented as planned as the use of regeneration harvest has nearly ceased due to social concerns.

**By Design: Isolation of Timber Production Lands—a Paradigm Shift**

When the NWFP was implemented, the two objectives of “…conserving the ecosystems upon which species depend, and at the same time providing raw materials and other resources that are needed to sustain the health and economic well-being of the people of this country” (USDA FS and USDI BLM 1994: 26) were viewed largely as direct conflicts. Harvest levels from the NWFP area in the late 1970s and 1980s were in the billions of board feet per year—levels that were not sustainable. But regardless of the long-term sustainability of these harvest levels, communities and industry expected and relied on these timber outputs. In 1994, the Forest Service's historic practice and paradigm of harvesting old-growth forests and cutting the largest, often oldest, trees with a singular purpose of wood production, was set against the environmental movement and ecosystem stewardship. What has changed most dramatically since 1994, besides timber output, is the idea that these two objectives are not necessarily in conflict, especially (Pipkin 1998) at landscape scales.

The 1994 NWFP and other amendments (Eastside Screens, PacFish, InFish) were meant to be planning tools to move federal lands, the public, and the courts past this historic gridlock. The NWFP amendment had the goal of creating a framework for achieving both timber production and protecting ecosystem function and diversity. In this 1980s and 1990s context, (1) timber production and (2) protection of biological diversity and ecosystem function were separate goals. The two goals were kept separate both as ideas and physically, through management area delineation in each forest plan and the NWFP land use allocations (late-successional reserves, riparian reserve, adaptive management area, and the remaining matrix).

The current goals and objectives for the Forest Service are multiple use and include ecosystems services and ecological integrity. Collaboration, shared stewardship and a shifted focus toward ecosystem management in our multiple objective lens are now at the heart of the Forest Service mission. Today the Forest Service, our partners, collaborative groups, and the public are all using, and continuing to develop, new language to describe this synergistic connection and the multiple objectives that are the source of these paradigm shifts.
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Phrases such as “active management,” “restoration,” “creating resilient landscapes,” “fuels reduction,” and “thinning” describe how timber harvest and timber production can be elements or outcomes of projects designed to do much more than get logs to the mill. In fact, for the Forest Service, timely and financially sustainable implementation of restoration, fuels reduction, and the creation of landscape resilience often require timber harvest and the revenue produced from salable logs. We acknowledge that even when the best available science indicates that active management restoration is needed, conflicting values around timber harvest or other types of active management still exist, which we will need to address in upcoming planning efforts.

Furthermore, ecosystem and landscape resilience are not integrated into our existing forest plans. Current plans generally acknowledge forest health, usually under the context of forest health for the continuation of primarily timber production purposes, not necessarily holistic ecosystem function, resilience, or integrity. Existing forest plans often conflict with modern desired conditions.

Call-out box 6. Shared stewardship

"Shared Stewardship is about working together in an integrated way to make decisions and take actions on the land." - USDA Forest Service Chief Vicki Christiansen.

Today's forest land managers face a range of urgent challenges, including catastrophic wildfires, more public demands, degraded watersheds, and insect and disease epidemics. The Forest Service is committed to a shared stewardship strategy to address these challenges by working collaboratively to identify priorities for landscape-scale treatments. The agency works with a variety of partners to do the right work in the right place and at the right scale. By coordinating at the state level to prioritize land management, the Forest Service will be able to increase the scope and scale of critical forest treatments that support communities and improve forest conditions.

The shared stewardship strategy builds on a foundation of collaborative work, such as the Joint Chief's Landscape Restoration Partnership, the National Cohesive Strategy for Wild/and Fire Management, and the Collaborative Forest Landscape Restoration Program. It also builds on authorities created or expanded in the 2018 Omnibus Bill and the 2018 Farm Bill, such as the Good Neighbor Authority. The Forest Service will build on this foundation, working more closely than ever with states, tribes, and other partners to set cross-boundary priorities.

Conflicting Management Direction

One of the primary reasons that timber outputs from the NWFP area have been below anticipated levels (PSQ estimates) is that the plan expectation that matrix lands under the NWFP would emphasize timber production has not been met. Today, an estimated 1.5 million acres (8 percent) of the NWFP area aligns with timber production emphasis (table 2) (example map 7). In 1994, about 16 percent of the NWFP area emphasized timber production.
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Figure 14. Conflicting demands in matrix lands

The land available for timber production under the Northwest Forest Plan (NWFP) is much less than what was left undesignated in the NWFP. All undesignated lands in the NWFP were assumed to be timber production emphasis and dubbed “matrix.” Underlying land management plan direction and inventoried roadless area designation often conflict with NWFP assumptions about timber production. Matrix was assumed to emphasize timber production, but timber production goals in matrix are superseded by more restrictive stipulations. Examples of more restrictive management direction include minimum canopy cover requirements for ungulates and requirements for tree retention or limited tree stumps. Inventoried roadless areas generally prohibit timber harvest. Unmapped riparian reserve, survey and manage, and northern spotted owl critical habitat designations create further complexity and unpredictability for timber production. Northern spotted owl critical habitat designation that overlaps with the matrix allocation has reduced the land base available for primary timber production.

Forest Plan Management Emphasis

Understanding individual forest plan management emphases and amendments, constraints and standards and guidelines as related to vegetation management on our forests is complex. All forest plans in the BioA area have been affected by one or more large-scale amendments. In addition, each forest includes areas that have been designated or administratively withdrawn/Congressionally reserved throughout the years since 1994, including wilderness designations that prohibit timber harvest, or designations that allow limited or no vegetation management (national recreation areas, research natural areas, wild and scenic rivers). The combination of underlying forest plans, amendments, and administratively withdrawn/Congressionally reserved areas is complex, and forest plan direction across the BioA area is diverse. To better understand and summarize the forest plan management emphases as related to active management of vegetation, we have summarized the entire BioA area into four broad categories (modified from Ringo et al. 2016):
A. **Forest plan timber production emphasis**: areas where both the underlying forest plan and NWFP matrix overlap to emphasize timber production. These forest plan areas do not necessarily restrict mechanical treatments beyond standard guidelines and the areas may be used to meet wood production targets (for example, general forest, suitable timber, timber production). Timber production emphasis areas include areas where Eastside Screens 21-inch diameter at breast height standard applies depending on old-forest landscape conditions on forests outside of the NWFP area.

B. **Forest Plan multiple objective emphasis**: areas where mechanical treatments may be restricted by forest plans or an amendment such as the NWFP for multiple objectives. Mechanical treatments are limited but possible depending on forest conditions or management objectives (for example, elk winter range, riparian reserves, managed old growth, visual corridors). **NWFP late-successional reserves and riparian reserves are considered part of the multiple objective category of forest plan management emphasis.**

C. **Inventoried roadless areas**: the Roadless Area Conservation Rule generally prohibits timber harvest and road construction in Inventoried Roadless Areas.

D. **Preservation**: areas designated for long-term preservation by act of Congress or forest plan allocation, and no mechanical treatments are permitted (for example, wilderness or research natural areas), and areas where vegetation is only managed for non-vegetation objectives such as those used for national recreation areas.

In 2001, the Forest Service promulgated national regulations (known as the Roadless Area Conservation Rule) that permanently codified the designation of inventoried roadless areas on about 530,000 acres of matrix lands and 1.6 million acres of late-successional reserves (figure 15) within the BioA area. The rule generally prohibits timber harvest and road construction in inventoried roadless areas.

Northern spotted owl critical habitat designation also overlaps with timber management emphasis, and this overlap shifted management to an old-forest habitat conservation focus that was not necessarily consistent with planned timber production emphasis. This overlap reduced the total timber production emphasis area to approximately 940,000 acres, or 5 percent of all Forest Service lands in the NWFP area (figure 16, figure 17, and figure 18).

### Table 2. Forest plan management emphasis on matrix lands

*This table displays the percentage of area for four categories of management emphasis within matrix lands in the Northwest Forest Plan area.*

<table>
<thead>
<tr>
<th>Forest Plan Management Emphasis Category</th>
<th>Percentage of Matrix with each General Emphasis</th>
<th>Acres of Matrix General Emphasis</th>
</tr>
</thead>
<tbody>
<tr>
<td>Multiple objectives</td>
<td>54%</td>
<td>2,765,285</td>
</tr>
<tr>
<td>Timber production</td>
<td>30%</td>
<td>1,567,555</td>
</tr>
<tr>
<td>Inventoried roadless area</td>
<td>10%</td>
<td>531,329</td>
</tr>
<tr>
<td>Preservation</td>
<td>5%</td>
<td>230,871</td>
</tr>
</tbody>
</table>
Map 7. Vegetation forest plan management emphasis categories for the Bioregional Assessment area.

Categories are a combination summary of underlying Forest Plan direction, amendments including the NWFP Land Use Allocations, inventoried roadless area extents, and administratively preserved or withdrawn areas for a purpose other than vegetation management (for example, national recreation areas).
Table 3. Bioregional Assessment area by percentages under different forest plan management emphases

<table>
<thead>
<tr>
<th>Forest Plan Management Emphasis Category</th>
<th>Percentage BioA Area</th>
<th>Approximate BioA Acres</th>
</tr>
</thead>
<tbody>
<tr>
<td>Multiple objectives</td>
<td>43%</td>
<td>11,900,000</td>
</tr>
<tr>
<td>Preservation</td>
<td>23%</td>
<td>6,300,000</td>
</tr>
<tr>
<td>Inventoried roadless area</td>
<td>16%</td>
<td>4,860,000</td>
</tr>
<tr>
<td>Timber production</td>
<td>18%</td>
<td>4,290,000</td>
</tr>
</tbody>
</table>

Figure 15. Acreage of inventoried roadless area by Northwest Forest Plan land use allocation

Inventoried roadless area that was designated in 2001 by primary NWFP land use allocation. Most inventoried roadless area was designated in late-successional reserves and administratively withdrawn areas, but about 530,000 acres also were designated in matrix areas.
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Figure 16. Percentage of the Northwest Forest Plan (NWFP) area with timber production emphasis

Timber production under our current forest plans conflicts with habitat protection and other objectives. In 1994, an estimated 16 percent of the NWFP area had a timber production emphasis (other = matrix). Of these lands, only about 5 percent remain in timber production emphasis today. About 8 percent (1.58 million acres) have multiple objectives emphasis as a result of forest plan direction (for example, visual corridors, ungulate habitat), inventoried roadless area designation, or other designation (national recreation area). With the most recent northern spotted owl recovery plan, about 3 percent (644,000 acres) have been designed as northern spotted owl critical habitat, leaving approximately 940,000 acres (5 percent of NWFP area) of matrix, or primary timber production lands.

Figure 17. Designated northern spotted owl critical habitat in each land use allocation

Designated critical habitat management emphasis for northern spotted owl is most aligned with late-successional reserve, congressional reserves, and administratively withdrawn allocation objectives.

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Figure 18. Proportion of northern spotted owl critical habitat by actively managed Northwest Forest Plan land use allocations

Northern spotted owl critical habitat designation is most proportionally aligned with late-successional reserves on the Mt. Baker-Snoqualmie, Siuslaw, Olympic, Fremont-Winema, Mendocino, and Deschutes National Forests; more than 60 percent of northern spotted owl critical habitat is within late-successional reserves on these national forests. Non-alignment of northern spotted owl critical habitat with late-successional reserves and other reserves indicates a potential need to adjust land allocations in coordination with designated critical habitat for northern spotted owls and other species.

The primary intent of the matrix land use allocation under the NWFP was to support desired and probable timber production goals set forth in the plan, while also interfacing with other land use allocations to sustain old-forest and riparian areas. Most of the area where timber has been harvested since 1994 has been in matrix. Conflicts have added both tangible and perceived restrictions to mechanical vegetation management on matrix lands that we consider here. Conflicts generally fall into three categories:

**Desired Conditions:** spatial and design constraints include existing plan definitions, some plan components, Survey and Manage, lack of clarity of desired conditions for consultation processes, northern spotted owl critical habitat designation, and interpretation.

**Process:** constraints may include multiple layers of internal process, including the Regional Ecosystem Office, Regional Interagency Executive Committee, and adaptive management area processes.

**Investment and Tolerance Limits:** limits are reached by agency personnel related to planning process and perceived risk. Implementation of projects has limitations related to high complexity, process levels, and perceived litigation risk for the agency.
Estimates of timber production for national forests and grasslands outside the NWFP are complex because forest plan direction regarding these estimates has never integrated the adoption of Eastside Screens, including the 21-inch-diameter at breast height standard or PacFish/InFish.

An important consideration is the regional implication of real and perceived restrictions on timber production on NWFP lands. Harvest records in the Forest Service’s Pacific Northwest Region indicate that within the BioA area, national forests that are partly or completely outside the NWFP footprint tend to produce more timber volume than most NWFP national forests (the Willamette National Forest is the exception) in recent years. Examples of forests east of the Cascade Range that have produced more timber than wetter NWFP forests include the Malheur, Fremont-Winema, and Colville National Forests.

Another trend that is important to consider is the large number of acres treated in fire-dependent landscapes. Of all the BioA national forests and grasslands, the Fremont-Winema, Deschutes, Klamath, Lassen, and Shasta-Trinity National Forests treated the most acres between 2010 and 2017. This shift in timber harvest from the most intrinsically productive national forests of the Cascades and North Cascades to the generally less productively capable East Cascades and Klamath Ecoregions is likely the result of two primary factors: (1) the national focus on restoration of fire-dependent ecosystems and (2) conflicting management direction of the NWFP with both timber production and restoration treatments.

Road System Context
Roads and motorized transportation networks provide access for commercial and noncommercial harvest activity and other contracted work on national forests and grasslands, in addition to private inholdings and infrastructure for transportation of timber to mills. The National Forest System portion of the road system is not sustainable at its current extent and maintenance level, given current agency capacity (photo 7, figure 19). National Forest System roads were built primarily with revenue from timber sales and such funding contributes to road maintenance, but this revenue has declined in the past three decades as timber sale volumes have declined (Charnley 2006). Road condition and distance to mills affects the cost of vegetation management activities and the amount of revenue that can be gained from them. Other sources for funding for road maintenance have also declined in the past three decades. From 2003 to 2018, there was a 59-percent reduction in appropriated funding for road maintenance but only a 2-percent reduction in the road system.

Bridges are an example of a potentially acute road maintenance constraint related to timber harvest implementation. Many bridges do not have the load capacity to support timber haul or are in poor conditions, making timber harvest machinery transport unsafe. There are many more bridges on national forests in the western Cascades than east of the Cascades.
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Photo 7. Example of road damage on a paved road

Figure 19. Road miles by maintenance level and national forest in the Bioregional Assessment area

Road mileage and maintenance levels vary widely and while not directly correlated provide an indication of the maintenance backlog and needs along with access to a given forest.
Forest Plan Elements that Reflect Underlying Trust Issues
Both the 80-year exemption of the NWFP and the 21-inch standard of the Eastside Screens are relatively simple policy rules developed to deal with complex forest biodiversity issues. These elements were developed to help describe old forest. In addition to helping describe old forests, these elements reflect the public’s underlying differing values around old forests and trust issues surrounding federal land management that existed when the management direction was developed, and which still exist to some extent today. We acknowledge that land management planning alone will not resolve conflicts in values, tradeoffs, or build trust among stakeholders and the Forest Service. We are committed to learning how and why stakeholders hold different values and to providing transparent public engagement opportunities throughout the entire planning process to increase shared learning and build trusting relationships. Science can guide the Forest Service in the modernization of these elements, but understanding, trust, and social and political durability are the primary drivers of how policy is developed or changed.

Late-Successional Reserve Context Related to Timber Production
There is a need to support continued management in late-successional reserves to meet desired ecological conditions. This includes the development of old forest in wetter regions and resilience of landscapes to climate change and disturbance, such as fire in frequent-fire dependent and fire diverse (mixed severity) ecosystems. Treatments aimed at resiliency-based density reduction, species composition, and landscape pattern shifts can contribute to the coproduction of timber in some of the most productive landscapes.

Timber volumes derived from late-successional reserves currently, and will increasingly, bump against the 80-year exemption as trees grow and age (figure 20 through figure 22). We discuss more fully the complexity of defining old forests in the “Forest Ecology” section below. It is important to point out that depending on the ecosystem and landscape, there is a tipping point somewhere between 80 and 200 years where forests tend to take on the characteristics of old forest. How and what we define as old forest has important implications for management and implementation.

Figure 20. Timber volume by national forest and Northwest Forest Plan late-successional reserve and matrix land use allocations.
Approximate acres of merchantable-size timber volume (more than 12,500 board feet per acre of standing tree volume) by national forest and two primary NWFP land use allocations: late-successional reserve and matrix. Not all national forests and grasslands have the same potential to sell additional timber volume. Total acres of standing merchantable timber in matrix is about 4 million and about 5 million acres in late-successional reserves. Forests with less than 100,000 acres of matrix or late-successional reserves were excluded from this graphic.

**Figure 21. Index of late-successional reserve and matrix land use allocations by national forest**

*Forests with an index of 1 have almost all potential active management land base in late-successional reserves (little to no matrix lands). Forests with a value of -1 have nearly all potential active management land base in matrix lands and have little to no LSR. The Olympic and Siuslaw National Forests have almost all their potential active management areas in late successional reserves (value of 1). The Rogue River-Siskiyou has almost 40 percent more LSR than matrix lands. The Mt. Hood has almost 40 percent more matrix than LSR lands. More lands in late successional reserves as compared to matrix means that constraints like the 80-year exemption will likely make restorative or old-forest enhancing treatments more difficult, especially as trees naturally age.*

With the 80-year threshold for exemption, more acreage will become too old for exempted vegetation treatments such as thinning. Of the 6.1 million acres of forests that are 80 to 200 years old, the Okanagan-Wenatchee has the most acres by almost 950,000 (figure 22). The Shasta-Trinity, Rogue River-Siskiyou, Klamath, Gifford Pinchot and Six Rivers National Forests have between 400,000 and 750,000 acres of forest aged 80 to 200 years old. In 1993 it was estimated that although late-successional reserve volumes are not included in PSQ calculations, an additional volume of 100 to 170 MMBF might be obtained from reserve areas (Johnson et al. 1993).
Figure 22. Acres of merchantable timber volume by age class, forest, and Northwest Forest Plan land use allocation

National forests and grasslands have most of their standing volume in the 80- to 200-year age class in both matrix and late-successional reserves across the NWFP area. Acres of forest that contain merchantable timber (small-tree and larger size classes) and are between 30 to 80 years old within the NWFP area by forest and land use allocation, including late-successional reserves and matrix, are estimated here. Some national forests and grasslands, including the Rogue River-Siskiyou, Olympic, Siuslaw, and Mt. Baker-Snoqualmie, have most of these younger forests. With the 80-year exemption, more acreage will become too old for exempted vegetation treatments such as thinning over time. Of the 6.1 million acres of forests 80 to 200 years old, the Okanogan-Wenatchee has the most acres by almost 950,000. The Shasta-Trinity, Rogue River-Siskiyou, Klamath, Gifford Pinchot, and Six Rivers National Forests all have between 400 and 750 thousand acres of forests aged 80 to 200 years old.
There is a need to support national forests with Eastside Screens contribution to timber volume output (see "Forest Ecology" section for more context on Eastside Screens). One component of the Eastside Screens that affects both restoration and resilience project implementation and timber product output is a forest plan standard that restricts harvest of trees greater than 21 inches diameter at breast height. The Eastside Screens require a landscape-scale analysis of forest structure. If the results of this analysis indicate there is a lack of old forest as compared with historic reference conditions, then there would be no harvest of trees, regardless of species, greater than 21 inches diameter at breast height outside of old forest. This standard also has been implemented to effectively restrict harvest of trees 21-inches diameter at breast height in old forest of all structural classes, even when harvest of some of these trees would maintain and conserve old forest. This standard, intended to conserve old and large trees on a landscape where the largest trees were historically cut, can conflict with attaining desired restoration across the landscape (Johnston et al. 2018, Merschel et al. 2019). This is especially true where ingrowth of white fir and grand fir undermines the protection and culturing of ecologically desired species such as ponderosa pine in frequent-fire dependent and fire diverse (mixed severity) ecosystems.

Forest Service Staffing and Budget Context

Increasing project costs, reduced agency staffing, and constrained budgets hamper efforts to increase the pace and scale of forest restoration activities. Forest budgets for restoration have not been increasing and less money is available because of constantly decreasing budgets and fire management taking up a larger portion of the overall agency budget (USDA FS 2015b) (figure 23 through figure 25).

![Figure 23. Northwest Forest Plan area staffing by agency and Forest Service regional staff from 1993 to 2017](image)

**Figure 23. Northwest Forest Plan area staffing by agency and Forest Service regional staff from 1993 to 2017**

*Staffing reductions indicate reduced capacity to plan and implement vegetation management activities (Adapted from Grinspoon et al., n.d.). Nationally, the Forest Service employs approximately 3,000 permanent non-fire employees in the forestry (1,036) and forestry technician (1,973) series (human resource roster report, February 7, 2018), which is 17 percent*
of the non-fire permanent workforce. There are fewer than 100 field level professional foresters in all of Forest Service Pacific Southwest and Pacific Northwest Regions (regions 5 and 6, respectively). One out of four permanent non-fire foresters and forestry technicians has at least 25 years of service and is nearing if not already eligible for retirement. Additionally, one out of three permanent non-fire foresters and forestry technicians has at least 20 years of experience. In Regions 5 and 6, there are about 260 foresters at all levels and about 600 forestry technicians.

Figure 24. Forest Service fiscal year 1995 budget appropriations

This chart provides an indication of funding focus and capacity at the beginning of management under the Northwest Forest Plan, PacFish and InFish plan amendments.

Figure 25. Forest Service fiscal year 2015 budget appropriations

This graph highlights the current funding focus and capacity and the changes when compared to figure 23 and figure 24. Note the much larger proportion of funding allocated to wildland fire management.
Conclusion
In this chapter, we recognized the significant values that national forests and grasslands deliver to all people, especially to those in communities that directly depend on economic benefits from these lands. Moving forward, we know that balancing complex ecological needs with the growing social and economic needs of communities within the BioA area will take a commitment to ongoing communication, collaboration, and coordination to develop solutions that address these challenges.

In chapter 2, you’ll read about our current relationship with American Indian tribes and opportunities to better work with these tribes to help secure their treaty rights, especially when ecosystems within the BioA area provide and support a broad range of ecocultural resources important to tribes, including foods, medicine, materials, and nonmaterial values.
Chapter 2: Tribal Rights and Interests

Introduction
The Forest Service understands the significance and interconnectedness of treaty rights and resources within American Indian tribal cultures. Many tribal members continue to foster longstanding, customary relationships with natural resources on national forests and grasslands, continuing an interdependent relationship whereby tribal practices nurture ecological systems, and those systems in turn nurture and sustain cultural identity and continuity. Laws protect American Indian rights to use and possess sacred objects; to protect their ancestral graves, archaeological, and cultural sites; to secure the freedom of worship through ceremonial and traditional practices; and to collect native plant and animal resources for traditional cultural purposes. Memoranda of understanding for collaboration, consultation, and cooperation in managing natural resources on national forests and grasslands are in effect between the Forest Service and multiple tribes within the Bioregional Assessment of Northwest Forests.

The Forest Service maintains government-to-government relationships with American Indian tribes that have sovereign governments. These relationships are important for protecting and managing ecological resources to honor, support, and respect cultural, spiritual, and community interests and to integrate these as much as possible into agency planning and project implementation. In their treaties with the United States, the tribes expressly reserved many of their aboriginal rights, including rights to harvest a range of resources both on and off reservation lands. The Forest Service has certain legal responsibilities to American Indian tribes beyond those identified in treaties; these are clarified in statutes, executive orders, and case law interpreted for the protection and benefit of federally recognized American Indian tribes. In meeting these responsibilities, the Forest Service consults with tribes whenever proposed policies or management actions may affect tribal interests.

More than 70 federally recognized American Indian tribes have tribal lands or territory within the Northwest Forest Plan area (Vinyeta and Lynn 2015), and there are additional tribes within the NWFP area that are not currently federally recognized (Long et al. 2018). A significant portion of lands ceded by tribes were established as part of the National Forest System by the Organic Administration Act of June 4, 1897. While federal law governs relations with all federally recognized American Indian tribes, each tribe is different and is recognized as a separate and unique government. While some tribes have treaties, other tribes were recognized under different authorities and have different rights.

Treaty rights and the historical relationships between tribes and the lands they aboriginally relied upon may differ greatly among tribes. Cultural differences between tribes can also be significant. In some cases, several tribes may each have legitimate interests in the same lands because they each may have occupied or otherwise used those lands during overlapping or different historical periods. These factors and others combine to make each Forest Service-tribal consultation relationship unique.
Tribal communities often rely on traditional foods, such as water, salmon, game (such as elk and deer), roots (such as cous, camas, and bitterroot), and berries (such as huckleberries and chokecherries). While these plants and animals, as well as water and other resources, have cultural significance, they also provide nutrition to many tribal people. Culturally significant foods are especially important and provide critical subsistence given the high incidence of poverty in American Indian communities.

Environmental justice policies are also relevant to American Indian populations, though some applications may not fully capture the dynamics of tribal populations particularly well. For example, American Indian population representations in the census are prone to inconsistent patterns as respondents change their self-identification of race and ethnicity.

The following section provides an overview of the relationships that the Forest Service has with tribes and offers general observations. This section outlines resource management themes that are frequent consultation topics and highlights ongoing resource concerns that will be discussed collaboratively with interested and affected tribes throughout planning processes, project implementation, research, and monitoring.

What is Working Well

**What is Working Well 1—Ecosystems Support Tribal Values**

Ecosystems within the BioA area provide and support a broad range of ecocultural resources important to tribes, including foods, medicine, materials, and nonmaterial values. Newer and emerging forest management concepts, including restoring frequent-fire dependent systems, align with tribal ecocultural resource perspectives, and are a central strategy for promoting and revitalizing tribal resources. Restoring fire, combined with other tending practices, perpetuates resource conditions that support tribal use. Landscape-scale restoration ensures long-term resource sustainability and availability and provides social and economic benefits, including resource security and opportunities for restoration-related work.

**Call-out box 7. First foods**

First foods are traditional foods that have been and remain significant in some American Indian tribal diets and cultures (Lynn et al. 2013). Culturally significant foods, including water, fish, big game, roots, and berries, are used in ceremonies as well as for sustenance and economic benefit to perpetuate American Indian sovereignty and cultures. The Forest Service understands the tribal significance of treaty rights and traditional resources, including first foods. Many tribal members have longstanding, customary relationships with natural resources on National Forest System lands. Tribal practices nurture ecological systems as part of interdependent relationships, which in turn nurture and sustain cultural continuity and identities. Ideas to promote tribal ecocultural resources, such as first foods, are consistent with emerging direction in forest management and the management options presented in the Bioregional Assessment.
What is Working Well 2—Aquatic Resource Benefits

The NWFP’s Reserve and Aquatic Conservation Strategy and its four main components, the riparian reserve system, key watersheds, watershed analysis, and watershed restoration, contribute to improved watershed conditions through both passive and active management. Active management strategies, including restoration actions in key watersheds, also provide needed restoration to targeted areas within the BioA area to improve refuge habitat for at-risk fish stocks and serve as sources of high-quality water. Combined, these management approaches provide long-term resource benefits that are central to tribal well-being.

Pacific Northwest native fish, particularly salmon, steelhead, redband trout, and lamprey have been important resources for tribal communities for thousands of years. These resources are often included in their consideration of first foods and tribal treaty rights protect them. Under the NWFP, habitat restoration for these species has increased, and such restoration is increasingly important to protecting species that are at high risk from continued climate change and other stressors.

Aquatic restoration projects often require timber acquired from National Forest System lands to place in streams and floodplains to restore habitat for fish and other aquatic species. Tribes frequently finance and implement such restoration projects.

What is Working Well 3—Tribal Collaboration

The Forest Service continues fostering collaborative partnerships with tribes built upon the legal foundations and the desire to include native knowledge in forest planning, research, implementation, and monitoring. Agency collaboration and consultation with tribes have benefited from the increased role of tribal liaisons and through the increased prominence of American Indians in key leadership positions—such as forest supervisors and district rangers—within the Forest Service.

Key Change Issues

Long et al. (2018) identified key change issues in Tribal Ecocultural Resources and Engagement.

Key Change Issue 1—Consultation, Tribal Rights and Access, and Federal-Tribal Forest Management Compatibility

The 20-year monitoring report, Strengthening the Federal-Tribal Relationship: A Report on Monitoring Consultation Under the Northwest Forest Plan (Vinyeta and Lynn 2015), assessed the federal-tribal relationship through interviews with tribes within the NWFP area, and through several case studies to reveal tribal experiences and perspectives on how their rights and interests have been affected by federal policy. The three primary categories for key findings include consultation, tribal rights and access, and federal-tribal forest management compatibility.
Chapter 2: Tribal Rights and Interests

The 20-year monitoring report provides a brief summary of these interrelated areas (Vinyeta and Lynn 2015):

“Of particular importance is the need to align tribal and federal visions on what constitutes consultation, the need to ensure that agency staff are culturally competent and informed on treaty rights, other tribal rights, the federal trust responsibility, the history of federal-tribal relations, and the need to ensure that tribes’ needs, knowledges, and practices shape not only tribal, but also federal forest management.”

Planning Considerations
Planning considerations applicable to land management planning from Long et al. (2018) include the following:

• Recognize each tribes’ unique vision for consultation and integrate into management actions.
• Continue government-to-government relationships.
• Some tribes may desire formalized consultation policies through memoranda of understanding or agreements that are individualized and include procedures, tribal contacts, and specifically identify issues of importance that affect tribal interests and rights.
• Customize tribal notification to allow tribes more efficiency in identifying topics they deem critical.
• Use more personal forms of communication, including face-to-face meetings and phone calls that may add legitimacy and can be more culturally appropriate for tribal members.
• The Forest Service should inform tribes of intergovernmental forum opportunities related to the planning process and encourage tribal participation.

Tribal rights and access recommendations and potential planning solutions include the following:

• Develop and adopt procedural frameworks that protect sensitive tribal and traditional knowledge.
• If needed, develop and use effective conflict resolution processes in federal-tribal agreements such as memoranda of understanding and memoranda of agreement.

Consult with tribes to find appropriate solutions in situations where management priorities and actions (such as permitting, road access, and interactions with commercial and noncommercial harvesters for nontimber forest products) interfere with tribal rights and access (Vinyeta and Lynn 2015). Improving federal-tribal management compatibility recommendations and planning solutions include the following:

• Consult and collaborate with tribes to (1) make federal and tribal forest management practices more compatible, (2) align federal and tribal management programs, and (3) improve time and cost efficiency for both agencies and tribes.
Chapter 2: Tribal Rights and Interests

- Compatible management approaches include a focus on ecosystem management rather than single species management; holistic forest conservation of forest resources, including riparian areas, water quality, and fish and wildlife habitat; and management practices that include prescribed fire.

Look for opportunities for tribal land management on ancestral lands and promote policies like the Tribal Forest Protection Act that fosters tribal management. Tribal communities maintain widespread interest in forest ecosystems. Tribal engagement, consultation, and partnerships are important to achieving land management objectives as prescribed in the Forest Service 2012 Planning Rule. Engagement also upholds tribal rights and federal responsibilities and recognizes the importance of tribal ecocultural resources on ancestral lands. Recommendations and potential planning solutions associated with engaging tribes in forest planning and management include the following:

- Designate special tribal stewardship areas to achieve both social and ecocultural objectives for tribes and the Forest Service.
- Coordinate land management planning with related planning efforts of federally recognized American Indian tribes and consider the policies of approved tribal land resource management programs. Coordination is an important element in land management planning and is required under agency regulation and policy as well as regulations outlined in the Federal Land Policy Management Act of 1976.

Key Change Issue 2—Availability of Ecocultural Resources

Both social and biophysical factors reduce the ability of tribes to obtain ecocultural resources from public lands in the desired quality and quantity. Important tribal resources considered “first foods” can be degraded by multiple factors, including changes in fire regimes, invasive species, altered hydrologic systems, species extirpation, and other historical legacies, including reduced or omitted tribal landscape practices and influence. The inability of the tribes to readily obtain ecocultural resources from national forests and grasslands is due to the following:

- Reduced fire frequency in drier, inland areas, and locations near historic tribal settlements, trade and travel routes, and harvesting and hunting areas have reduced the production of ecocultural resources unique to such areas. This includes areas representative of unique ecological transition zones, including areas defined by the edges between forests and grasslands or wetlands.
- Additional Endangered Species Act listings of fish species and distinct population segments, especially salmonids and other anadromous fish, have occurred even after the NWFP and other aquatic strategies in the BioA area, such as PacFish and InFish, were developed and adopted.
- Barriers to fish passage from forest roads continue to affect aquatic system connectivity and productivity, where road-stream crossings are not passable to native fish, and a large percentage of these locations have not been sufficiently evaluated.
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- Hydrologic systems, instream flows, and water quality sustain important fishery, aquatic, and riparian resources for tribes. Existing impacts to these systems include irrigation, mining operations, and water diversions for illicit marijuana-based operations, and may be further impacted by population growth and climate change.

Planning Considerations
Management recommendations for forest plan revision consideration were presented by Long et al. (2018) and include the following:

- Improved productivity and availability of tribal ecocultural resources is needed, and results from increased active forest management, including understory and variable overstory thinning and increased use of fire. Activities can focus on conserving large old trees, cultural sites, and other resources vulnerable to severe disturbances.

- Habitat maintenance and restoration for species reintroductions should be consistent with promoting direct tribal needs, providing for species’ habitat needs, and working with conservation partners to achieve multiple benefits.

- Address aquatic organism passage and habitat restoration projects on National Forest System lands and work cooperatively with tribes to accelerate the program.

- Tribal ecocultural resources are associated with diverse vegetation communities, including hardwood-dominated forests and woodlands (including “old growth”), and non-forest communities such as bogs, prairies, wetlands, and meadows. Classifications and management goals that focus on vegetation type and seral stage should capture these diverse and unique conditions desired by tribes.

- Continue and improve resource monitoring that is relevant to promoting tribal ecocultural resources.
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Ecological Integrity and Resilience

Call-out box 8. Ecological integrity

Ecological integrity is the quality or condition of an ecosystem when its dominant ecological characteristics (for example, composition, structure, function, connectivity, and species composition and diversity) occur within the natural range of variation and can withstand and recover from most perturbations imposed by natural environmental dynamics or human influence (36 CFR 219.19).

The Forest Service 2012 Planning Rule incorporates ecological integrity and directs forest plans to align approaches across the broader landscape for an all-lands approach to ecological sustainability. Planning for ecological integrity supports ecosystem resilience under changing conditions, and systems that can recover from disturbance (36 CFR 219: 21176). Integrity is a course-filter designed to maintain biological diversity.

Resilience (figure 26) is a concept applied throughout this assessment and is also defined in the Forest Service 2012 Planning Rule Handbook 1909.12. (The term describes a state of being or a potential, not a specific end condition. Resilience is the ability of a system to regulate itself as it transitions among stable states (Gunderson 2000, Holling 1973) and can be especially powerful in our era of rapid change (Scheffer et al. 2015, Standish et al. 2014). Examples of these transition triggers, or disturbances, are fire, insects, disease, invasive species, and climate change (O'Hara and Ramage 2013, Puettmann 2011). Resilient ecosystems maintain keystone structuring processes, which are sources of renewal and function across multiple scales (Hessburg et al. 2015; Spies et al. 2018a, 2018b). One of these keystone structuring processes in the Bioregional Assessment of Northwest Forests is fire (Keane et al. 2009) (photo 8 and photo 9).

It is important to note that ecological resilience is not always a desired condition. In fact, ecological resilience may directly conflict with desired conditions. For example, homes and communities exist near an ecologically resilient forest where wildfire may occur within normal ecological ranges. In this case, a natural process that is part of maintaining ecological integrity may not be socially desirable for that community. We can use concepts

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8 This concept is defined in the scientific literature as a means of evaluating ecological conditions in terms of their sustainability. The concept of ecological integrity is required by use in National Forest Service management planning (36 CFR 219 §219.8(a)). “Plan components for ecological integrity would be required to take into account the interdependence of ecosystems, impacts from and to the broader landscape, system drivers and stressors including climate change, and opportunities to restore fire adapted ecosystems and for landscape scale restoration. Plan components would be also be required to maintain or restore air, soil and water resources, and to maintain or restore the ecological integrity of riparian areas” (36 CFR 219 §219.9(a)).

9 Resilience: The ability of an ecosystem and its component parts to absorb or recover from the effects of disturbances through preservation, restoration, or improvement of its essential structures and functions and redundancy of ecological patterns across the landscape (USDA FS 2015b: 16).
like ecological integrity and resilience as anchors for management and desired conditions, but they have limitations in the context of communities and changes in climate and land use. Landscape resilience measures the stability and response of an ecosystem state to disturbance. Vegetation structure, species, composition, and pattern are ways to measure landscape resilience. It is used in this document without the integration of social values of those landscapes.

**Ecosystem Conditions: Resilience**

The comparison of current conditions to known resilient conditions generates powerful insights into how ecosystems function, developed, and even how they manifest the conditions seen on the ground today (Higgs et al. 2014, Hobbs et al. 2014, Safford et al. 2012). Often we use past resilient conditions as indicators of resiliency because they are known and measurable. The use of these measurements of the alignment or departure of ecosystems from past resilient conditions to assess resilience requires the assumption that past conditions point us in the right direction, toward ecosystem function. In addition, we need to consider predictions about future resilient conditions to prepare landscapes and ecosystems to cope with changing environmental conditions.

For these reasons, ecosystem conditions are compared here with both past and current condition, in addition to indices of future conditions, drivers, and stressors.

**Future Ecosystem Resilience**

Reference conditions providing guidelines for how to manage our ecosystems for resilience are essential to translate a concept of resilience into practice. Using historic reference conditions related to multiple aspects of our landscapes and forests is a critical foundation for understanding resilience. Given our current changing climate, particularly as it relates to wildfire behavior, extent and seasonality, projections of future conditions are also essential when managing toward forest resilience. The changes highlighted in future large fire suitability in the BioA are important to describe not only how the environment is projected to change, but to project and draw meaningful conclusions about the degree of change and impact to forests, specifically in terms of structure, species composition, and landscape patterns.

Subsequent sections of this chapter will highlight the ways existing land management plans promote ecosystem resiliency as related to forest ecology, wildlife, aquatics, and fire, and will provide recommendations to promote resiliency where it is lacking. The chapter ends with a climate change section, one that demonstrates how National Forest System Lands are equipped, to varying degrees, to cope with climate change, followed by concerns that will challenge ecosystem resiliency.
When a disturbance occurs, even a natural one such as fire, drought, insects or disease, an ecosystem responds in different ways. It can be resistant (very little change occurs, and the system stabilizes quickly), resilient (more change occurs but the ecosystem stabilizes eventually), or unstable (the ecosystem changes states completely). One example of this is a frequent-fire dependent forest historically dominated by ponderosa pine that now has dense stands of white fir or grand fir. The forest could experience a low-severity groundfire and change very little (resistant), or it could experience a mixed-severity fire and take years to recover, but eventually return to a state resembling conditions before the fire. Finally, a large high-severity fire could burn a large area of forest to the ground with the ecosystem unable to regenerate ponderosa pine, instead transitioning to a fir-dominated forest or a grassland or shrubland ecosystem (unstable). This figure does not incorporate climate change, changes in land use, or social factors.
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Photo 8. Fire on the Rogue River-Siskiyou National Forest, 2018

Photo 9. Fire on the Olympic National Forest in the Hamma Hamma area, 2018
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Forest Ecology

What is Working Well

What is Working Well 1—Reserve Systems
The reserve network of the NWFP, including late-successional reserves, riparian reserves, and congressionally reserved lands, is part of a designation-based landscape-scale approach that has worked well for conserving a network that supports many aspects of overall ecosystem health. This includes support for aquatic habitat and conservation of habitat for wildlife species that use dense, multi-layered old forest. The reserve network also ensures consistent management direction in each type of land use allocation.

Other plan amendments, such as PacFish, InFish, Eastside Screens (USDA FS 1995) and Sierra Nevada Framework (USDA FS 2004), have also been successful in achieving some desired network outcomes, including connecting and conserving aquatic habitat and dense, multi-layered forest.

While land use allocations and the reserve network have benefited multiple resources, some adjustments to create landscape resilience, especially in frequent-fire dependent and fire diverse (mixed severity) ecosystems are needed. Furthermore, to ensure a well-connected network of reserves that will persist, the application of best available science is needed to ensure that climate change refugia and fire refugia are incorporated into the reserve network.

What is Working Well 2—Conservation of Dense, Multi-layered, Old-growth Forests
The NWFP conservation strategies and other strategies, such as the Eastside Screens and the Sierra Nevada Framework, have been effective in stemming the loss of old trees and the type of old-growth forest that is a focus of the NWFP, mainly dense, multi-layered forest (Spies et al. 2018a). This type of old-growth forest is generally considered stable on federal lands and has increased slightly since 1993, providing the abundance, diversity, connectivity, and availability needed to support ecosystem functions and specific types of old-growth-dependent species in the BioA area (figure 27). This reversal of the pre-NWFP trend of old-growth loss is mainly a reflection of stopping clear-cutting practices and allowing trees to grow. Compared to preindustrial logging levels, old-growth forests are now described as islands of federally owned old forest among other land ownerships (figure 28).

While this old-growth conservation approach has been successful in some respects, old-growth forest types not defined or emphasized in the NWFP, Eastside Screens, or Sierra Nevada Framework are increasingly at risk of loss to fire (Spies et al. 2018a). These include old-growth forests in frequent-fire dependent and fire diverse (mixed severity) ecosystems. Loss of old growth from recent wildfires in California, southern Oregon, and east of the Cascade crest have been masked by gains in old-growth forest west of the Cascade Range where trees initiated after large 20th century fires have now grown into the old-forest category; when examining frequent-fire dependent ecosystems separately, acres of old growth have declined (Davis et al. in progress).
Regeneration harvest of old forest has been significantly reduced since the 1990s, particularly on federal forest lands in the Pacific Northwest, where the practice “virtually ceased in the early 21st century due to social objections and litigation.” Currently, harvest systems that “retain significant structural elements of the pre-harvest stand have largely replaced clear-cutting” (Franklin et al. 2018:108).

**Figure 27. Snapshots (1993 and 2017) of old-growth abundance and change**

Old-forest levels 1 and 2 represent areas where the amount of old forest on federal lands is at levels above (1) or within (2) the long-term average that occurred before logging and extensive fire exclusion. Level 3 represents areas where the amount of old forest is below the range expected for the area. No area with very low levels of old forest exists at this analysis scale. The “Change” map displays the percentage change in old-forest levels: Overall, old-growth habitat has remained stable but there have been changes in where the habitat is found through losses and additions.
Figure 28. Historic and current amounts of old forests
Landscape portrait of how estimated levels of old forest have changed through the 20th century. Before industrial logging (circa 1940) much of the BioA area was old forest. Today, old forest tends to be better described as islands of federally owned old forest among other land ownerships and land uses. Current old-growth forest in the NWFP area was estimated here using Old Growth Structural Index 80 from the NWFP 25-year monitoring report. Outside the NWFP area old-growth was estimated based on large tree (greater than about 20 inches diameter at breast height) densities greater than 8 trees per acre based on Merschel and others (2019).

Key Change Issues

**Key Change Issue 1—Landscape Scale Restoration is Needed**
Restoration is needed to improve and maintain ecological function and resilience so that our national forests and grasslands can meet the needs of our communities and society now and in the future.

About 65 percent (17.8 million acres) of the national forests and grasslands across the BioA area lack historic structural diversity and resilience and do not adequately contribute to ecological integrity (map 8). Unless this issue is actively addressed, ecological integrity could continue to decline and the benefits from the national forests and grasslands could be reduced for future generations. For methodologies, see DeMeo et al. 2018 and Ringo et al. 2019.
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Approximately 9.6 million acres within the BioA area need some type of forest structure restoration to improve and maintain ecological function and resilience. Of the 9.6 million acres that require restoration, about 2.2 million acres, primarily in our wettest, most fire infrequent ecosystems, need succession restoration\(^\text{10}\) to enhance tree growth and snag development. Natural ecosystem processes and treatments that maintain and enhance old-growth forest are needed to develop and maintain ecosystem function (map 9).

The remaining 7.4 million acres have denser vegetation than they would have naturally supported, and they need either mechanical or fire treatment\(^\text{11}\) or both to return to more natural, sustainable, and resilient densities (map 10). About 3 million of the 7.4 million acres also lack large trees, and therefore need a combination of disturbance and succession restoration to alter density along with time for the development of old-forest attributes.

Restoring riparian areas in concert with upland forests can be difficult because there is not necessarily synchronization of management direction between the two, even when best available science indicates needed restoration. Current land management plans offer descriptions of desired conditions for riparian areas that are too general and are not linked to landscape-level vegetation restoration or resiliency needs.

It is important to note that this analysis only focuses on the broad structural shifts needed across the BioA area. Specific forest restoration design occurs at the project level, based on site-specific data. The specific complexities of landscape and stand structures are not outlined here.

Restoration need is summarized in the *Synthesis of Science to Inform Land Management Within the Northwest Forest Plan Area* (Spies et al. 2018) and includes the following specific needs:

- Increase vegetation diversity in plantations and accelerate development of older forest structure and composition.
- Reduce fragmentation and increase connectivity of older forest patches.
- Create or promote early-seral vegetation where needed to provide seral stage and landscape diversity.
- Restore disturbance processes (for example, fire) where feasible.
- Restore low- and mixed-severity fire as key ecological processes where appropriate.
- Increase areas of open old forests where ecologically appropriate to promote resilience to fire and climate change and meet needs of species.

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\(^{10}\) Succession restoration essentially means to restore forests to resilient ecological conditions by allowing or culturing forests to grow into later seral stages. Later seral stages vary by forest type but generally include larger trees, more complex spatial arrangements, elements of dead and dying trees, and various ecologically appropriate species compositions.

\(^{11}\) Disturbance restoration is needed when some elements of earlier seral stages are needed to restore forests. Some type of missing natural process, primarily fire, needs to be restored or mimicked through active management, or allowed to occur. Disturbance restoration treatments could take the form of some combination or independent application of mechanical treatment, prescribed fire, or wildfire management.
• Develop landscape-level strategies to create desired mosaics of open and dense old forest and to increase resilience and meet simultaneous needs of wildlife species and ecological integrity.

Road systems have also been shown to negatively affect terrestrial and aquatic biological diversity and ecosystem processes (Forman and Alexander 1998, Trombulak and Frissell 2000). So, although roads are a critical element of restoration implementation, reducing roads through decommissioning is important for meeting many biodiversity goals (Franklin and Johnson 2012, Trombulak and Frissell 2000). We acknowledge that conflict can exist between the use of roads and other objectives, such as habitat. No decisions have been made, and these tradeoffs will need to be addressed in upcoming planning efforts.
Map 8. Structural departure from more resilient reference conditions by watershed within the Bioregional Assessment area

Reference conditions are based on fire regime condition class assessment by vegetation type and landscape unit. Highest absolute departure occurs along the east Cascade Range slopes and foothills, Klamath Mountains/California high north Coast Range Cascade Mountains and northeast Washington, and the Coast Range (adapted from Ringo et al. 2019).
Map 9. Succession restoration needed within the Bioregional Assessment area

About 5.2 million acres within the BioA area need succession restoration, including tree growth and snag development. This includes 2.2 million acres of tree-growth-only needs (primarily national forests and grasslands west of the Cascade Range crest) and a 3-million-acre subset of lands where disturbance and succession are needed (for example, thinning, under burning, and tree growth) (adapted from Ringo et al. 2019.)
Map 10. Disturbance restoration needed within the Bioregional Assessment area

This map shows that 7.4 million acres within the BioA area need mechanical or fire treatments (for example, thinning or under burning) to reduce density and restore forests (disturbance restoration need). If a watershed has a 45- to 70-percent restoration need, that watershed needs urgent restoration to maintain ecosystem resilience (adapted from Ringo et al. 2019).
Figures 29 through 32 describe restoration needs in different ways using four factors that can be considered when creating strategies for forest plan modernization. They include 1) forest plan management emphasis (figure 29), 2) NWFP land use allocation (figure 30), 3) fire ecology category (figure 31), and 4) by national forest (figure 32). The first two factors are related directly to current forest plan direction, both in the underlying forest plans and in the NWFP amendment. This shows that forests generally have the most restoration need (both disturbance and succession) in areas where there is multiple, often conflicting, direction. In the NWFP area, the primary land use Allocation of adaptive management areas, congressionally reserved, administratively withdrawn, late-successional reserve, and matrix all have restoration needs, with almost equal amounts within matrix, late-successional reserves, and administratively withdrawn areas.

When examined in terms of fire ecology categories (frequent-fire dependent, fire diverse (mixed severity), and fire infrequent), restoration need is found mostly in frequent-fire dependent zones and is also apparent in fire diverse areas. This makes sense, because the structural effects of fire exclusion, historic timber harvest, and grazing have tended to be most apparent in frequent-fire dependent systems where tree density is much higher than historical reference conditions and old, large trees are less common on the landscape than they were historically, although younger shade-tolerant species continue to grow. Fire diverse ecosystems have had structural changes that are often less apparent as compared to species composition shifts, which are not measured in the BioA.

Restoration need is also summarized by national forest, and is an important consideration given that forest plan revision under the 2012 planning rule is completed at the forest level. Therefore, considering the amount of restoration need by forest could be a factor contributing to strategies to modernize forest plans. The Okanogan-Wenatchee, Fremont-Winema, Shasta-Trinity, Rogue River-Siskiyou, Deschutes, and Klamath National Forests each have more than 30,000 acres of disturbance-only or disturbance-then-succession restoration need.
Figure 29. Acres of disturbance restoration and succession restoration need by forest plan management emphasis bin within the Bioregional Assessment area

Almost 3.5 million acres of disturbance restoration need is within lands that have multiple-objective management direction (Ringo et al. 2019).

Figure 30. Acres of disturbance restoration and succession restoration need by Northwest Forest Plan land use allocation

More than 3.2 million acres of both matrix and late-successional reserve need some combination of fire or mechanical treatments to change forest structure (Ringo et al. 2019).
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**Figure 31. Forest restoration and fire dependency**

Forest restoration need is categorized by fire dependency, including frequent-fire dependent, fire diverse (mixed severity) and fire infrequent. Both succession restoration and disturbance restoration are needed, sometimes in combination. Most current need can be seen in frequent-fire dependent landscapes and includes high levels of disturbance restoration need (Ringo et al. 2019).

**Figure 32. Forest restoration need by national forest within the Bioregional Assessment area**

The Okanogan-Wenatchee, Fremont-Winema, Shasta-Trinity, Rogue River-Siskiyou, Deschutes, and Klamath National Forests each have more than 30,000 acres of disturbance-only or disturbance-then-succession restoration need. Number of acres of restoration need will likely increase through time. Given the sheer volume of current and future needs, modern forest plan direction related to restoration could help address this trajectory.
Current departure from resilient ecological function is also apparent when examining the abundance, distribution, and diversity of various tree species (species composition) and landscape pattern, primarily in frequent-fire dependent and increasingly in fire diverse (mixed severity) ecosystems. Recent research has taught us the following:

- An understanding of forest species composition, including the dynamic interactions between species composition and structure, is integral to gauging and managing for landscape resilience (Hessburg et al. 2016, Tepley et al. 2013).

- Fire-dependent landscapes that historically were frequently disturbed by wildfire have had major shifts in species composition over about the past 100 years (Spies et al. 2018).

- In moister, fire-diverse landscapes, the effect of fire exclusion has been to significantly reduce the amount of early- and mid-successional vegetation that otherwise would now exist on the landscape as well as landscape-scale heterogeneity in forest composition, structure, and patch sizes (Haugo et al. 2019, Spies et al. 2018, Tepley et al. 2013).

- Some natural processes, including fire, are excluded across much of the BioA area (Balch et al. 2017).

Photo 10. Disturbance restoration need on Deschutes National Forest

*Ponderosa pine overstory with grand fir understory.*
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Photo 11. Disturbance restoration need on Okanogan-Wenatchee National Forest
A typical dense, multilayered mixed-conifer patch in the eastern Washington Cascade Mountains. Note the relict ponderosa pine tree in the mid-ground (a remnant of a former forest); dead western larch—killed by dwarf mistletoe and inter-tree competition—fir-engraver bark beetle mortality, and western spruce budworm defoliation in the foreground; and a glut of pole-size trees that have grown into the understory during a period of fire exclusion.

Photo 12. Disturbance restoration in action on Okanogan-Wenatchee National Forest
A crew marks small Douglas-fir and grand fir in-growth for removal.
Photo 13. Disturbance restoration in action

After mechanical treatment, many areas also need prescribed burns to reduce surface and ladder fuels in addition to reinitiating ecosystem process such as nutrient cycling. Repeated burning favors the survival of medium- and large-size trees with thick bark and that are well spaced from neighboring trees.

Photo 14. Succession restoration need in fire infrequent forests

Larger trees and more complex structure are needed in this Olympic National Forest stand.
Climate Change Context
This section highlights the predictions that “environmental suitability” for large wildfires is projected to increase and the potential effects of that change on forest vegetation and landscape resilience. Environmental suitability for large wildfires is basically the measure of if conditions exist to support large wildfires. It is based on conditions under which large wildfires have manifested in the past and may be a predictor of when and where they might occur in the future. These conditions are associated with fire season precipitation, maximum temperature, slope, and elevation. Davis et al. (2017) projected changes in environmental suitability for large forest fires over the 21st century, and their methods have been applied across the BioA area (figure 33). Environmental suitability for large wildfires is described as low, moderate, or high and is displayed historically (1980) and as projected through the end of the century (2100).

Modelling indicates that the area that is highly suitable for large wildfires will steadily increase through time in the southern and eastern portions of the BioA area. Furthermore, the greatest increases in relative change in suitability for large wildfires is projected to occur in the higher elevation Cascades (fire diverse [mixed-severity] ecosystems), Coast Range (fire diverse [mixed severity] and fire infrequent), northern Cascades (fire diverse [mixed severity] and fire infrequent) and northern eastern Cascades (frequent-fire dependent and fire diverse [mixed-severity]).

Although changes discussed in the climate section (including fire season precipitation, maximum temperature, slope and elevation, and water-balance deficit) do not necessarily mean that fires will burn in these areas more frequently, as that depends on ignition, they do provide us with a good index of how much change these forests are likely to undergo in the coming decades. This will likely be increased fire frequency and seasonality, and changes in severity and patch size, as well as synergistic changes related to drought stress.

Climate change is likely to increase native insect and pathogen activity and affect population dynamics and geographic distributions of pathogen and insect species. Warmer winters and more intense droughts are expected to enable insects to move into previously unsuitable habitat (Bentz et al. 2016, 2010). Drought and insects may also interact to further stress trees and predispose them to mortality.

Climate change is not considered in our forest plans and is a primary reason for the need to create higher levels of ecological resilience to stressors and options for resilient ecological pathways. Climate change may continue to alter the composition, structure, and function of forested and non-forested ecosystems in the BioA area (Vose et al. 2012). Tree mortality is expected to increase (Allen et al. 2010, 2015). Tree growth and viability will also be affected by warmer winters, earlier snowmelt, and changing water availability.

Climate change is likely to compound, increase, and expand the departure of ecosystems from resilient conditions in the immediate, mid-term, and long-term future. The landscape is becoming increasingly vulnerable to uncharacteristic vegetation shifts in the face of climate change. Changes in temperature, precipitation, and effects on natural ecosystem processes such as wildfire, insects, disease, and windstorms, in addition to complex interactions with invasive species will be compounding factors for increasing the departure of ecosystems from resiliency (figure 35).
Figure 33. Past and future fire suitability in the Bioregional Assessment area

Past and future projected change in fire environment and relative change in fire suitability as an index of changes to resilient reference conditions (adapted from Davis et al. 2017). The area that is highly suitable for large wildfires (as measured by fire season precipitation, maximum temperature, slope, and elevation) steadily increases through time in the southern and eastern portions of the BioA area (comparing 1980 conditions to 2100 panels 1 and 2). Furthermore, the relative change in suitability for large wildfires increases most greatly in the higher elevation Cascade Range (moister fire-diverse [mixed-severity] ecosystems), Coast Range, northern Cascades and northern eastern Cascades (panel 3).
Figure 34. Past and future suitability for large wildfires in frequent-fire dependent (A) and fire diverse (mixed severity) forests (B)

In frequent-fire dependent forests, the area highly suitable to large wildfires grows from 27 percent in 1980 to 50 percent by 2060 and 70 percent by 2100. At the same time the proportions of moderate- and low-suitability to large wildfire declines. In fire diverse (mixed severity) forests, the areas highly suitable to large wildfires grows from 1980 levels at just 6 percent through 2060 at 18 percent and 35 percent by 2100. In fire diverse areas with low suitability to large wildfire decreases from 63 percent of the landscape in 1980 to 47 percent of the landscape in 2060 to 35 percent of the landscape in 2100. These charts both indicate an urgent need to address landscape conditions in a changing climate to limit the increase in large wildfire suitability. For variability of trends see Davis et al. (2017).
Figure 35. Past and future suitability for large wildfires in fire infrequent forests

In fire infrequent landscapes the highly suitability environments for large wildfires gradually increases through time from 2 percent in 1980 to 8 percent in 2060 and 18 percent in 2100. For variability of trends see Davis et al. 2017.

Planning Considerations

Refer to BioA Chapter 2 Management Recommendations

Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.

Recommendation 2: Address the dynamic nature of ecosystems to be better positioned to respond to future environmental uncertainties.

Recommendation 6: Recognize that fire is a natural process and plays an important role in reducing the risk of uncharacteristic fire and in promoting ecosystem health.

Recommendation 7: Expand the use of timber harvest as a restoration tool to provide economic and social benefits to communities.

Recommendation 8: Shift from single-species management to maintaining and restoring habitat for multiple species to help ensure we are managing our ecosystem to be resilient in the face of future change.

Recommendation 9: Promote active management in plant and animal habitats to restore and promote ecological resilience.

Geographic Considerations

As described in this section, based on best available information, we anticipate the area that is highly suitable for large wildfires will expand in the southern and eastern portions of the BioA area, with the greatest increases in suitability for large wildfires in the higher elevation Cascades, Coast Range, and northern and northeastern Cascades.
Key Change Issue 2—Management Options are Incompatible With Ecosystems

Current management options in forest plans are not always compatible with the unique ecosystems across the BioA area. This incompatibility creates situations where managers are working to restore and protect forests without helpful direction from forest plans. There is a need to eliminate current constraints that are no longer aligned with best available science and management objectives.

Fire exclusion has resulted in the exclusion of natural fire across much of the BioA area, leading to unintended ecological consequences, but the restoration work needed to address the consequences is difficult to plan because forest plans limit options. Loss of old forest from high-severity wildfire has been concentrated in frequent-fire dependent ecosystems where large wildfires have occurred, resulting in localized loss of old forest (Davis et al. 2015, in progress).

The amount of tree-density reduction and prescribed fire in frequent-fire dependent ecosystems to reduce risk of old-forest loss has been less than anticipated (Spies 2009). These activities would promote resilience to both climate change and fire within and outside late-successional reserves (Spies et al. 2018).

The use of timber harvest as a tool to achieve desired ecological outcomes is not broadly or consistently included in current plan direction. Plans tend to focus on timber outputs rather than desired landscape conditions and therefore do not often provide helpful direction on how to connect the two when they are complementary.

Harvest restrictions on lands within the NWFP reserve network and amendments such as the Eastside Screens management direction, which restricts the size or age of trees harvested, are not consistent with management recommendations based on the best available science. In the case of the Eastside Screens, this can limit our ability to capitalize on using timber harvest as a tool to reach ecological integrity.

One example of management direction that is not necessarily aligned with current best available science is the 80-year exemption associated with NWFP late-successional reserves. Spies et al. (2018) state the following:

Somewhat arbitrarily, 80 years after conifer forest establishment has been used as the onset for “mature” (for example, OGSI 80) Douglas-fir forests…Eighty years was used as the threshold for late-successional/old growth in the NWFP (USDA FS 1994) because that is about the earliest time when such stands begin to resemble maturing forests in the moist forest (does not apply to the dry forest zone)…. The variability in structure with stand age indicates that at a regional scale, age or time alone is only a partial predictor of forest structure. The structural features of mature and old-growth forests would have included medium- to large-size (for example, greater than 40 inches) shade-intolerant tree species; smaller shade-tolerant trees of similar and lesser age in the mid to lower canopy layers; large standing and down dead tree boles; and horizontal and vertical structural heterogeneity of live and dead trees.
This exemption does not reflect our current understanding about many ecosystems within the NWFP area. In addition, under the NWFP, proposed forest management activities in stands older than 80 years in a late-successional reserve are subject to review processes, administered by the Regional Ecosystem Office, which tend to require more time and resources than activities in younger stands. Because it often requires higher levels of project management and project-related risk taking, needed restoration work in older stands may not be accomplished. In and of themselves, stand age or tree size are not the only, or necessarily the best, measures of ecological health. For example, forests more than 80 years old on lower productivity sites may be dense and lack understory diversity as a result of slow growth and development and exclusion of fires that would have contributed to structural variation; in this case, human intervention could help move the stand toward a more natural condition despite the age-based harvest limit. Consequently, needed restoration work in older stands may be hindered, or not accomplished. Looking across the diversity of forest habitats on the landscape, there is increasing recognition that stand age or tree size are not the only, or necessarily the best, measures of ecological health.

Another example is the 21-inch standard associated with the wildlife standards of the Eastside Screens. In this case, many relatively young shade-tolerant trees (for example, grand fir/white fir) that need to be removed to work toward ecological resiliency purposes are overabundant on the landscape, but cannot be cut under current plan direction (without a forest plan amendment) because they have grown beyond 21 inches diameter at breast height.

Outdated management direction can hinder our ability to reduce the risk of insect and disease mortality in late-successional reserves, scenic corridors, and habitat managed for deer and elk cover; this is another example of the need to consider updating plan direction to reflect the best available science (figure 36 and figure 37; map 11).

**Insect and Disease Context**

In 2012, insect and disease risk was assessed nationally. On the National Insect and Disease Risk Maps, “risk” or “hazard” are defined as the potential that, without remediation, 25 percent or more of the standing live basal area of trees larger than 1 inch diameter at breast height will die over the next 15 years because of insects and disease (Krist et al. 2014). This risk assessment was updated in 2017. This assessment has identified geographic areas where prevention opportunities exist in synergy with forest plans (figure 61 and figure 62). It also shows how prevention opportunities intersect with national forests in the BioA area and fire dependency categories (figure 60) (see the “Forest Plan Management Emphasis” section and map 7). Current and near future risk (2012 to 2027) of forest mortality from insect and disease agents is particularly high in frequent-fire dependent and fire diverse (mixed severity) ecosystems.
Figure 36. Insect and disease prevention need identified in 2017 as overlapping with Bioregional Assessment fire ecology categories and forest plan management emphasis areas

Forest health prevention needs about 4 million acres in frequent-fire dependent ecosystems and more than 2 million acres in fire diverse (mixed severity) landscapes. In fire infrequent landscapes, more than 600,000 acres need forest health prevention measures. Most of the need is in forest plan emphasis areas that have multiple objectives, including late-successional reserves, ungulate habitat, or viewsheds. See also the “Forest Plan Management Emphasis” section under “Sustainable Timber” in chapter 1.
Map 11. Intersection of forest health prevention needs with forest plan management emphasis areas within the Bioregional Assessment area

This map also displays all other forest health prevention needs outside of congressionally reserved areas. Risk, or hazard, in the National Insect and Disease Risk Maps (source of forest health prevention need) is defined as the potential that, without remediation, 25 percent or more of the standing live basal area or trees over 1 inch diameter at breast height will die over the next 15 years due to insects and disease (Krist et al. 2014). This risk assessment was updated in 2017. See also the “Forest Plan Management Emphasis” section under “Sustainable Timber” in chapter 1.
Late-Successional Reserve Context

Late-successional reserves were designed to protect and enhance old-growth forest conditions. In combination with other land use allocations and standards and guidelines, late-successional reserves are intended to maintain a functional, connected, late-successional and old-growth forest ecosystem and create a network of habitat for old-forest-dependent species, including the northern spotted owl (USDA FS and USDI BLM 1994: 6–8). By design, timber harvest for purposes other than the development and protection of late-successional reserves or riparian reserve objectives is restricted through plan elements. Some of those elements are highlighted below.

Today, late-successional reserves, particularly in dry/fire-frequent/fire-dependent landscapes, are at risk of not fully meeting the purpose and need outlined in the NWFP to, “maintain and enhance late-successional forests...that are retained in their natural condition with natural processes, such as fire” The desired conditions of late-successional reserves generally identified in current forest plans do not capture old-forest structure, species composition, or development potential, depending on site conditions and history (Pabst et al. 2008, Reilly and Spies 2015). Our knowledge of the role of fire has advanced considerably from the early 1990s. The lack of fire in drier forests types over the past century has created hazardous, unsustainable conditions and treatments of some kind are needed in many of these forest types, including reserve areas, to restore resiliency.
The 2012 planning rule introduces a new ecological objective: ecological integrity and related resilience. This vocabulary was not part of the direction for the NWFP. This element changes the frame through which we see late-successional forests and emphasizes putting forests in the context of their ecosystem function. For the NWFP area, this is a significant change, especially for frequent-fire dependent forests.

**Regional Ecosystem Office and Regional Interagency Executive Committee Review**

Under the NWFP, national forests are required to prepare an assessment for late-successional reserves, or groups of small reserves. The late-successional reserve assessment describes the late-successional reserves’ existing conditions, criteria for developing appropriate treatments, a fire management plan, and a monitoring plan. This document is then required to undergo review by the Regional Ecosystem Office (Regional Ecosystem Office) and Regional Interagency Executive Committee. Following the development and approval of the late-successional reserve assessment, proposed activities within the late-successional reserve must be consistent with the late-successional reserve assessment; if they are not, the late-successional reserve assessment or forest plan must be updated or amended, respectively, or a consistency review of the proposed activity is undertaken by the Late-Successional Reserve workgroup. Proposed activities that require a consistency review or amendment could include those designed for the beneficial creation and maintenance of late-successional forest in fire-dependent forests as well as activities that reduce the risk of large-scale disturbance. Depending on the complexity of the proposed activities, the late-successional reserve review process can add additional time to project timelines and may discourage national forests from proposing beneficial activities because of the time commitment. In recent years, the Regional Ecosystem Office and late-successional reserve workgroup have been working to expedite the amount of time it takes to conclude consistency reviews. The late-successional reserve Workgroup has identified many late-successional reserve assessments that may need to be updated to allow additional beneficial activities that would reduce large-scale wildfire risk and promote maintenance of late-successional forests. The standard for the consistency review is based on how proposed activities in late-successional reserves will protect or enhance old-growth forest conditions. This means that projects with multiple objectives (for instance, old-growth enhancement plus fire-risk reduction near the wildland urban interface) and projects that include portions of a late-successional reserve and other lands such as matrix, may not fully meet the standards set forth in the NWFP. In effect, late-successional reserves may be left out of otherwise “all lands” restoration projects and, if included, may add additional coordination requirements and increase the timeline for the completion of individual projects.

Under the 2012 planning rule, the overall ecosystem function as related to ecological integrity and resilience is emphasized where this concept was not part of the NWFP. This underpinning focus on ecosystem resilience may change how old-growth forest management goals and objectives are set in the future.
Chapter 3: Caring for the Land

80-Year Exemption: Late-successional reserve standards and guidelines empower the Regional Ecosystem Office to create exemptions related to review of vegetation management in late-successional reserves. Certain precommercial thinning and reforestation activities were exempted from the Regional Ecosystem Office review process in 1995 and, in 1996, some commercial thinning activities were exempted from review. This 1996 exemption from Regional Ecosystem Office review for commercial thinning in stands under 80 years of age has become known generally as the “80-year exemption.” This means that for stands over 80 years of age, there is still a Regional Ecosystem Office review process needed in most late-successional reserves. Proposed management needs to be consistent with the late-successional reserve assessment and the NWFP. There is an exception in that the North Cascades adaptive management area/late-successional reserve stand age is 110 years, and additional activities are allowed east of the Cascades and in the Oregon/California Klamath Province.

The intent of this age limitation was to serve as a rough proxy for old-growth stand structure and development potential and was based on expert opinion informed by observations of natural forests (Spies et al. 2018). The 80-year exemption is applied across the NWFP area (with the exception of the North Cascades AMA/late-successional reserve where it is 110 years). This is also associated with Regional Ecosystem Office direction limiting harvest of trees to less than 20 inches diameter at breast height in late-successional reserves. Late-successional reserve criteria stipulate that,

Individual trees exceeding 80 years in those provinces or exceeding 20-inches diameter at breast height in any province, shall not be harvested except for the purpose of creating openings, providing other habitat structure such as downed logs, elimination of a hazard from a standing danger tree, or cutting minimal yarding corridors (Knowles 1996).

The 80-year exemption, which applies to a diverse landscape, lacks the flexibility given our current environments and new insights around how old-forest composition, structure, and development proceed regardless of an age threshold. For example, the 80-year exemption applies to frequent-fire dependent, fire diverse (mixed severity), and fire infrequent systems although old forest develops and manifests very differently across all three of these categories.

New science about frequent-fire dependent and fire diverse (mixed severity) ecosystems (Spies et al. 2018) may suggest the need to modernize the 80-year exemption. Early-seral species such as white fir or grand fir have established primarily as a result of more than a century of fire exclusion; many of these trees are now more than 80 years of age. More trees regenerate through time and the ingrowth of these shade-tolerant and relatively fire-intolerant species has resulted in more acres of dense and multi-strata forests than historically existed (Camp et al. 1997, Hagmann et al. 2017, Merschel et al. 2014, Taylor and Skinner 1998, 2003). This changes the effects of fire and increases the risk of large patches of higher severity fire effects where historically they were not a major component (Spies et al. 2018). The 80-year exemption may be helpful to manage young, simplified stands, but the process associated with Regional Ecosystem Office review for management in stands more than 80 years of age may limit restoration needs in more complex stands.
Opportunities to use Timber Harvest as a Restoration Tool

Timber output predicted in forest plans does not necessarily incorporate modern desired outcomes for resilient landscapes, but opportunity exists to use active management, including timber harvest and prescribed fire, to achieve desired landscape and habitat conditions (figure 38). Implementation of ecologically appropriate landscape prescriptions based on resiliency, particularly in the face of climate change, could create a synergy where resiliency management and timber outputs both benefit. We acknowledge that even when the best available science indicates that active management for restoration and resiliency is needed, conflicting values around timber harvest or other types of active management still exist. During upcoming planning efforts, we will engage the general public, local and state governments, and American Indian tribes to help address this complex issue.

Figure 38. Merchantable timber acres in Northwest Forest Plan (NWFP) late-successional reserve and matrix

Approximate acres of merchantable size timber volume (more than 12,500 standing board feet per acre of standing tree volume) by forest and NWFP land use allocation: late-successional reserve and matrix. Not all forests have the same potential to sell additional timber volume based on future planning framework.

Eastside Screens Context

One component of the Eastside Screens that affects implementation of restoration and resilience projects and timber production is the restriction on harvest of trees greater than 21 inches diameter at breast height. The ecosystem standard requires a landscape-scale analysis of forest structure. If the results of this analysis indicate that there is a lack of late- and old-structure forest as compared with historic reference conditions (Scenario A), then no trees, regardless of species, greater than 21 inches diameter at breast height may be harvested within late- and old-structure forest. This component of the wildlife standard, intended to conserve large trees on a landscape where the largest trees were historically cut, can conflict with attaining desired conditions across the landscape. This is especially true where ingrowth of white fir and grand fir undermines the protection and culturing of ecologically desirable old trees and species fire tolerant species compositions in frequent-
fire dependent and fire diverse (mixed severity) ecosystems. Currently, the only pathway for harvesting younger, shade-tolerate conifers more than 21 inches diameter at breast height under Scenario A of the wildlife standard, is with a forest plan amendment. We acknowledge that even when the best available science indicates that removal of trees more than 21 inches diameter at breast height is needed for restoration and resiliency, conflicting values around timber harvest or other types of active management still exist. Upcoming planning efforts will engage the general public, local and state governments, and American Indian tribes to help address this complex issue.

Planning Considerations

Refer to BioA Chapter 2 Management Recommendations

Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.

Recommendation 2: Address the dynamic nature of ecosystems to be better positioned to respond to future environmental uncertainties.

Recommendation 6: Recognize that fire is a natural process and plays an important role in reducing the risk of uncharacteristic fire and in promoting ecosystem health.

Recommendation 7: Expand the use of timber harvest as a restoration tool to provide economic and social benefits to communities.

Recommendation 8: Shift from single-species management to maintaining and restoring habitat for multiple species to help ensure we are managing our ecosystem to be resilient in the face of future change.

Recommendation 9: Promote active management in plant and animal habitats to restore and promote ecological resilience.

Geographic Considerations

The geographic considerations associated with the aforementioned key change issue “Management Options are Incompatible With Ecosystems” vary by ecosystem and interaction with management direction and are therefore complex. The section above highlights some of these considerations given fire, active management for landscape resilience, and timber output.

Key Change Issue 3—Dynamic Ecosystems Need Dynamic Plans

Our current forest plans are relatively static, but our ecosystems are not. There is a need to create planning mechanisms that integrate uncertainty and reduce risk of ecosystem or management failure in the face of anticipated change. This could include creating scenarios or processes that are triggered under various circumstances. Integration of upland forests with riparian and aquatic systems when planning is also needed for the overarching resilience of forest ecosystems. Ecosystem resilience, along with many ecosystem processes such as regulation of structure and nutrient cycling via naturally occurring fire, landslides, or insect and disease activity and habitats, can only be evaluated at a landscape or larger scale.
This dynamic need is also tied to evaluating and managing invasive species, including plants, insects, diseases, and animals; invasive species are increasing in diversity and extent across the BioA area and will have increased capacity to change forest ecosystem function and processes in the future. Invasive species generally reduce ecosystem resilience, which is of heightened concern in the face of climate change.

**Planning Considerations**
Integrate modern management concepts and tools to help us manage ecosystems toward resilient states and reduce risks associated with ecological and social uncertainty.

*Refer to BioA Chapter 2 Management Recommendations*
Recommendation 2: Address the dynamic nature of ecosystems to be better positioned to respond to future environmental uncertainties.

Recommendation 3: Update and consolidate the existing aquatic direction processes and analysis requirements.

Recommendation 4: Reduce the introduction and spread of plant, animal, and other invasive species.

Recommendation 6: Recognize that fire is a natural process and plays an important role in reducing the risk of uncharacteristic fire and in promoting ecosystem health.

**Geographic Considerations**
Within the BioA area, the effects of climate change are anticipated to be the greatest in northern California, southern Oregon, the eastern Cascades, and high elevation zones.

**Key Change Issue 4—Ecosystem Restoration Tradeoffs**
There is a need to strategically consider that some ecosystems will be more difficult than others to restore, stabilize function in, or proactively prepare for climate change. Turning the tide of rapid change in some frequent-fire dependent and fire diverse (mixed severity) ecosystems will be difficult given our current toolbox and ecological realities. Major, often irreversible, changes include broad landscape species composition shifts (for instance, shade-tolerate seed banks in large high-severity fire footprints), swings toward non-forest (for instance, rapid succession of fire in chaparral), or transition into novel states or landscape patterns (for instance, large-scale mortality of old-forest and legacy trees from a combination of insect, disease, drought and uncharacteristic fire (Franklin et al. 2018, Johnstone et al. 2016).

Under our new “Anthropocene”\(^\text{12}\) epoch of human activity and climate change, “there are some historical ecological patterns and processes difficult or impossible to reestablish” (Corlett 2015 in Spies et al. 2018). The disproportionate increase in high-severity fire (without proportionate increases in low- and mixed-severity fire) in frequent-fire dependent and fire diverse (mixed severity) zones (Haugo et al. 2019) is contributing to the trend of moving frequent-fire dependent ecosystems toward novel alternative states (Lydersen et al. 2017, Prichard et al.2017), including non-forest and non-native (Hessburg et al. 2015).

\(^{12}\) Relating to or denoting the current geological age, viewed as the period during which human activity has been the dominant influence on climate and the environment.
Planning Considerations

Refer to BioA Chapter 2 Management Recommendations

Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.

Recommendation 2: Address the dynamic nature of ecosystems to be better positioned to respond to future environmental uncertainties.

Recommendation 4: Reduce the introduction and spread of plant, animal, and other invasive species.

For Further Consideration

Strategically consider that some ecosystems will be more difficult than others to restore or proactively prepare for climate change. Create regional and subregional strategies for ecological risk management. Balance investments for restoration and resilience management, given tradeoffs in value and opportunity cost related to (1) highly altered ecosystems, (2) ecosystems with inertia toward an altered state, and (3) ecosystems where there is opportunity and likelihood to create landscape resilience in the face of climate change.

Geographic Considerations

Fire infrequent forests are relatively functional and forest attributes that contribute to resilience are generally improving through time. These forests have a high capacity for carbon sequestration and the development of structural complexity. Currently, our most productive, fire infrequent and historically clear-cut forests (Coastal and Olympic) are still missing old-forest structural elements that are projected to develop through time and succession as trees grow larger. Within historic plantations, options to continue to enhance complex structural components, species composition or more complex landscape patterns through silvicultural methods are likely limited under current late-successional reserve management direction.

Frequent-fire dependent forest ecosystems do not currently reflect resilient reference conditions (East Cascades slopes and foothills, Klamath Mountains/California high north Coast Range) and are highly departed, given the metrics of forest structure, species composition, and landscape pattern. Frequent-fire dependent forests are becoming increasingly vulnerable to uncharacteristic vegetation shifts in the face of climate change. This includes the Cascades, Klamath, North Cascades, and portions of the Coast Range ecoregions. These areas are projected to continue to rapidly depart from resilient conditions because of interactions between fire exclusion, increased large fires with larger patches of high-severity wildfire, insects, disease, and changing climate.

There is a need to mitigate and adapt frequent-fire dependent landscapes to improve the ecosystem resiliency in the face of climate change. This could include assisted migration of plant species, creation of climate refugia, or landscape-scale planning for increased fire in transitional forest-grassland interface areas. In frequent-fire dependent ecosystems, multi-aged management, including the use of mechanical harvest, prescribed fire, and
wildfire,\textsuperscript{13} could help with retention of soil organic matter, microbial communities, and water storage to support long-term soil quality where shade-tolerant tree species dominate the landscape outside of reference conditions.

Fire diverse (mixed severity) forests are currently somewhat reflective of resilient reference conditions, but ecosystem function is becoming increasingly vulnerable to uncharacteristic vegetation shifts in the face of climate change—especially historically uncharacteristically large, high-severity wildfires. This is especially true in the Cascades, South Coast and portions of the north Cascades. Without proactive management to mitigate the effects of climate change, these national forests and grasslands will likely be measurably less reflective of resilient conditions soon because of interactions between changing climate, increased wildfire, insects, disease, and invasive species.

There is a need for strategic planning and ecosystem management direction related to proactively prepare for climate change. In fire infrequent forests there are special needs to manage in the face of projected increases in dramatic windthrow events, especially along the Washington coast. Across the BioA, there is a need to create strategic climate and fire refugia.

Soil types vary across the BioA, but in general, forest plan direction is similar across the whole area; it would be possible to address these needs at the BioA scale. There is variability, such as sandy or serpentine soils, which will need to be addressed on a sub-regional or forest scale.

\textbf{Background Information to Support Key Change Issues}

The BioA area is composed of highly diverse forest and non-forest ecosystems characterized by a broad variety of vegetation zones and species (map 12). In addition to providing a high-level characterization of the diversity of forests and ecosystem function and processes across this area, this document aims to describe how these forests have changed in the past 200 years, are changing now, and are projected to change into the future. Since European settlement, forest ecosystem conditions and functions have been considerably altered through a combination of fire exclusion, grazing, roadbuilding, invasive species, historic timber harvest, and other activities (DeMeo et al. 2018, Haugo et al. 2015, Keane et al. 2009, Landres et al. 1999, Morgan et al. 1994; Swetnam et al. 1999; Wiens et al. 2012). Change is a continuing theme for our BioA ecosystems. The gap between resilient conditions and existing conditions is projected to continue to widen in the future, primarily as a result of climate change, the legacy of fire exclusion, and the interactions of these with insects, disease, and invasive species (Davis et al. 2017, Spies et al. 2018; map 12).

\textsuperscript{13} Related terms: “let burn,” “natural fire management,” prescribed natural fire”; “wildland fire use for resource benefit,” also referred to as WFU or fire use (1995). Any natural fire could be managed for multiple objectives (2009)—“appropriate management response fires,” “multiple objective fires,” “managed wildfire.”
Map 12. Broad and diverse vegetation zones within the Bioregional Assessment area

These vegetation zones are named for the tree species representing land potential to support ecosystems and produce resources. The categorization serves as a framework for the intersection of climatic and productivity gradients across the landscape, including disturbance regimes (notably fire). Also represented on this map are Environmental Protection Agency class III ecoregions and Northwest Forest Plan physiographic provinces.
Photo 15. Complex early-seral forest after a stand-replacing fire in the Metolius Basin, Deschutes National Forest

Early-seral species that established after fire include manzanita, ceanothus, and conifers, such as ponderosa pine. Also pictured are fire-created snags and residual trees that survived the fire.

Photo 16. Forest in winter on Mt. Bachelor, Deschutes National Forest
Chapter 3: Caring for the Land

Photo 17. South Fork Skokomish Trail on the Olympic National Forest

Photo 18. A meadow below Butler Butte lookout on the Umpqua National Forest
Call-out box 9. Non-forested lands

Non-forested ecosystems are found across the Bioregional Assessment area, and those landscapes provide resiliency, habitat, and a variety of culturally significant species. Major herbaceous and shrubland vegetation include grasslands, meadows, wetlands, vernal pools, chaparral, coastal and mountain scrub, sagebrush, alpine, and sparse vegetation. These systems support high levels of diversity, provide habitat for wildlife, recreational opportunities, and important ecosystem services. Non-forested lands should be managed to maintain or improve wildlife habitat and rangeland conditions based on ecological parameters. Studies show that in many of our drier landscapes we have less non-forest and more forest than historically. In some cases, managing for national range of variation could mean reducing forest cover. There is a potential need to address desired ecological conditions in forest plans for terrestrial non-forested lands. “Essentially all natural temperate forest landscapes include areas that are not currently dominated by the tree life-form, although these historically may have varied widely in nature and extent with the forest region and disturbance regime” (Franklin et al. 2018).


Potential Vegetation

Potential vegetation serves as a useful integrator of climate, geomorphology, and soils to describe land capability to support ecosystems and associated resources. Whereas existing vegetation can vary with age and disturbance, potential vegetation serves to capture and describe the endpoint of where the sere (set of ecosystem stages over time) ends up. The modern view of potential vegetation focuses less on the ideal climax vegetation at the end of the sere, and more on land productivity, hydrologic and nutrient cycles, and disturbance regimes.
Natural Range of Variation
We can estimate if a landscape is resilient by comparing current forest structure, species composition, and landscape pattern to resilient reference conditions. Natural range of variation and historic range of variation are different names used for reference conditions. Understanding reference conditions is critical to putting resilience into management practice. For example, natural range of variation could provide percentages for how much old, mid-aged, young, and early-seral forest might be present on a resilient landscape. Reference condition values are usually predicted through modeling.

One of the most important applications of potential vegetation in recent years has been to serve as the framework for disturbance regimes, notably fire. Fire regimes are characterized by a recurring pattern of fire frequency and severity that can be described and quantified. For example, ponderosa pine ecosystems (potential vegetation) historically featured frequent low-intensity fire. In contrast, wet western hemlock potential vegetation is associated with infrequent but high-severity fire. Mixed-conifer vegetation is associated with a mixed-severity fire regime and has proven challenging both to describe and to manage on our landscapes. Through modeling techniques using tree-ring and other data, we can estimate what the range of historic fire frequency and seral stage abundance was for each potential vegetation type. By comparing this with current seral stages, estimates of ecological departure from natural range of variation can be developed. This becomes an important tool in planning restoration needs.

Past and Current Departure from Resilience
Naturally sustainable forest conditions, or natural range of variation, have been estimated for our forests. Departure or distance from this natural range of forest conditions is a measurement of how far and in what direction forests have moved away from resilient states.

A simplified analogy to help describe forest ecosystem resilience would be to the human body and overall health. When humans become weakened internally or externally, we become less resilient and less able to recover from disease, accidents, or normal stresses in our daily lives; we may become sick. This is akin to loss of forest ecosystem resilience. Forests ecosystem function can be weakened in a variety of ways and forests become less able to recover from forest disturbances, such as insects, disease, wind, and wildfire that forests would normally be able to bounce back from. Stresses that are a normal part of everyday forest life now can trigger ecosystem state changes. The healthy states defined in the natural range of variation were sustained for thousands of years and therefore give us a scientifically sound foundation for managing our forests toward more healthy and sustainable conditions into the future. The information in this section provides an ecological basis for how and where management could create more healthy forests. This ecological basis does not consider social acceptability as discussed in the introduction of this chapter.

Many ecosystems are currently departed from resilient conditions, and the gap between resilient conditions and existing conditions is projected to continue to widen in the future because of a variety of factors, including fire exclusion, the effects of legacy timber harvest and grazing, and climate change, and the interactions of all these with insects, disease, and invasive species. Shifts in forest structure and species composition and changes in disturbance processes have resulted in forests with too many trees, an unstable increase in
shade-tolerant tree species, and systems that are out of sync with current and projected fire, insects, and disease conditions. The direction of change in a given ecosystem varies by primary drivers and productivity factors. Both restoration activities and succession support could be applied to move ecosystems toward more resilient reference conditions.

Current definitions of old forest are not reflective of the broad diversity of forests within the BioA and are not supportive of managing for resilient landscapes. Desired conditions (old-forest definitions) identified in existing land management plans are not fully aligned with resilient historical and future projected structure and composition, especially in systems that have been altered by fire exclusion, which include fire-dependent and more mixed fire diverse (mixed severity) landscapes (Spies et al. 2018).

**Forest Structure**

The term “departure” is used to describe how different current forest structural conditions are from historical or reference conditions. For this discussion we rely on standard Fire Regime Condition Class metrics (Barrett et al. 2010) to describe the natural range of variation. Current conditions that are less than 33 percent different than reference conditions considered within the natural range of variation. Conditions that are 33 to 66 percent different from reference conditions are considered moderately departed, and conditions that are more than 66 percent different than reference conditions are considered highly departed.

Current (2017) departure of forest structure from natural range of variation has been measured for the BioA. In addition, the need for forest restoration has been characterized. Restoration need has been described as (1) a need for succession restoration (that is, time to grow larger trees) and (2) a need for disturbance restoration (that is, reducing the number of trees via thinning or fire). Approximately 9.56 million acres in the BioA area currently need some type of restoration. This analysis is based on a comparison of current conditions to natural range of variation from LANDFIRE biophysical settings. About 7.4 million acres need some type of fire or mechanical treatment to reduce the number of trees; about 3 million of these acres also need succession restoration to grow larger trees. About 2.2 million acres within the BioA area currently need only succession restoration because the forests lack complexity and the trees need to grow larger.

Of the 7.4 million acres across the BioA area that need some type of disturbance restoration, about half are located within land management areas that restrict the implementation of mechanical forest restoration activities. About 46 percent of the identified area that needs disturbance restoration is in multiple-objective management emphasis areas, 20 percent is in forest plan timber production emphasis areas, 19 percent is in preservation, and 15 percent is in inventoried roadless areas. Of the acres in multiple-objective areas, about 470,000 are in riparian reserves.

Of the 2.2 million acres across the BioA area that need only succession restoration, 45 percent is in multiple-objective management emphasis areas, 20 percent is in preservation emphasis areas, 18 percent is in forest plan timber production emphasis, and 17 percent is in inventoried roadless areas.
Overall, frequent-fire dependent forests are consistently in need of disturbance restoration and are relatively highly departed from resilient conditions. About 4.8 million acres of the fire-dependent forests within the BioA area need some type of disturbance restoration and about 900,000 acres are also in of succession restoration. Fire-diverse systems are currently less departed, with only 1.7 million acres of departure needing some type of disturbance, and 650,000 acres needing succession restoration.

It is important to note that this departure analysis is an estimate that is likely underestimating departure in moister fire diverse (mixed severity) systems. Spies et al. 2018, find that, “Historical fire occurrence in these regimes (mixed severity) varied at centennial scales with climate and human population density (for example, Weisberg and Swanson 2003). Thus, given the occurrence of warm, dry conditions during much of the contemporary fire period, a rotation exceeding the upper end of the range suggests we are currently experiencing much less fire than would have occurred historically under a similar climate” (Spies et al. 2018 Vol. 1 p.139).

Vegetation Species Composition
Species composition of the forest, including the dynamic interactions between species composition and structure, are integral to gauging and managing for landscape resilience (Hessburg et al. 2016, Tepley et al. 2013). We know that species composition in fire-dependent landscapes historically were frequently disturbed and have had major shifts in species composition since the implementation of fire exclusion. For example, ponderosa pine, which thrive with and depend on frequent, low-severity fire, have declined throughout their range due to fire exclusion, while more shade-tolerant species such as white fir, grand fir, and Douglas-fir have expanded.

In moderately frequent or fire diverse (mixed severity) zones, species composition has shifted less dramatically since historical times, but has great potential for change owing to the inherent productivity of these areas. In moister fire diverse (mixed severity) landscapes, the effect of fire exclusion has been to reduce the amount of early-successional vegetation and change the quality and distribution of mid-successional forests that otherwise would now exist on the landscape. This has changed overall landscape-scale heterogeneity in forest composition, structure, and patch sizes (Haugo et al. 2019, Spies et al. 2018, Tepley et al. 2013).

Mixed-severity fires burning at intervals of 50 to 200 years would have created a mosaic of forest successional stages, including multi-cohort, old-growth stands. Differences and complexity inherent in mixed-fire frequency and severity have contributed historically to creating multiple development pathways for forests (Harvey et al. 2014, Stevens-Rumann et al. 2018). But because humans have altered fire frequency, seasonality (Balch et al. 2017) and severity, vast changes in ecosystem function and vegetation have also occurred (Haugo et al. 2019). This means that, in fire diverse (mixed severity) zones, fire exclusion has and will continue to reduce the variation of not only stand structure but also in the species composition and the dynamic roles these elements play in creating resilient forests now and into the future (Tepley et al. 2013).
Species composition is also critical when considering resilience to insects and disease. Species composition, in concert with forest density and structure (canopy layers), is the main driver of forest mortality from insects and disease. Species composition has shifted through time because of fire exclusion, extensive grazing, and selective harvest. Species such as Douglas-fir, grand fir, and white fir have moved down slope and into areas where their densities were historically kept low (large, fire-resistant individuals cultured) in a predominately ponderosa pine system (Hagmann et al. 2013, Heyerdahl et al. 2019).

**Ecological Inertia**

The inertia of species composition is also to be considered in the context of managing for resilient landscapes. Dry forests that have already had significant shifts in species composition are showing signs of changed successional pathways toward the regeneration of more shade-tolerant, less fire-adapted species. Under the new Anthropocene epoch of human activity and climate change, “there are some historical ecological patterns and processes difficult or impossible to reestablish” (Corlett 2015 in Spies 2018). “For example, field observations suggest that after recent wildfires, instead of regenerating to ponderosa pine or western larch, some areas now quickly regenerate to Douglas-fir and white fir, grand fir, subalpine fir, or lodgepole pine, despite intentional efforts (which often fail unless done well) to reestablish ponderosa pine or larch. The presence of abundant seed from shade-tolerant tree species (for example, firs) provides this inertia (Stine et al. 2014: 140).” Additionally, Franklin et al. (2018) give examples of frequent-fire forest ecosystems that, when disturbed after a long period of alteration (fire exclusion), are no longer able to return to their former state (Franklin et al. 2018: 352 and 378).

This is also the case with compounding disturbances (such as reburns) (Johnstone et al. 2016). Spies et al. (2018: 139) conclude that this “trend is likely to continue unless climatic changes alter the disturbance regime and the growth or survivorship of tree species.” Additionally, the disproportionate increase in high-severity fire (without proportionate increases in low- and mixed-severity fire) in frequent-fire dependent and fire diverse (mixed severity) zones (Haugo et al. 2019) is contributing to the trend of moving fire-dependent ecosystems toward novel alternative states (Lydersen et al. 2017, Prichard et al. 2017), including non-forest and non-native (Hessburg et al. 2015, Millar 2015).

**Landscape Pattern**

The pattern of patches of forested areas, along with their size, arrangement, juxtaposition, edge, and other descriptors, is another critical element to consider when measuring and managing for landscape resilience. Landscape patterns are made up of combinations of structure and species composition. The scale at which pattern is measured is important and ranges from fine-scale (tree clumps) patches or stands, neighborhoods, and landscapes (Hessburg et al. 2015). Landscape patterns can be influenced through use of mechanical and nonmechanical silvicultural systems, methods, and tools, including prescribed and natural fire, even- and uneven-aged silviculture and intermediate methods, such as variable-density thinning.
Focusing on landscape patterns as we develop strategies to increase resilience ensures that our efforts are consistent with the best available science, which supports increasing heterogeneity and forest structure and composition at multiple scales (Hessburg et al. 2016). Across the landscape, there appears to be a correlation between the loss of pyrodiversity (diversity of fire size, frequency, and severity) and loss of habitat, biodiversity, and ecosystem resilience (Hessburg et al. 2016, Kelly and Brotons 2017, Perry et al. 2011, Spies et al. 2018; Tingley et al. 2016). Management of landscape patterns also has implications for wildlife habitat connectivity and juxtaposition (distance between suitable habitat patches) (Cansler and McKenzie 2014, Kolden et al. 2015).

Frequent-fire dependent ecosystems tend to have homogenization at the finest scales and largest scales, including trees clump and patches due to ingrowth and densification (Churchill et al. 2013, Fry et al. 2014, Lydersen and North 2012) and the impact of large wildfires (Spies 2018, Stine et al. 2014).

Patches in fire diverse (mixed severity) landscapes are naturally complex; however, much of this complexity has been lost through fire exclusion, historic grazing, and timber harvest (figure 39). Fire diverse (mixed severity) zones have experienced recognizable homogenization at all scales, but most recognizably at the stand, neighborhood, and landscape scales because mixed-severity fires operate at all these scales (Tepley et al. 2013). Disease and insects tend to operate at the finer sub-stand and stand scales when at endemic levels. Therefore, the effects of fire exclusion in fire-diverse forest tend to be less recognizable, but the effects are important. Unsuppressed fires of the past would have created more of a landscape mosaic of forest successional stages than exists today.

Although some stand level diversity exists in moister forests (for instance, clear-cuts from the 1980s next to older managed or unmanaged forest), these are not the same patterns or scale of disturbance that would have occurred historically. Fire exclusion has reduced landscape-scale heterogeneity in forest composition, structure, and patch sizes. This very complex mix is not easily measured or recognized. But, if we are to estimate the departure from resilient conditions, changes in fire diverse (mixed severity) landscape patterns needs to be acknowledged as divergent, where the changes have created less than desired levels of heterogeneity, and therefore resilience, currently and into the future.
Figure 39. Patches in fire diverse landscapes

*Figure 39. Patches in fire diverse landscapes are complex. This complexity has been lost through fire exclusion. This example is from a Douglas-fir and western hemlock landscape in the western Cascade Range of Oregon. Black = a high mortality area (greater than 70 percent), vertical lines = moderate mortality (30 to 70 percent) and stippled = low mortality areas (less than 30 percent). From (Morrison and Swanson 1990).*

**Defining Old Forest**

The interplay and connection between how we plan for and manage old forest and how we define and discuss old forest is important, especially when considering a definition for resilient old forest.

For example, the NWFP standards and guidelines provide the example that western Oregon Douglas-fir stands may begin the mature phase of stand development in around 80 years (p B-2) but offer little specificity on how to measure old forest among diverse physiographic provinces and ecosystems. Other general elements include live old-growth trees, snags, fallen trees, multiple canopy layers, small understory trees, canopy gaps, and patchy understories, with the caveat that all these characteristics will vary across vegetation types, disturbance regimes, and developmental stages. Distinctions between “older forests,” “late-successional,” and “old-growth forests” in forest plans, amendments, and subsequent guidance are also layered with either complexity or ambiguity.

Descriptions and definitions of old-forest structure and composition continue to be detailed in policy and scientific literature (Franklin et al. 1986; USDA FS 1989, 1992, 1993; Dunbar-Irwin and Safford 2016; Franklin and Johnson 2013; O'Hara et al. 1996; Safford and Stevens 2016; Spies et al. 2006; Youngblood et al. 2004). Since the implementation of the NWFP, our understanding and ability to measure and describe old forest and distinguish between the many types of old forest across the landscape have progressed, but most current definitions are still not reflective of historical structure and composition, especially for fire-dependent landscapes (Spies 2018).
The Old Growth Structural Index (Davis et al. 2015) has improved our ability to measure old forest; however, it is based on current forest conditions, not necessarily old forest as it would have looked before historic harvest, grazing, and fire-exclusion effects (figure 40, figure 41). Further work is needed to be able to measure and implement restoration of functional old forests across wet as well as frequent-fire dependent landscapes.

Call-out box 10. Old forest

The Northwest Forest Plan Science Synthesis defines old-growth forests based on live and dead structure and tree species composition. Old-growth forests in the plan area differ with age, forest type, environment, and disturbance regime (Reilly and Spies 2015, Spies and Franklin 1991). The variability and complexity of site conditions, forest succession, and disturbance processes make defining old growth difficult or impossible under a single definition. Current definitions of old forest also are not reflective of resilient conditions. This includes misalignment of desired conditions (old forest definitions) with resilient historical and future projected structure and composition, especially in frequent-fire dependent and fire diverse (mixed severity) landscapes. "Old forests are inextricably intertwined in space and time in a continuum of forest development, just as young, mature, and mixed-age forests are. Focusing on only one part of the continuum is like trying to understand light by examining only one color or wavelength, or like trying to understand a river by looking only at the deep, quiet pools and ignoring the rapids (Spies 2009)."

Photo 20. Old-growth forest on the Gifford Pinchot National Forest
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Photo 21. Old-growth forest on the Deschutes National Forest

Figure 40. Old-growth structural index (OGSI) components

These components include large live trees, snags, down wood and diameter diversity.
Figure 41. Forest types, old-growth structural index, age and compositional elements

These figures display how different forest types (here western hemlock and grand fir/white fir) differ in forest compositional elements, including diameter diversity, down wood, snags, and live trees at the different age thresholds of old-growth structural index and age including 80 and 200 years.

Other Vegetation Conditions, Including Quality Early Seral

Considering early-seral forest and habitat, along with all the other classes of forest structure and development, is part of whole ecosystem management (Spies et al. 2018). The NWFP Record of Decision (USDA FS and USDI BLM 1994: 47) points out that in 1994, there was more early-successional habitat within the NWFP area than at any other time in history, and did not necessarily create mechanisms for creating more, especially on federal lands, as early-seral conditions were assumed to be created on non-federal lands through timber harvest practices. An assessment of landscape-scale restoration needs in the BioA area indicated that from a restoration perspective, there is not necessarily a regional need for development of early-seral forest, as there is an abundance of relatively simple early-seral forest on private lands (DeMeo 2018: 26).

However, there is a need for “complex” early-seral forest with higher ecological and habitat value. Complex early-seral habitat has been examined by Swanson et al. (2011), and in essence, complex early-seral forest differs from more simplified early-seral forest. Complex early-seral forest is often naturally occurring; it has high species diversity and is made up of legacies, including live and dead trees that provide habitat for surviving and colonizing organisms. Traditional forestry practices such as clear-cutting, salvage logging, and tree planting can reduce species richness and key ecological processes associated with complex early-seral habitat (Swanson et al. 2011). Phalan et al. (2019) found that early-seral habitat has remained stable overall on federal lands but declined on the Coast Range and in the Cascades ecoregions.

Old-forest conservation on its own will likely not be the only answer to conservation of old-forest habitat and there is a need to take a broader look at the balance of all structural and seral stages, their arrangement, and projected persistence through time (Spies et al. 2018).
Call-out box 11. Complex early-seral forest

While complex early-seral habitat was not a focus for the development of the standards and guidelines for the Northwest Forest Plan, there is now a compelling amount of scientific research documenting the importance of complex early-seral habitat for a variety of plant and wildlife species (Swanson et al. 2011, 2014; Phalan et al. 2019). Complex early-seral habitat is most often derived from stand-replacing forest disturbances, including wildfires, severe insect infestations, volcanic eruptions or extreme weather events, that leave behind residual large-diameter green trees, snags, and an abundance of downed wood, which are soon joined by understory grasses, hardwoods, shrubs and flowering plants (photo 22).

Photo 22. Complex early-seral habitat created by fire along the Fish Lake Trail 1570 on the Umpqua National Forest
Insects and Disease

Insects and disease are part of functioning ecosystems and can also be helpful indicators of ecosystems that are becoming less functional. Population dynamics and feedback loops between various disturbance agents like root diseases and bark beetles play an essential role in regulating, shaping, and driving our forested ecosystems. Insects respond when vegetative conditions move away from resilient conditions; levels of insect- and disease-caused mortality outside of natural ranges of variation or expected norms can indicate that forest structure, density, species composition, and landscape patterns are divergent from resilient conditions.

In combination with changing climate and wildfire, insects and disease will likely have changing effects across the region, especially where species like mountain and western pine beetles can broadly affect old forest by killing the largest relict trees. Nationally, Oregon tops the list in both absolute acres and proportion of risk (Idaho and Montana are also in the top three states). Primary agents of risks in the BioA area align in fire-dependent forests and include mountain and western pine beetles, along with a combination of root diseases, Douglas-fir beetle and spruce budworm (Krist et al. 2014).
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**Insects, Pathogens, and Climate Change**

Biotic disturbances, such as insects and pathogens, can generally increase stand-scale mortality and may erupt into epidemic outbreaks with high levels of tree mortality (Raffa et al. 2008). They do not always result in immediate mortality, but instead depress tree growth and vigor and make trees less resistant to wind disturbance, predisposing them to stem breakage (Hansen and Goheen 2000, Larson and Franklin 2010, Manion 1981). Although mortality rates associated with insects and disease are generally much lower than those associated with fire in this region, insect activity has resulted in more loss of live carbon and canopy mortality than has fire in recent years at the scale of the BioA (Berner et al. 2017, Hicke et al. 2016, Reilly and Spies 2016).

Native insect and pathogen activity is likely to increase as trees experience more growing season drought; however, the magnitude of effects will likely vary geographically as well as among species (Chmura et al. 2011, Kolb et al. 2016, Sturrock et al. 2011). In addition to affecting host species, climate change will also affect population dynamics and geographic distributions of pathogen and insect species. Increases in insect activity are driven by drought and extreme weather events. Pathogen activity is likely to increase in areas where they typically infect drought-stressed host species, while the effects of climate change on pathogens that proliferate under moist conditions may be more variable and difficult to predict (Sturrock et al. 2011). Warmer winters and hotter droughts are expected to enable insects to move into previously unsuitable habitat (Bentz et al. 2010, 2016). Drought and insects may also interact to further stress trees and predispose them to mortality. Several nonnative pathogens and insects are of particular concern in the BioA area (see Reilly et al. 2018).

In 2012, insect and disease risk was assessed nationally. Risk, or hazard, in the National Insect and Disease Risk Maps is defined as the potential that, without remediation, 25 percent or more of the standing live basal area of trees over 1 inch in diameter will die over the next 15 years due to insects and disease (Krist et al. 2014). This risk assessment was updated in 2017 to reflect tree mortality in the current risk assessment period (2012–2027); the Forest Health Assessment and Applied Sciences Team recently updated the 2012 risk map to indicate areas where mortality has been averted through forest management, has already occurred, is ongoing, or has yet to be observed, associating those latter categories with opportunities for restoration, exclusion and prevention.

For the BioA area, we have identified areas where prevention opportunities exist in synergy with forest plan management emphasis categories and how prevention opportunities intersect with national forests and grasslands in the BioA and fire-dependency categories.
Call-out box 12. Non-native forest invasive insects and diseases

“Nonnative invasive plants, insects, and disease can have major economic and ecological effects on forests. …there are several species of plants and pathogens that are having or could have significant impacts on forests within the NWFP [Northwest Forest Plan] area” (Spies et al. 2018). Invasive diseases and insects with significant effects on forests of the BioA area include pine blister rust, Port Orford cedar root disease, sudden oak death, balsam woolly adelgid, and emerald ash borer. Adaptive land management direction has allowed for forests to address these changing threats to individual trees and forested areas across the Bioregional Assessment area. Including similarly adaptive planning components in future land management plans can help forests address current evolving and as-of-yet unknown risks.

Harvest Methods Context

Timber harvest history has dramatically shifted over the life of the NWFP (photo 23, photo 24, photo 25). Timber sold volume levels on Forest Service lands within the NWFP area have totaled about 456 MMBF over the past 10 years (2009–2018), and 381 MMBF over the previous 10 years (1999–2008), with a low of 213 MMBF in 1995. The type of harvest method, or silvicultural method, has also shifted dramatically. The use of clear-cut regeneration harvest has been virtually eliminated.

Over the past 20 years, harvest methods have shifted primarily to commercial thinning. This is an intermediate type treatment meant to grow larger trees and increase stand health, but not to regenerate new trees. From a forest management perspective, this shift toward commercial thinning meets stand-level needs to grow larger trees; preserve and promote old-forest architecture and components; and, where appropriate, reduce stand stocking levels to promote forest health, individual tree growth, and reduce fuels. This shift has worked in many areas to both meet ecological needs and sustain the timber industry.

Photo 23. Historic clear-cut

This historic 1990s clear-cut in a mixed-conifer stand in California shows the shaded southern edge (left side of photo) of the clear-cut contrasted by the well-lit interior and northern edge.
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Photo 24. Variable density thinning
A current example of ground-based variable density thinning on the Mt. Baker-Snoqualmie National Forest.

Photo 25. Prescribed fire-treated stand on the Fremont-Winema National Forest
Call-out box 13. Evolving timber harvest methods

Forest Service harvest methods have shifted from primarily clear-cutting in the 1980s and early 1990s to mainly commercial thinning for the past 20 years. In the 21st century, ecological forestry has become central to how foresters approach landscape planning. This includes the use of natural forest ecological models and natural forest development and disturbances such as fire in project design and harvest methods.

More diverse commercial thinning harvest methods such as variable density thinning; individuals, clumps and openings; multi-aged management; and variable retention harvest have evolved since 1994 and become broadly studied and implemented. The Forest Service is dedicated to supporting the study and use of these more diverse silvicultural techniques and systems to meet today's need for resilient landscapes and multiple land management objectives.

Ecological forestry, multi-aged management, variable retention and other modern harvest methods are likely our best options moving forward, especially in the face of climate change and social and ecological uncertainty.

“Silviculture carries the implication of active rather than passive human participation in the initiation and development of forest ecosystems. This has probably never been more appropriate than in the 21st century when humankind has altered so many of the fundamental conditions under which forest ecosystems have evolved. We believe forests in this century will often require human participation to assist them in their continued adaption to shifting environments and disturbance regimes (Franklin et al. 2018: 92).”

The summary here is based on definitions of silvicultural methods as historically and currently defined by the U.S. Forest Service Forest Activities Tracking System database (FACTS). As Franklin et al. (2018: 91) state, “Traditional terminology often fails to convey clearly the treatments that are being done, because over time, these terms have been used in many diverse ways.” Although the Forest Service works to track activities within a historical database, one of the drawbacks of this method is that the details of silvicultural prescriptions are not necessarily conveyed. For example, variable density thinning is likely categorized simply as commercial thinning, although various levels of retention and removal occur across the broad range of prescriptions implemented in various ecosystems (figure 42 and figure 43). As Forest Service silviculturists have moved into the post-NWFP era, landscape ecology has become central to how foresters approach landscape planning; land managers now use ecological models from natural forest systems as a basis for managing forests, incorporating principles of natural forest development, including the role of natural disturbances, in the initiation, development, and maintenance of forests and forest landscape mosaics (Franklin et al. 2018: 92).

Desired ecological outcomes are unlikely to be met with commercial thinning alone. Generally, commercial thinning includes keeping the dominant trees in the stand, and may integrate forest health evaluations; however, intermediate thinning prescriptions do not include regeneration of any tree species as a primary purpose. This means that commercial thinning is an incomplete method for changing the dominant species composition of a forest in the long run, especially where more drought-tolerant and fire-adapted species are desired. For example, many forests that used to be dominated by ponderosa pine are now dominated by white fir or grand fir. Thinning in these forests will not necessarily recruit
ponderosa pine back into the forest. If regeneration of trees does occur, the method promotes shade-tolerant (usually fire-intolerant) species. Even in variable density thinning, small created openings are not planned to necessarily regenerate a new cohort of trees, rather the purpose is primarily spatial heterogeneity. This means that when variable density thinning is applied as a part of a Forest Service silvicultural system, it is planned as an intermediate treatment and the tending and care of regenerating trees is not considered for the short or long term.

Regeneration of trees, especially sun-loving and relatively drought-tolerant Douglas-fir (in wetter sites), ponderosa pine and western larch, will not be successful with commercial thinning. Additionally, commercial thinning is meant to fit within a larger plan for how to work with forest succession. Since little mechanical regeneration of forests has occurred since about the mid-1990s, wildfire is currently the primary regeneration agent. Wildfires, especially uncharacteristically severe wildfires, often create more extreme regeneration conditions than would otherwise occur. This includes highly altered microclimates for tree regeneration and distance to seed sources, which may result in alternate pathways of forest development that may not be optimal, particularly in the face of climate change.

From a timber production and ecological restoration standpoint, relying solely on commercial thinning is also unlikely to be sustainable into the future, especially given existing limitations on vegetation management. The board-feet-per-acre volume production for a thinning from below (commercial thinning) project is generally much smaller than for another harvest method, such as single tree selection, group selection, or shelterwood harvest. When thinning from below, the smallest diameter trees are generally harvested. It takes many small trees to equal the volume of one larger tree, and the product may be completely different (for example, pulp or biomass for small trees and sawlogs for larger trees). The balance between logging operation costs, the value of the timber product, haul distance to mill, and market usually creates a situation in which smaller diameter tree thinning tends to be economically borderline. Depending on market values, timber sales can provide a profit or be a deficit. When even just a few larger trees are included in sales such as these, volumes and marketability increase, creating more stumpage value in the timber sale.

Skog et al. (2006) show that prescriptions that are not thinning from below can reduce the subsidy required for treating forests, increase the volume in the local market, and increase the value (sawlogs) of sold material. This increased public harvest volume benefits consumers, loggers/buyers, and mills, although private landowners may experience lower stumpage prices. Prescriptions that go beyond thinning from below can improve the value-to-cost comparison for a treatment and more volume can be produced. More diverse prescriptions that include tree species selection, gap creation, or various levels of regeneration such as single-tree selection, group selection, or shelterwood harvest can make some sales more viable and could benefit local wood processing infrastructure and workforce.

Furthermore, as forests continue to grow, the management direction in our forest plans, including age limitation on harvest within late-successional reserves (80-year exemption) and Eastside Screens 21-inch-diameter at breast height harvest restriction in old-forest-deficient landscapes, are increasingly broad. At a regional scale, these rules may not allow
implementation of best available science on the forests for which they were designed. This is especially true when we consider desired conditions related to species composition, density, structure, and landscape pattern in old forests, and old-forest resilience through time. For example, shade-tolerant trees that are a product of fire exclusion in Eastside Screen areas will continue to grow larger than 21 inches diameter at breast height but would not be harvestable under current plans without forest plan amendment.

The integration of approaches such as ecological forestry (Franklin et al. 2018) and tools like multi-aged management harvest into the forest plans would help create modern desired outcomes across the landscape. Continued reliance solely on traditional intermediate commercial thinning at the current rates across the BioA area limits options for the creation of resilient landscapes through time. This could be especially valuable in frequent-fire dependent and fire diverse (mixed severity) ecosystems, where multi-aged silviculture and ecological forestry offer options for successfully coproducing timber and positive timber sale revenues where inherent productivity may limit the volumes produced per acre under traditional intermediate thinning practices.

![Figure 42. Acres of the primary harvest types within the Bioregional Assessment area from 1947 to 2017](image)

_Harvest types include clear-cutting, commercial thinning, salvage, sanitation, and single-tree selection. Peaks of activity include more than 75,000 acres of clear-cutting in 1988. In 2007 there was a peak in commercial thinning at just less than 82,000 acres. Timber harvest practices have moved to mostly commercial thinning, with almost no stand clear-cutting. Even with area harvested being similar to that in the 1970s, the volume is much lower due to the differing harvest types._
Knowledge Gaps

Forest ecology, particularly landscape ecology, is a relatively new and rapidly developing field of study. Modern definitions of the discipline originated in the mid-1980s (Turner et al. 2001). Knowledge gaps, or topics of which we know there is incomplete understanding or lack of information, are common as they are with many aspects of natural resource management. We can point to knowledge gaps in places where new study and understanding are needed. We also know that knowledge gaps exist in locations where the public and land managers need to make recommendations and decisions based on the best information available in conjunction with strategic thinking, planning, and implementation. Primary areas discussed here with data gaps include the following:

- Mixed-severity fire regimes
- Moist and mixed-conifer forests
- Ecological forestry and multi-aged silviculture techniques
- Landscape patterns
- Vegetation species composition
- Forest structures, including complex early-seral and old forest
- Ecosystem resilience, reference conditions, and the natural range of variation
Call-out box 14. Uncertainty and risk in land management

Managing natural resources is especially challenging due to the inherent complexity of managing living things that are affected by randomly determined processes, such as fires and floods. It was important to consider uncertainty and risk in any projections and effects of land management decisions in this Bioregional Assessment, such as climate change and its impact on vegetation and disturbances. Not acknowledging uncertainties and tradeoffs can result in an underinformed decision. Incorporating uncertainty into decision-making and communications with the public to acknowledge risk can help update plans and meet new challenges nimbly.

Methods for dealing with uncertainty:

- Focus on ecosystem resilience
- Short-term and long-term goal setting
- Less focus on hypothetical desired future conditions and more on desired ranges that can be adapted to changing conditions.
- Plan for potentially dramatic changes in social, economic, and ecological changes
- Develop early warning systems
- Accelerate learning through communication and use of available technology
- Adaptive management

“The future has always been unknowable, but the levels of uncertainty about both future environmental and social conditions are extraordinary in this century” (Franklin et al. 2018).

Terrestrial Wildlife

Introduction

The Forest Service has a critical role in contributing to biological diversity by managing habitat for a broad array of species. Land management plans provide direction and a framework for habitat management. Maintaining or restoring ecological integrity provides for the habitat needs of species and the diversity of plant and animal communities (“coarse-filter” approach). Sometimes, when this is not enough, there is a need for additional, species-specific management that focuses directly on one organism’s population and habitat (“fine-filter” approach).

Although the protections created by the NWFP reserve networks have been effective in stemming the loss of old-growth forest, dependent species continue to decline, and new species have been listed since the amendment was enacted. There is a need to evaluate the effectiveness of the current reserve network in providing habitat for northern spotted owl, marbled murrelet and other old-forest-dependent species, while also recognizing the value of healthy and well-represented forest types of multiple successional and seral stages, including complex early seral, and maintaining habitat linkages and corridors.
Call-out box 15. Benefits of wildlife

The importance of a biologically diverse ecosystem is difficult to quantify but should be acknowledged. Diverse ecosystems are generally thought to be more resilient to change than non-diverse systems. Communities of plant and animal species are interdependent.

Besides the clear relationship to a healthy ecosystem, wildlife provides a variety of social and economic values to humans. For example, some species provide inspiration (aesthetic values), some provide information and resources (economic values), and for some people, wildlife have inherent value. In recent years, the value of wildlife to Northwest Forest Plan-area communities has been estimated in terms of the economic benefits that wildlife-based recreation provides. Spending by anglers, hunters, and wildlife watchers on National Forest System lands within the Bioregional Assessment area contributes considerably to both local and regional economies (USFWS 2011b, 2016).

For example, a 2011 survey conducted by the U.S. Fish and Wildlife Service found that 2.7 million people age 16 or older in Washington, 1.7 million people in Oregon, and 7.8 million people in California participated in fishing, hunting or wildlife viewing. The revenue generated from these activities was $24.79 billion in Washington, $21.7 billion in Oregon, and $7.58 billion in California.

What is Working Well

What is Working Well 1—Reserve Network

Much of what is working well for terrestrial wildlife habitat and biological diversity is closely integrated with ecological integrity. The reserve network, including late-successional reserves, riparian reserves, and congressionally reserved lands, is part of a landscape-scale approach that has worked well in supporting the conservation of habitat for wildlife species. Creation of designated areas and land use allocations that focus on species recovery has included clear and effective management direction in the context of habitat protection.

What is Working Well 2—Conservation of Dense Multi-layered Old-growth Forests

The NWFP conservation strategies and other strategies, including the Eastside Screens and the Sierra Nevada Framework, have effectively stopped the loss of old trees and old-growth forest on federal lands from timber harvest, mainly in dense, multi-layered forest (Spies et al. 2018a). Old-growth forest is generally considered stable on federal lands and has increased slightly since 1993, providing the abundance, diversity, connectivity, and availability needed to support ecosystem functions and specific old-growth-dependent species in the BioA area (Davis et al. 2015; Davis, in progress).

14 The “working well” of “Conservation of Dense Multi-layered Old-growth Forests” may seem like a dichotomy as the “Forest Ecology” section (“Key Change Issue 2”) outlines how many types of old forest, other structural stages, and overall ecosystem resilience is declining across much of the BioA area. In many respects, the NWFP has been successful in conserving old-forest ecosystems. However, conservation of a specific types of older forests does not necessarily ensure ecosystem resilience and integrity across the diversity of landscapes in the planning area. Threats to old-growth forests vary, depending on forest type and geographic location, but are heightened due to changing climate.
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The northern spotted owl, listed as threatened under the Endangered Species Act (ESA) in 1990, is one of many species that relies on old-growth forest habitat. The reserve network established by the NWFP has been effective in stemming the loss of old-growth habitat from timber harvest on federal lands (Lesmeister et al. 2018); however, the owl population continues to decline. The reserve network has also been effective in maintaining and enhancing marbled murrelet habitat on federal lands; however, the birds continue to experience population declines in the northern portion of the BioA area (Raphael et al. 2018). Since the NWFP was adopted, additional conservation focus has been placed on other species, including marten, fisher, wolverine, and other mammalian carnivores who also depend, in part, on late-successional forest habitats.

What is Working Well 3—Survey and Manage

The survey and manage guidance in the NWFP requires that surveys for certain species be conducted before initiating management actions, and actions are limited based on the results of the surveys. Survey and manage has added much to our knowledge about rare and uncommon late-successional and old-forest-dependent species in the NWFP area (Marcot et al. 2018) and has resulted in species not being listed under the ESA. Also, some species have been removed from the survey and manage list after the increased survey efforts resulted in the discovery that they are more common than previously expected. Survey and manage mitigation measures help us focus on certain species and contribute to the modernization of forestry practices, such as leaving more dead trees, downed wood, and refugia habitat. However, the survey and manage program has not been without its challenges, and improvements and updates are needed.

Key Change Issues

Key Change Issue 1—Northern Spotted Owl Habitat and Planning Direction

When critical habitat was designated for the northern spotted owl in 1992 under the ESA, the habitat was designed to be consistent with the late-successional reserve network. However, under the 2012 critical habitat designation, some matrix lands were identified as critical habitat. There is a need for land management plan direction that better aligns with the U.S. Fish and Wildlife Service’s northern spotted owl recovery plan (USFWS 2011a) (figure 44 and map 13) and critical habitat final rule (Spies et al. 2018b). Better alignment is needed between designated critical habitat for spotted owls and the late-successional reserve network; this could help simplify management direction and better protect high-quality habitat for owls and other species that depend on dense, multi-layered, old-growth habitat, such as marbled murrelet. Additionally, better alignment with northern spotted owl recovery plan guidance, specifically recovery actions 10 and 32, and land management plan components could help streamline project planning and consultation (recovery action 10 involves conserving spotted owl sites and high-value spotted owl habitat to provide additional demographic support and recovery action 32 calls upon land managers to maintain and restore all high-quality northern spotted owl habitat that has large diameter trees, high amounts of canopy cover, and decadence components such as snags, downed wood, and broken-top live trees). In addition, management direction that is consistent with both the recovery plan and critical habitat final rule and that calls for active management to restore and improve ecosystem resilience could help conserve and develop northern spotted owl habitat in the long term.
The protections created by the NWFP reserve networks have been effective in stemming the loss of northern spotted owl habitat from timber harvest on federal lands (Lesmeister et al. 2018). National Forest System lands serve as the primary federal habitat for the northern spotted owl due to the large areas of connected suitable habitat and regulatory mechanisms that are protective of threatened or endangered species. There is a need to ensure the reserve network continues to function as intended in the face of a dynamic and ever-changing ecosystem, including addressing the alignment of designated northern spotted owl critical habitat and the current distribution of late-successional, old-growth habitats across the entire planning area (map 14).

Call-out box 16. Northern spotted owl

The historic range of the northern spotted owl stretches from southwest British Columbia through the Cascade Mountains and Coastal Ranges in Washington, Oregon, and northern California. A key component of northern spotted owl habitat is structurally complex older forest, especially for nesting and roosting. Early-seral habitat (for example, openings) is also important, especially in the drier portions of the northern spotted owl range. The Northwest Forest Plan and the Endangered Species Act have helped protect and enhance spotted owl habitat on federal lands, but habitat protection will not be enough to ensure long-term viability of the species (Lesmeister et al. 2018). Despite current protections, the species’ population continues to decline each year throughout its range. This decline is thought to be driven by continued reduction of nesting and roosting habitat, particularly on non-federal lands, increasing habitat loss from wildfires, and competition with expanding populations of barred owls. Climate change projections indicate that suitable habitat for spotted owls will shift northward and to higher elevations (Carroll et al. 2010). The complex topography of the Cascades and Klamath Mountains could provide extensive refugia in shaded canyons and other microclimates. Future management considerations should include northern spotted owl competition with other species as well as the complex and dynamic habitat within the Bioregional Assessment area.
Figure 44. Designated critical habitat in each Northwest Forest (NWFP) land use allocation

Desgnated critical habitat for the northern spotted owl is most in line with late-successional reserves, congressional reserves, and administratively withdrawn allocations.
Map 13. Northern spotted owl designated critical habitat and range within the BioA area
Map 14. Northern spotted owl critical habitat overlaying Northwest Forest Plan land use allocation
Planning Considerations

Refer to BioA Chapter 2 Management Recommendations

Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.

Recommendation 2: Address the dynamic nature of ecosystems to be better positioned to respond to future environmental uncertainties.

Recommendation 6: Recognize that fire is a natural process and plays an important role in reducing the risk of uncharacteristic fire and in promoting ecosystem health.

Recommendation 7: Expand the use of timber harvest as a restoration tool to provide economic and social benefits to communities.

Recommendation 8—Shift from single-species management to maintaining and restoring habitat for multiple species to manage for ecosystem resilience under future uncertainty.

Recommendation 9: Promote active management in plant and animal habitats to restore and promote ecological resilience.

Geographic Considerations

The proportion of northern spotted owl designated critical habitat is least aligned with late-successional reserves in the Mt. Hood, Umpqua, Six Rivers, and Willamette National Forests (see figure 44). Plan direction that allows active management to restore and improve ecosystem resilience could help enhance and protect northern spotted owl critical habitat where it overlaps frequent-fire dependent and fire diverse (mixed severity) forests.

Key Change Issue 2—Northern Spotted Owl Habitat and Need for Active Management

Modification of land management plan desired conditions associated with old-forest management in drier, frequent-fire dependent ecosystems is needed. Loss of old forest from high-severity wildfire has been concentrated in frequent-fire dependent ecosystems (Spies et al. 2018a). It will be important to update land management plans to reflect ecological resilience and expected ecosystems in these areas.

The current capability of National Forest Service lands to provide late-successional forest habitat is strong; however, that capability is threatened by uncharacteristic wildfire, climate change, and invasive species. Uncharacteristic disturbances that broadly alter stands can remove features critical to northern spotted owl survival and reproduction, including sufficient canopy cover, large trees and snags. Climate change is projected to have a significant impact on the agency’s ability to maintain and sustain ecosystems and associated habitats in their current distribution.

As in figure 18, northern spotted owl critical habitat designation is most proportionally aligned with late-successional reserves on the Mt. Baker-Snoqualmie, Siuslaw, Olympic, Fremont-Winema, Mendocino, and Deschutes national forests with more than 60 percent of northern spotted owl critical habitat within such reserves on these national forests. Nonalignment of northern spotted owl critical habitat with late-successional reserves and other reserves indicates a potential need to adjust land allocations in coordination with designated critical habitat for northern spotted owl and other species.
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While the NWFP acknowledges the dynamic nature of fire-prone landscapes, there is a need for active management consistent with the Northern Spotted Owl Recovery Plan and Critical Habitat Final Rule that promotes the resiliency and ecological integrity of current and future cover types in the more fire-prone portions of the species range.

Planning Considerations

*Refer to BioA Chapter 2 Management Recommendations*

Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.

Recommendation 2: Address the dynamic nature of ecosystems to be better positioned to respond to future environmental uncertainties.

Recommendation 6: Recognize that fire is a natural process and plays an important role in reducing the risk of uncharacteristic fire and in promoting ecosystem health.

Recommendation 7: Expand the use of timber harvest as a restoration tool to provide economic and social benefits to communities.

Recommendation 8: Shift from single-species management to maintaining and restoring habitat for multiple species to manage for ecosystem resilience under future uncertainty.

Recommendation 9: Promote active management in plant and animal habitats to restore and promote ecological resilience.

Geographic Considerations

Within the frequent-fire dependent ecosystems common in northern California, southwest Oregon and the eastern Cascade Range, there is a challenge between providing forest structure suitable for nesting and roosting cover and the risk of habitat loss from high-severity fire (map 15). Habitat in fire diverse (mixed severity) ecosystems will increasingly face this challenge under a changing climate. Climate change will influence all areas, including the fire environment in some fire diverse (mixed severity) systems, particularly in south-central and southwestern Oregon, including the Umpqua.

Throughout the BioA area, there is geographic variation in northern spotted owl prey and foraging habitat patterns. In the southern and drier portion of the owl range, woodrats make up a high proportion of northern spotted owl diet. Woodrats are most abundant in shrubby habitats and owls forage from perches in older forest that borders early-seral habitat; in these areas, a mix of early-seral and late-seral forest is more typical. Woodrats require the retention of downed woody debris and other understory shrub species as habitat. In wetter portions of the range to the north, flying squirrels compose a higher portion of northern spotted owl diet and foraging habitat is more uniform, closed canopy older forest. There is also an elevational gradient to this pattern, with woodrats dominating in lower elevations and flying squirrels at higher elevations, especially in the southern extent of the northern spotted owl range. Consideration in this geographic distribution of northern spotted owl prey species is important when choosing active management activities across the range of the northern spotted owl in the BioA area.
Map 15. Broad fire ecology groups with northern spotted owl critical habitat overlaid
Key Change Issue 3—Northern Spotted Owl Habitat Restoration and Barred Owls

Despite the protections afforded by the NWFP, old-growth-dependent species such as the northern spotted owl continue to decline because of factors that were not anticipated. One of these factors is the barred owl, an invasive species in the BioA area. The barred owl has expanded its range in the past 25 years to cover the entire NWFP footprint and has become a significant threat to northern spotted owls. The expansion of the barred owl's range, in combination with disturbances outside of federal lands, has led scientists to conclude that the protections in the NWFP alone are not sufficient to ensure spotted owl recovery (Lesmeister et al. 2018).

In light of the additional impact on northern spotted owls from barred owls, there is an amplified need to continue to promote and conserve northern spotted owl habitat and increase treatment of currently unsuitable habitats to accelerate the attainment of suitable nesting/roosting habitat, while at the same time cooperatively addressing the barred owl threat. Habitat restoration, conservation, and enhancement will continue to provide for the needs of the northern spotted owl as options to mitigate barred owl impacts are evaluated and developed. There is also a need to develop projections of habitat change over time to better understand how northern spotted owl nesting habitat will change as currently unsuitable habitat matures within the reserve system. Managers need better tools to forecast how the total amount of nesting habitat will change, and whether total habitat might increase even if habitat losses continue on the non-federal landscape.

Planning Considerations

In order to retain and enhance northern spotted owl habitat while solutions to the barred owl threat are developed and evaluated, conservation and restoration of habitat should be considered in order to serve as a buffering agent against the compounding threat. Especially in fire-prone forests, there is a need to assess the increased risk that conserving more northern spotted owl habitat poses to fire extent and severity, ecological resilience and integrity, and to other forest management activities. Landscape-level approaches are needed to reconcile the potentially competing goals of forest resilience and northern spotted owl habitat.

Regardless of the status of the owl, it is important that the Forest Service continues efforts to retain and promote diverse and resilient late-successional habitats for the broad and diverse suite of species that rely upon them.

Refer to BioA Chapter 2 Management Recommendations

Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.

Recommendation 2: Address the dynamic nature of ecosystems to be better positioned to respond to future environmental uncertainties.

Recommendation 4: Reduce the introduction and spread of plant, animal, and other invasive species.
Call-out box 17. Barred owl invasion

A threat to the northern spotted owl that was not anticipated by the Northwest Forest Plan is the invasion and establishment of a non-native competitor. The barred owl, once confined to eastern North America, now co-occupies habitat and outnumbers spotted owls throughout much of their range and continues to increase in population density. Barred owls have higher annual survival, produce more offspring, and inhabit smaller home ranges than spotted owls. They compete for resources that would otherwise be available to spotted owls, including nest sites. They are slightly larger, and are strongly aggressive toward their native counterparts, usually quickly excluding spotted owls from territories and habitat. The competitive relationship between the two owl species has become a key limiting factor to spotted owl recovery (Lesmeister et al. 2018).

Recommendation 6: Recognize that fire is a natural process and plays an important role in reducing the risk of uncharacteristic fire and in promoting ecosystem health.

Recommendation 7: Expand the use of timber harvest as a restoration tool to provide economic and social benefits to communities.

Recommendation 8: Shift from single-species management to maintaining and restoring habitat for multiple species to manage for ecosystem resilience under future uncertainty.

Recommendation 9: Promote active management in plant and animal habitats to restore and promote ecological resilience.

Geographic Considerations

A threat to the northern spotted owl that was not fully anticipated by the NWFP is the invasion and establishment of a non-native competitor. The barred owl, once confined to eastern North America, now co-occupies and outnumbers northern spotted owls throughout much of their range and continues to increase in density (map 26). The barred owl’s newly extended geographic range now completely overlaps that of the northern spotted owl (Gutiérrez et al. 2007). Barred owls use the full range of forest types used by spotted owls, and a broader range of forest cover types outside of areas historically occupied by spotted owls. Barred owls have higher annual survival, produce more offspring, and inhabit smaller home ranges than spotted owls. They compete for resources that would otherwise be available to spotted owls, including nest sites. They are slightly larger, and are strongly aggressive toward their native counterparts, usually quickly excluding spotted owls from territories and habitat. The competitive relationship between the two owl species has become a key limiting factor to spotted owl recovery (Lesmeister et al. 2018). However, systematic studies have yet to quantify the full range of forest conditions that support barred owls in the Pacific Northwest. Incidental field data show a rapid increase in barred owls as they expanded their populations westward and southward into the range of the spotted owl (Dugger et al. 2016) (figure 45).
Photo 26. Juvenile barred owl

Figure 45. Mean annual local extinction rates (with 95 percent confidence intervals) for northern spotted owls on 11 study areas relative to presence of barred owl (Dugger et al. 2016)

It appears that northern spotted owls are more at risk of extinction with barred owls living in the area.
Key Change Issue 4—Marbled Murrelet Habitat

Call-out box 18. Marbled murrelet

The marbled murrelet is a small seabird that spends the majority of its time on the ocean but depends on old forest for nesting. Ranging from central California to the Aleutian Islands in Alaska, murrelets roost and forage primarily on small fish and krill in the nearshore marine environment. They may nest up to 55 miles inland, typically nesting in large trees with limbs containing moss or piles of needles large enough for laying a single egg and raising a nestling.

Murrelet populations are limited by available suitable marine and old-forest habitats. Nest predators such as crows, ravens, and jays are another management consideration since they are a primary cause of murrelet nest failure and will most likely limit murrelet populations in many areas. These predators are commonly associated with forest edges and open, thinned forests and are attracted to litter and food waste in human recreation areas. Murrelet populations are declining in the state of Washington but appear to be generally stable in Oregon and California.

The marbled murrelet occupies the coastal portions of the BioA area (map 16), and the current large reserve design on the coast has proven effective in maintaining and enhancing marbled murrelet habitat (Raphael et al. 2018). However, marbled murrelet populations have continued to decline, primarily due to rapid loss of habitat on non-federal lands from logging and threats to the marine environment, such as harmful algal blooms, oil spills, gillnet fishing and climate change. While there is much uncertainty about the role of murrelet movement along the coast and the influence this movement has on observed numbers and trends, this is largely outside of agency control. There is also uncertainty in how changes in marbled murrelet populations are related to the relative importance of change in the amount and distribution of nesting habitat versus changes in marine habitat, including predictions of how marine prey will respond to climate change scenarios.

Even with the success of maintaining and enhancing marbled murrelet habitat in the coastal regions of the BioA area, the shrinking habitat on non-federal lands heightens the importance of securing and protecting existing and additional old-growth habitat.

Planning Considerations

There is a need to maintain large, contiguous blocks of marbled murrelet habitat within its range. Enhancements and modifications to the current late-successional reserve network in the marbled murrelet range needs to focus on the maintenance of large, contiguous blocks of densely canopied late-successional habitat with ample nesting platforms. There is also a need to develop projections of habitat changes over time to better understand how marbled murrelet nesting habitat will change as currently unsuitable habitat matures within the reserve system. Managers need better tools to forecast how the total amount of nesting habitat will change in the future and whether total habitat might increase even if habitat losses continue on non-federal land.
Furthermore, minimizing corvid conflicts especially in developed recreation areas and other management activities areas in and adjacent to marbled murrelet nesting habitat will be an important planning consideration.

Refer to BioA Chapter 2 Management Recommendations
Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.

Recommendation 2: Address the dynamic nature of ecosystems to be better positioned to respond to future environmental uncertainties.

Recommendation 4: Reduce the introduction and spread of plant, animal, and other invasive species.

Geographic Considerations
Marbled murrelet populations have remained steady in Oregon and California but continue to experience steep declines in Washington (Falxa and Raphael 2016). Habitat losses continue on all non-federal lands and climate change is likely to affect future marbled murrelet populations, their nesting habitat, and their food resources.
Map 16. Extent of marbled murrelet critical habitat within the Bioregional Assessment area

NWFP = Northwest Forest Plan.
Key Change Issue 5—Mammalian Carnivores Habitat Connectivity

In the decades since the NWFP was designed, additional conservation focus has been placed on a broader suite of species for which late-successional forest habitats are important (Marcot et al. 2018). These include species such as Pacific marten, fisher, Canada lynx, Sierra Nevada red fox, grizzly bear, and wolverine (photo 27 and photo 28). Potential changes in management should include consideration of the needs of these additional species.

Potential modifications to the reserve network need to account for and provide for habitat linkages and connectivity sufficient for these species’ persistence at the range of the species distribution. Planning should emphasize retention of snags, large trees, mistletoe brooms and damaged trees to provide resting and denning structures for fisher and marten; downed wood is also an important habitat feature for resting sites, maternal denning, and prey habitat. These modifications would help align future plans to be consistent with the 2012 planning rule, which requires plans to include components to maintain or restore ecological integrity, including connectivity. Plan direction should emphasize coarse-filter and use fine-filter components where needed to address the connectivity needs of multiple species.

Planning Considerations

Refer to BioA Chapter 2 Management Recommendations

Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.

Recommendation 2: Address the dynamic nature of ecosystems to be better positioned to respond to future environmental uncertainties.

Recommendation 6: Recognize that fire is a natural process and plays an important role in reducing the risk of uncharacteristic fire and in promoting ecosystem health.

Recommendation 8: Shift from single-species management to maintaining and restoring habitat for multiple species to manage for ecosystem resilience under future uncertainty.

Recommendation 9: Promote active management in plant and animal habitats to restore and promote ecological resilience.

Geographic Considerations

Geographic trends for mammalian carnivores are not currently well known in the BioA area; more research on range and distribution would help address the needs of these species.
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Photo 27. Wolverine climbing to a feeding platform

Photo 28. Pacific marten perched in a tree
Key Change Issue 6—Habitat Diversity

Beyond old-forest-dependent species, national forests and grasslands provide important habitat for a broad range of species. The NWFP was designed to address the large-scale and rapid decline in late-successional habitats and placed an important emphasis on the conservation and promotion of dense multi-layered older forest conditions, in particular, the closed-canopy, structurally complex forests associated with northern spotted owl nesting and roosting habitat. While this emphasis continues to be critically important, the value of a distribution of forest types that represent resilient landscapes is becoming increasingly apparent. This usually includes multiple successional and seral stages distributed in ecologically significant patterns across the landscape. For example, there is now a greater recognition of the importance of complex early-seral habitats and the key habitat components (snags and downed wood) that they provide through later successional stages (Phalan et al. 2019; Swanson et al. 2011, 2014). These early-seral habitats are vitally important to invertebrate and vertebrate pollinator species and game species, such as deer and elk, which have suffered some population declines due to changes in federal land management that favor older forest conditions.

Planning Considerations

Refer to BioA Chapter 2 Management Recommendations

Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.

Recommendation 2: Address the dynamic nature of ecosystems to better respond to future environmental uncertainties.

Recommendation 6: Recognize that fire is a natural process and plays an important role in reducing the risk of uncharacteristic fire and in promoting ecosystem health.

Recommendation 8: Shift from single-species management to maintaining and restoring habitat for multiple species to manage for ecosystem resilience under future uncertainty.

Recommendation 9: Promote active management in plant and animal habitats to restore and promote ecological resilience.

Geographic Considerations

Fires have created complex early-seral habitat within the historic range of variability in the north central Cascades on the Okanogan-Wenatchee National Forest and southwest Oregon national forests and grasslands (Rogue River-Siskiyou, Umpqua, Fremont-Winema National Forests, and the Middle Fork Ranger District on the Willamette National Forest) and northwest California (particularly, the Klamath, Six Rivers, Shasta-Trinity and Modoc National Forests). Complex early-seral habitat is below historical ranges along the Coast Range (Siuslaw and Olympic National Forests), as well as the central Oregon and Washington Cascades (northern Willamette, Mt. Hood, Gifford Pinchot, and Mt. Baker-Snoqualmie National Forests (Phalan et al. 2019). In those areas in the Coast Range and central Cascades that have not experienced as much wildfire and are therefore lacking in complex early-seral habitat, there are opportunities to create and maintain complex early-seral habitat in previously managed plantations, in particular, plantations planted with off-site conifers.
Call-out box 19. Birds as indicators

Birds can be monitored relatively easily to measure ecological effects of land management actions. A group of bird species can be chosen to represent the key habitat attributes in a given habitat type, such as a mixed-conifer forest, and be monitored to ensure key habitat attributes are present. These bird species are often used as ecological indicators.

The Forest Service is working with several partners to integrate Partners in Flight decision support tools that inform management planning and measure outcomes. We are using birds as indicator species to inform planning and measure forest trajectories based on management outcomes that are expected to be achieved far into the future (for example, old-growth development).

Partners in Flight bird species monitoring, part of a partnership focused on helping species at risk and keeping common birds common, will help inform future plan revision or modernization efforts within the Bioregional Assessment area.

The hermit warbler, along with other indicator species, such as the Hammond’s flycatcher, Pacific wren, and brown creeper, are highly associated with mature forests. The habitat attributes these species are associated with represent a range of conditions that are important in these forests (for example, closed canopy, open mid-story, deciduous understory, and forest floor complexity) (Altman and Alexander 2012).
Aquatics, Fish, and Water

Introduction
In the BioA area, 6.7 million acres were designated as riparian management areas through four plan amendments, including (1) the NWFP Aquatic Conservation Strategy (ACS) (USDA FS and USDI BLM 1994); (2) Sierra Nevada Framework Aquatic Management Strategy (USDA FS 2004); (3) PacFish (USDA FS and USDI BLM 1995b); and (4) InFish (USDA FS and USDI BLM 1995a). Seventy-two percent of the BioA area and 85 percent of the aquatic and riparian habitats are managed under the NWFP ACS. Ten percent of the BioA area and 6 percent of the aquatic and riparian ecosystems fall under the Sierra Nevada Framework Aquatic Strategy. The PacFish and InFish aquatic strategies, which are nearly identical, combine to cover 17 percent of the land area and 8 percent of the aquatic ecosystems (map 17). The four amendments will be collectively referred to as the BioA aquatic strategies.

The BioA aquatic strategies include goals, objectives, standards and, guidelines to prevent damage to riparian areas, along with four management components: riparian management areas, key or refuge watersheds, watershed analysis, and aquatic restoration, all of which work in concert to maintain or restore aquatic ecosystems. The NWFP ACS is a good example of a multi-species coarse-filter approach that has improved conditions within aquatic and riparian ecosystems on which anadromous fish and other organisms depend.

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16 Links to electronic versions of the four BioA aquatic strategies.
NWFP: https://www.fs.fed.us/r6/reo/acs/
Sierra Nevada Plan: https://www.fs.usda.gov/detail/r5/landmanagement/planning/?cid=STELPRDB5349922
PacFish: https://www.fs.usda.gov/Internet/FSE_DOCUMENTS/fseprd591470.pdf
InFish: https://www.fs.usda.gov/Internet/FSE_DOCUMENTS/fseprd591470.pdf

17 About one percent of the riparian management acres are not managed by any one of the BioA aquatic strategies.
Map 17. Bioregional Assessment (BioA) aquatic strategies and coverage area
What is Working Well

**What is Working Well 1—Bioregional Assessment Aquatic Strategies**

The BioA aquatic strategies are working; after 20 years of NWFP implementation, the Aquatic and Riparian Effectiveness Monitoring Program detected improving trends for aquatic physical habitat, aquatic macroinvertebrates and 7-day average maximum water temperatures, along with improvements in upslope and riparian conditions (photo 29) (Miller et al. 2017). While complex, the improvements in physical habitat noted by the Aquatic Riparian Effectiveness Monitoring Program across the NWFP national forests show an improving shift in the physical habitat scores. This improvement indicates that changes in land management have been effective at improving aquatic habitat (figure 46).

Outside of the NWFP area, Roper et al. (2019) documented that nine of 10 stream attributes are trending upward or stable on federal lands managed by PacFish and InFish aquatic strategies, supporting the conclusion that 20 years of management under these strategies likely played a role in improving stream conditions. Most apparent was a reduction in fine substrates in the downstream end of pools and an increase in in-stream large wood (Roper et al. 2019). Further, the most current monitoring report for the Sierra Nevada Framework rated the condition of 78 percent of evaluated streams as “good” to “excellent” (Furnish 2013).

The Aquatic Conservation Strategy authors (USDA FS and USDI BLM 1994: B-9) recognized that it may take at least 20 years or even more than a century to restore ecological processes across the NWFP area, a timeframe that can likely be applied to PacFish, InFish, and Sierra Nevada Framework areas as well. Recent monitoring results are promising and help to confirm, along with science reviews (Naiman et al. 2000, Reeves et al. 2018, Spence et al. 1996), that the four primary components that form the structure of the BioA aquatic strategies are functionally sound and provide a solid foundation to move forward with continued improvements. During the 2015 listening sessions,18 public participants offered general support for the continuation of existing aquatic programs that protect and improve water quality, habitat for salmon and other aquatic species, and overall watershed health.

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Photo 29. Aquatic and Riparian Effectiveness Monitoring Program field staff collecting data

Figure 46. Aquatic habitat improvements in the Northwest Forest Plan (NWFP) area

This graph highlights the improvements in physical habitat noted by the Aquatic Riparian Effectiveness Monitoring Program across the NWFP national forests and grasslands. It shows a rightward (improving) shift in the physical habitat scores from sampling period 1 (2002–2009) to sampling period 2 (2010–2013). (Miller et al. 2017).
**What is Working Well 2—Riparian Management Areas**

Riparian management areas,\(^{19}\) a cornerstone to the BioA aquatic strategies,\(^{20}\) are a primary reason for watershed improvements across the BioA aquatic ecosystems. Composed of approximately 6.7 million acres of land that border 154,000 miles of streams and surround 211,388 acres of lakes, ponds, and wetlands, riparian management areas protect and naturally restore watershed and ecological processes (for example, stream shade, large wood input, floodplain functions). The vast majority of these acres (about 85 percent) occurs on national forests and grasslands managed by the NWFP (table 4). The size of riparian management areas is dependent on several criteria associated with a given waterbody. For example, fish-bearing streams receive greater protection than non-fish-bearing streams.

<table>
<thead>
<tr>
<th>Strategy</th>
<th>Acres of Riparian Management Area</th>
<th>Percentage of Acres</th>
</tr>
</thead>
<tbody>
<tr>
<td>Aquatic Conservation Strategy (Northwest Forest Plan)</td>
<td>5,717,362</td>
<td>85</td>
</tr>
<tr>
<td>Sierra Nevada Framework</td>
<td>540,716</td>
<td>6</td>
</tr>
<tr>
<td>PacFish and InFish</td>
<td>424,623</td>
<td>8</td>
</tr>
<tr>
<td>Outside the above aquatic strategies</td>
<td>45,788</td>
<td>1</td>
</tr>
<tr>
<td><strong>Total</strong></td>
<td><strong>6,728,489</strong></td>
<td><strong>100</strong></td>
</tr>
</tbody>
</table>

The primary management tool used to attain watershed improvements is passive restoration, meaning that riparian management areas are left to recover naturally without influence from other forest management actions. Trees that might have been targeted for timber harvest before 1994 are now left to grow and provide stream shade and aquatic and terrestrial habitat and help create a network of migration corridors for animal species throughout and between watersheds. Where passive restoration may not be sufficient to restore riparian conditions, silvicultural treatments can be applied within riparian areas; for example, the ACS allows vegetation management to “acquire desired vegetation characteristics needed to attain ACS objectives.”\(^{21}\)

To facilitate this transformation from timber production to a more balanced management approach, standards and guidelines were developed to ensure that any action in riparian management areas that overlay or intermix with other land use allocations, where active management is permitted, must maintain, protect, or restore watershed and ecological

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\(^{19}\) The term “riparian management areas” is a collective term used to describe NWFP riparian reserves, Sierra Nevada Framework riparian conservation areas, and PacFish and InFish riparian habitat conservation areas.

\(^{20}\) Citations with page numbers to each of the four BioA aquatic strategies where riparian management area descriptions are located: NWFP aquatic conservation strategy (USDA FS and USDI BLM 1994: C30-31), Sierra Nevada Framework Aquatic Management Strategy (USDA FS 2004: 42), PacFish (USDA FS and USDI BLM 1995b: C6-9), and InFish (USDA FS 1995a: A5-6).

\(^{21}\) NWFP ACS (USDA FS and USDI BLM 1994: C-32, TM1-c.).
processes. Photos 30 through 36 show the diversity of stream and riparian types that occur throughout the BioA area, from northern Washington’s relatively wet Mt. Baker-Snoqualmie and Olympic National Forests south to northern California on the more arid Klamath and Shasta-Trinity National Forests.

Photo 30. Stream on Mt. Baker-Snoqualmie National Forest

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Photo 31. Stream on Olympic National Forest

Photo 32. Stream on Willamette National Forest
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Photo 33. Stream on Deschutes National Forest

Photo 34. Stream on Fremont-Winema National Forest
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**What is Working Well 3—Watershed Restoration**

To complement passive restoration and standards and guidelines provided through riparian management areas, the existing aquatic strategies in the BioA area include a watershed and aquatic restoration component, which targets impaired watershed processes and habitats negatively affected by past management. Under the NWFP, priority projects include treatment or removal of roads prone to landslide and erosion. Photo 37 shows an aquatic restoration project on the Gifford Pinchot National Forest where a road was decommissioned and a culvert removed to reclaim a natural stream channel, restore aquatic organism passage, and reduce road-related sediment in the broader stream network. Photo 38 shows a project in which the Olympic National Forest and partners placed large wood along a river’s edge and floodplain to create pools and hiding cover for ESA-listed fish. Large wood placement improves habitat for a variety of other riparian-dependent species, both aquatic and terrestrial, resulting in an abundance of ecosystem benefits.

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23 Citations with page numbers where aquatic restoration sections can be found in each of the four BioA aquatic strategies: NWFP Aquatic Conservation Strategy (USDA FS and USDI BLM 1994, p. B 30-32); Sierra Nevada Framework Aquatic Management Strategy (USDA FS 2004, p. 34 Rod); PacFish (USDA FS and USDI BLM 1995a, p. C 21-22); and InFish (USDA FS 1995b, p. E 15).

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Photo 37. Road decommissioning on Gifford Pinchot National Forest

Photo 38. Large wood placement on Olympic National Forest
Prior to 2017, Staley Creek on the Willamette National Forest was intentionally confined by placement of berms to control channel movement and lateral flooding. The berms crowded the stream into a straightened channel, increased stream velocity, and simplified aquatic habitats and the stream’s ability to accommodate the broad array of aquatic species native to the area. After restoration in 2017, Staley Creek now flows across a wide floodplain during high-flow events, creating an assortment of habitat types for aquatic and terrestrial species, including ESA-listed bull trout, and recharges groundwater across an expansive area (photo 39 and photo 40).

Photo 39. Staley Creek, before restoration, on Willamette National Forest

Photo 40. Staley Creek, after restoration, on Willamette National Forest

Significant amounts of active riparian management have occurred in the BioA area since the initiation of the aquatic strategies. For example, from 2014 through 2018, restoration projects improved 1,968 miles of stream habitat, restored fish access to 357 miles of stream, and removed 490 miles of roads that affected water quality, local hydrology, and other watershed processes. Projects were, and continue to be, integrated with other actions throughout watersheds, addressing impacts to watershed processes and aquatic habitat.

**What is Working Well 4—Watershed Analysis**

Watershed and aquatic restoration projects are identified through another aquatic strategy component, watershed analysis, which is an interdisciplinary approach to assess statuses and trends of physical and ecological processes in a watershed. A watershed analysis identifies what is and what is not working for a given area and sets the foundation for remedial actions to treat chronic and systemic problems that alter important watershed

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24 Citations with page numbers where watershed analysis sections can be found in the four BioA aquatic strategies: NWFP Aquatic Conservation Strategy (USDA FS and USDI BLM 1994, B 20-30); Sierra Nevada Framework Aquatic Management Strategy (USDA FS 2004, p. 33); PacFish (USDA FS and USDI BLM 1995a, p. C 19-21); and InFish (USDA FS 1995b, p. E 14-15).
functions. Under the NWFP, the analysis is mandatory for project planning and implementation in riparian reserves and key watersheds.

More than 200 watershed analyses have been conducted on national forests and grasslands in the NWFP area, most of which were completed over a 10-year period from the mid-1990s through the early 2000s. See photo 41 for a watershed view located on the Rogue River-Siskiyou National Forest.

![Photo 41. Watershed on Rogue River-Siskiyou National Forest](image)

*Watershed analysis, an essential part of the Bioregional Assessment aquatic strategies, considers the entire landscape, from ridge top to valley bottom, to identify where watershed processes are working well and where past management actions negatively affect terrestrial and aquatic habitats, to help determine sites for future restoration.*

**What is Working Well 5—Key Watersheds**

The aquatic strategies in the BioA area include key watersheds,\(^\text{25,26}\) which generally are priority areas for aquatic restoration because they are designated as refuge areas for aquatic species and are important sources of water. From 2012 to 2017, nearly 60 percent of watershed and aquatic restoration projects on NWFP national forests in the Forest Service’s Pacific Northwest Region were conducted in key watersheds (map 18). In the NWFP area, about 8.5 million acres were designated as key watersheds and currently provide habitat for 23 of the 27 aquatic species listed under the ESA (Reeves et al. 2018). These include 20 species of salmon and steelhead, three distinct populations of bull trout, two sucker species, Pacific eulachon, and the Oregon spotted frog. See photo 42, photo 43, and photo 44 for examples of common fish species in the BioA area.

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\(^{25}\) The term “Key Watersheds” are used in the NWFP and PacFish, while Critical Aquatic Refuges are used in the Sierra Nevada Framework, and Priority Watersheds is the term used under InFish.

\(^{26}\) Citations with page numbers where key watershed sections can be found in each BioA aquatic strategy, NWFP Aquatic Conservation Strategy (USDA FS and USDI BLM 1994, pages B 18-20); Sierra Nevada Framework Aquatic Management Strategy (USDA FS 2004, pages 43-44); PacFish (USDA FS and USDI BLM 1995a, page C 19); and InFish (USDA FS 1995b, pages E 13-14).
Photo 42. Adult bull trout in a stream

In the Bioregional Assessment area, Washington and Oregon national forests offer the primary strongholds for bull trout, a species listed as threatened under the Endangered Species Act. Resident bull trout require clear, cold-water streams, and National Forest System lands often provide the sole source of this essential habitat. (U.S. Fish and Wildlife Service).

Photo 43. Chinook salmon in a stream

Chinook salmon, the iconic and most recognized Pacific Coast salmon, can be found on 16 of the 19 units in the Bioregional Assessment area. The spring, summer, and fall Chinook, which often exceed 20 pounds, migrate to national forest streams to lay their eggs, and the juveniles seek cool water tributaries to rear and grow before migrating to the ocean. Chinook salmon streams are usually a primary target for national forest aquatic restoration projects. (U.S. Fish and Wildlife Service).

Photo 44. Steelhead trout leaping a barrier

Steelhead, an anadromous trout, moves up the Sacramento River in search of spawning grounds. Steelhead can migrate back and forth from the ocean more than once to spawn, and national forest lands in the Bioregional Assessment area provide much of the headwater spawning and rearing habitat preferred by these fish. (U.S. Fish and Wildlife Service).
Key watersheds and other National Forest System lands provide most of the water supply that sustains the majority of Western U.S. communities with drinking water, and supply irrigation water to nationally important agricultural areas like California’s Central Valley, Oregon’s Willamette Valley, and the Columbia River valley (map 18 and map 19).

Map 18. Northwest Forest Plan and PacFish key watersheds and InFish priority watersheds in the Bioregional Assessment area
Map 19. Percentage of water supply from National Forest System lands

Percentage of total water supply that originates on National Forest System lands (Luce et al. 2017). National forests and grasslands are an important source of abundant high-quality water for many uses, including aquatic habitat, drinking water, and irrigation. They typically provide much of the water for a given basin. BioA = Bioregional Assessment.
Key Change Issues

Key Change Issue 1—Management Efficiency and Need for Single, Unified Aquatic Conservation Strategy

Aquatic and riparian ecosystems in the BioA area are managed by one of four aquatic strategies: NWFP ACS, Sierra Nevada Framework Aquatic Management Strategy, PacFish, and InFish. All four strategies are similar in their architecture and approach, but the level and types of analysis and compliance requirements vary amongst the strategies, increasing Forest Service planning costs when a project area is covered by more than one strategy. Nine national forests and grasslands in the BioA area, for example, operate under more than one BioA aquatic strategy. On the Okanogan-Wenatchee National Forest, the Chinook salmon, which spawns in the Chewuch River within the forest’s borders, can readily swim back and forth across a NWFP-PacFish boundary. When the forest developed the Chewuch Transportation Plan Environmental Assessment, it was required to conduct separate analysis and compliance reviews for the same project, the same fish on the same stream to accommodate two different strategy directives, the NWFP and PacFish.\(^{27}\) The need to conduct multiple compliance reviews increases direct costs to the affected forest as well as social and ecological costs, which increase when a project area expands to a state, Forest Service region, or larger level, and include an increasing number of aquatic strategies that must be addressed.

Planning Considerations

To eliminate financial, ecological, and social costs required to address multiple aquatic strategies, develop a unified aquatic conservation strategy for the BioA area.

Refer to BioA Chapter 2 Management Recommendation

Recommendation 3: Update and consolidate the existing aquatic direction processes and analysis requirements.

Geographic Considerations

Eight national forests and grasslands and the scenic area, most of which occur on the eastern portion of the BioA, are covered by more than one aquatic strategy and would benefit the most from a unified aquatic strategy: Okanogan-Wenatchee, Gifford Pinchot, Mt. Hood, Deschutes, Ochoco, Fremont-Winema, Modoc, and Lassen National Forests and Columbia River Gorge National Scenic Area.

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\(^{27}\) The Endangered Species Act consultation requirements for fish listed as threatened or endangered are complicated when addressing more than one aquatic strategy.
Key Change Issue 2—Fish Species Adaptation and Need for Detailed Descriptions of Desired Future Conditions and Disturbance Regimes

The NWFP ACS objectives, as well as Sierra Nevada Framework, PacFish, and InFish aquatic management goals and objectives, often serve as ecosystem-based desired future conditions that highlight habitats under which aquatic species are uniquely adapted.²⁸ For example, ACS objective 8 states “Maintain and restore species composition and structural diversity of plant communities in riparian areas....” The general nature of this and other objectives often leads to disputes over the true character of riparian reference conditions and a potential need for active management. There is a need for active management in riparian areas throughout the NWFP area; this includes thinning dense riparian Douglas-fir stands to promote growth of hardwoods and expediting growth of large conifers that could serve as key structural components in stream channels (Reeves et al. 2018). There is a growing concern in frequent-fire dependent forests, where many watersheds that are inhabited by fish contain large areas of densely stocked, mid-seral stands susceptible to stand-replacement fire (map 20). These large stand-replacement fires could negatively affect fish habitat by reducing stream shade, water quality, and aquatic habitat complexity. However, mechanical and fire treatments to reduce stand densities to more sustainable conditions are sometimes avoided in densely stocked riparian areas because of differing views on reference conditions, leaving these areas vulnerable to stand-replacement fires.

Although native fish populations are adapted to natural fire regimes (Flitcroft et al. 2016, Reeves et al. 1995) and demonstrate resiliency even when faced with stand-replacement fires (Dunham et al. 2003), the growing number of uncharacteristically large and intense wildfires could increase adverse impacts and even local extinctions of ESA-listed fish that are already constrained by habitat degradation (Rieman et al. 2003) and fragmentation (Dunham et al. 2003).

²⁸ This footnote provides page numbers where aquatic and riparian objectives or goals can be found in the BioA aquatic strategies. NWFP Aquatic Conservation Strategy (USDA FS and USDI BLM 1994, p. B-11); Sierra Nevada Framework Aquatic Management Strategy (USDA FS 2004, p. 42-43); PacFish (USDA FS and USDI BLM 1995b, p. C 3-6); and InFish (USDA FS 1995a, p. E 2-4).
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Map 20. Fish-bearing watersheds and departed forests

This map displays the convergence of highly departed forests (Ringo et al. 2019), unusually dense timber stands relative to historic conditions, and fish-bearing streams in watersheds on the Klamath and Shasta-Trinity National Forests.

Planning Considerations

To promote a common understanding of riparian reference conditions, revised plan components should detail prevalent disturbance regimes for geographic areas and include descriptions of resulting riparian vegetation (structure, species composition, and landscape patterns), riparian and aquatic habitats, and water quality over space and time.

Refer to BioA Chapter 2 Management Recommendations

Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.

Recommendation 3: Update and consolidate the existing aquatic direction processes and analysis requirements.

Recommendation 6: Recognize that fire is a natural process and plays an important role in reducing the risk of uncharacteristic fire and in promoting ecosystem health.

Additional planning considerations that were not addressed by the BioA include developing desired future conditions in collaboration with the U.S. Fish and Wildlife Service and National Marine Fisheries Service.
Geographic Considerations

A description of desired future conditions and foundational disturbance regimes would help address forest and non-forest conditions that are departed from more natural conditions across the BioA area. In the BioA area, two groups of fire-dependent forests stand out as having high rates of vegetative departure from the natural range of variation in watersheds inhabited by fish.\(^2^9\)

**Group 1—High Urgency for Action:** In 95 percent of fish-bearing watersheds in northwest California’s fire-dependent national forests (Klamath, Mendocino, Shasta-Trinity, and Six Rivers), vegetative departure is moderate to high. To compound the urgency for action in northwest California, the Mendocino, Shasta-Trinity, and Six Rivers National Forests have relatively high numbers of ESA-listed fish species\(^3^0\) and occur in a geographic region that is projected to experience the most extreme climate change impacts, rendering them more susceptible to large, high-severity wildfire, in the NWFP area (refer to “Key Finding 2” in the “Climate Change” section below).

**Group 2—Moderate to High Urgency for Action:** Approximately 70 to 75 percent of the fish-bearing watersheds in southwestern, south-central, and central Oregon national forests (Rogue River-Siskiyou, Fremont-Winema, and Deschutes) have high departure rates and occur in the geographic region that is projected to experience the most extreme climate change impacts in the BioA area.

**Key Change Issue 3—Key Watersheds and Need to Realign with Critical Habitat and Current Science\(^3^1\)**

The NWFP key watersheds and their Sierra Nevada Forest Plan, PacFish, and InFish counterparts were identified as the best places to protect and recover at-risk fish stocks and other riparian-dependent species and to provide high-quality water on and from federal lands. Since that time, new methods to identify valuable habitats for ESA-listed salmon, steelhead, bull trout, and suckers strongly suggest that some key watersheds may not be aligned with important recovery areas for these fish; nearly 40 percent of critical habitat for ESA-listed fish on NWFP national forests and grasslands lie outside key watersheds. This raises uncertainty about the effectiveness of key watersheds in serving as refuge areas for at-risk fish stocks (Reeves et al. 2018). Further uncertainty exists because streams with high intrinsic potential to support fish, based on a methodology developed after the establishment of key watersheds (Burnett et al. 2005), have not been fully mapped across the BioA area, and climate change assessments that provide insights as to where climate impacts will be most adverse to native fish have yet to be incorporated for the entire area.

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\(^2^9\) At the time of this report, fish distribution data were not readily available for the Modoc and Lassen National Forests outside of the NWFP area. Therefore, these two forests were not included in this analysis.

\(^3^0\) Number of fish listed under the U.S. Endangered Species Act on high urgency national forests: Mendocino, 5; Shasta-Trinity, 4; Six Rivers, 3. The Mt. Hood and Gifford Pinchot National Forests have higher numbers at 7 and 6, respectively; these forests, however, occur in a geographic region where the urgency for action for this key change issue is low. Fish numbers are taken from Reeves et al. (2018).

\(^3^1\) The term “key watersheds” is used in the NWFP and PacFish, while “critical aquatic refuges” are used in the Sierra Nevada Framework, and “priority watersheds” is the term used under InFish. The term “key watersheds” is a collective term used to describe these watersheds.
One example of a critical habitat and key watershed disconnect can be found on the Siuslaw National Forest’s Siuslaw River basin, where nearly 60 percent of the designated critical habitat for the ESA-listed Oregon Coast coho salmon occur outside of key watersheds that were identified in 1994 (map 21). It is in basins such as this where better alignment between key watersheds and critical habitat may be warranted.

Map 21. Critical habitat alignment with key watersheds

This map displays the amount of critical habitat in and outside of key watersheds in the Siuslaw River basin on the Siuslaw National Forest. Nearly 60 percent of the critical habitat lies outside of key watersheds that were designated in 1994, emphasizing the need to reevaluate whether key watersheds are located in the best areas to address recovery of Endangered Species Act- (ESA)-listed fish. NWFP = Northwest Forest Plan.

Planning Considerations
Where needed, locations of key watersheds could be adjusted to better align with places best suited for recovery of ESA-listed fish and other areas of critical importance to aquatic biodiversity (for example, species of conservation concern). Any adjustment can be based on linkages to critical habitat, high intrinsic potential streams, bull trout core areas, climate change vulnerability assessments, presence of vulnerable aquatic- and riparian-dependent species not listed by the ESA, and priority refuge areas documented in federal ESA-listed fish recovery plans. Assessment actions could be coordinated with the National Marine Fisheries Service and U.S. Fish and Wildlife Service ESA-listed fish recovery programs.
Geographic Considerations

The national forest units with the greatest percentage of critical habitat that occurs outside key watersheds are the Modoc, (100 percent), Siuslaw (61 percent), Willamette (56 percent), Mt. Hood (45 percent), Olympic (38 percent), and Mendocino (32 percent) National Forests and the Columbia River Gorge National Scenic Area (greater than 75 percent). Critical habitat maps were not available for the southern Oregon/northern California coho salmon, which occurs on the Rogue River-Siskiyou, Klamath, Six Rivers, and Shasta-Trinity National Forests. Thus, these national forests could not be included in the key watershed/critical habitat assessment.

Key Change Issue 4—Management Inefficiencies and Need for Revised Watershed Analysis Process

Watershed analysis played a significant role in institutionalizing ecosystem-based management on forest system lands in the BioA area. These analyses provide an interdisciplinary look at a watershed, culminating with a list of resource management actions to help reverse degraded watershed processes, habitats, and undesirable trends. Since the completion of watershed analyses in the mid-1990s and early 2000s, the condition of many watersheds has been altered primarily through extensive wildfires and insect and disease outbreaks. When a watershed analysis is required to plan and implement subsequent projects in these newly disturbed areas, the existing analysis requires revision. The current watershed analysis development and revision process is often unconstrained in scope and level of detail, creating excessive timelines for management activities that demand its use, such as forest plan consistency assessments required to support active management in riparian areas and key watersheds.

Planning Considerations

A plan component could be developed to ensure that new or updated watershed analyses or similar documents address only the most critical current issues and questions and that the type and level of analysis is aligned with current management needs and opportunities, financial resources, and staff capacity. Creation or revision of a watershed analysis may be initiated by project or management requirements.

Geographic Considerations

National forests and grasslands that have and continue to experience large stand replacement fires (northern California, southern Oregon, and eastern Cascade Mountains of Oregon and Washington), which results in altered watershed conditions, have the greatest need to operate under an updated watershed revision process.

Key Change Issue 5—Aquatic Species Connectivity and Need for Stream Passage for all Aquatic Species

The NWFP, PacFish, and InFish aquatic strategies contain a standard and guideline that directs national forests and grasslands to provide fish passage where new culverts are to be installed at road stream crossings (USDA FS and USDI BLM 1994: C-33, 1995b: C-11; USDA FS 1995a: A-8), but does not address the need to provide passage at road stream crossings for all aquatic species to help ensure their long-term viability. For instance, culverts can restrict salamander movement in and along streams.
Call-out box 20. A riparian vegetation management case study

Throughout the Bioregional Assessment area, mechanical and fire treatments in riparian areas are often challenging as we seek to better understand local riparian functions and project tradeoffs. In some cases, restoring one part of a riparian area may appear to degrade another. Conifer thinning in riparian areas to increase space for deciduous trees and shrubs, for example, reduces one habitat type and increases another. Therefore, it is important to weigh tradeoffs in the context of natural disturbance regimes and associated plant species, structure and function, and resulting habitats.

In the Dry Hills Forest Restoration Project on the Lassen National Forest, it was critical to understand and clearly describe the natural range of variation of local riparian areas. The Lassen’s land and resource management plan and its associated amendment (Sierra Nevada Framework) authorizes active management in riparian areas—similar to the Northwest Forest Plan—as long as it meets or improves riparian conditions. Because the plan’s desired future conditions are general in nature, staff time and expertise were spent providing detailed desired future conditions that would identify the need for riparian restoration. In this case, the project team clearly described the natural range of variation for the project’s riparian areas, including the fire regime and associated forest plant species and structure. This necessary yet often difficult step helped to develop shared understanding and agreement for the need to do the work in riparian conservation areas.

Planning Considerations
Include direction similar to Sierra Nevada Framework standard and guideline101 (USDA FS 2004: 63): “Ensure that culverts or other stream crossings do not create barriers to upstream or downstream passage for aquatic-dependent species.” A similar plan component to provide passage at road stream crossings for all life stages of aquatic organisms where the lack of connectivity is considered to be a limiting factor to species viability could be incorporated.

Geographic Considerations
This issue applies to all BioA national forests and grasslands where stream connectivity has been raised as a limiting factor to species viability, excluding those covered by the Sierra Nevada Framework. Currently, a greater urgency for this type of plan component exists on national forests and grasslands occupied by amphibians that are listed or petitioned for listing under the ESA, species that rely on unencumbered passage to carry out life history stages. These units include the Mt. Hood (5 species), Willamette (5 species), and the Gifford Pinchot (3 species) National Forests and the Columbia River Gorge National Scenic Area (4 species), all of which occur within the Cascade Mountains from southern Washington to central Oregon (Reeves et al. 2018).
Key Change Issue 6—Amphibian Migrations and Need for Riparian Reserve Connectivity Across Watershed Divides

While the NWFP and Sierra Nevada Forest Plan include components that promote unencumbered riparian connectivity within and between watersheds, the PacFish and InFish strategies do not (USDA FS and USDI BLM 1994: B-11, USDA FS 2004: 33). The NWFP ACS objective 2 states the following:

“Maintain and restore spatial and temporal connectivity within and between watersheds. Lateral, longitudinal, and drainage network connections include floodplains, wetlands, upslope areas, headwater tributaries, and intact refugia. These network connections must provide chemically and physically unobstructed routes to areas critical for fulfilling life history requirements of aquatic and riparian-dependent species.”

Aquatic and riparian-dependent species include the Oregon slender salamander (photo 45). However, connectivity corridors typically have been viewed in the context of fish passage requirements, where corridors are synonymous with the stream network and associated riparian designations. Thus, providing fish passage within and between other watershed stream networks was viewed as a primary objective, and because road stream crossings were found to restrict connectivity, culvert replacements or removals were targeted to reunite stream connectivity. This view of corridors addresses connectivity within and between watersheds through the stream network, but it does not provide for connectivity over watershed divides.

Current research has shown that aquatic-dependent species such as salamanders travel over watershed divides and require connectivity corridors with downed trees (Olsen and Kluber 2014) to access riparian areas (Olsen and Burnett 2013). This has broadened the definition of connectivity between watersheds. The aquatic strategies in the BioA area, therefore, require clarification to directly address the migration needs of aquatic species, such as salamanders, or terrestrial species, such as martens and fishers, which require migration corridors over watershed divides that connect riparian reserves in different watersheds (see also “Key Change Issue 5” in the “Terrestrial Wildlife” section above).

Planning Considerations

Refer to BioA Chapter 2 Management Recommendations

Recommendation 2: Address the dynamic nature of ecosystems to better respond to future environmental uncertainties.

Recommendation 3: Update and consolidate the existing aquatic direction processes and analysis requirements.

Recommendation 8: Shift from single-species management to maintaining and restoring habitat for multiple species to manage for ecosystem resilience under future uncertainty.

To better establish migration routes between watersheds, we can create a plan component to establish corridors that connect riparian reserves across watershed divides in priority areas needed for aquatic and terrestrial species migration (figure 47).
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Photo 45. Oregon slender salamander

This amphibian is petitioned as threatened under the Endangered Species Act and lives in moist, closed-canopy areas primarily in the western Cascade Mountains on the Gifford Pinchot, Mount Hood, and Willamette National Forests and the Columbia River Gorge National Scenic Area. (Dede Olson, U.S. Forest Service).

Figure 47. Watershed divide corridors

This figure demonstrates the way designated corridors can connect riparian reserves across watershed divides. (Olsen and Burnett 2013). The arrow points to an example over-ridge area where the distance between headwater riparian reserves in different watersheds is small and over-ridge connectivity may be more easily achieved for aquatic species, such as salamanders, and terrestrial species, such as martens and fishers.

Geographic Considerations

This issue applies to the entire BioA area, especially where species viability is a concern and where amphibians are ESA listed or petitioned for listing. All Forest Service Pacific Northwest Region NWFP national forests are occupied by amphibians that are ESA listed or petitioned for listing. National forests that have the highest number of these species are Mt. Hood (5), Willamette (5), and the Gifford Pinchot (3) National Forests and Columbia River Gorge National Scenic Area (4), all of which occur within the Cascade Mountains from southern Washington south to central Oregon (Reeves et al. 2018).

Key Change Issue 7—Species Distribution and Need for All Lands Restoration

The authors of the NWFP recognized the limitations of forest system lands as the sole source of recovery for at-risk fish stocks, noting that fish habitat spans multiple land ownerships and jurisdictional boundaries (Reeves et al. 2018). This is even more apparent 25 years after the completion of the NWFP with the subsequent listing of 24 additional ESA-listed fish and designation of critical habitat. During the 2015 listening sessions, public participants expressed an interest in aquatic restoration across land ownerships. This need can be highlighted by examining the distribution of designated critical habitat of ESA-listed salmonids across multiple land ownerships. For example, of 20,475 miles of designated critical habitat for ESA-listed fish within the NWFP area, the Forest Service manages about 3,147 miles (15 percent), which is more than any other single landowner or manager; however, 17,328 miles (85 percent) of the critical habitat occurs off National Forest System...
lands and is distributed amongst numerous land owners and jurisdictions. In the McKenzie River sub-basin, 228 miles of salmonid critical habitat span the landscape; 118 miles are on private land, 105 miles are on the Willamette National Forest; 4 miles are on Bureau of Land Management land, and 1 mile is on State and local land (map 22). It is apparent that recovery of fish populations depends on Forest Service participation in an all-lands, stewardship approach to riparian and watershed management.

Map 22. Critical habitat and all lands restoration

This map of the McKenzie Watershed (Willamette National Forest) depicts a typical scenario across the Bioregional Assessment area, where critical habitat for Endangered Species Act-(ESA)-listed fish species is distributed across multiple land ownerships, and highlights the need for landowners to join forces to recover fish species, in this case the ESA-listed upper Willamette Chinook salmon, upper Willamette steelhead, and bull trout.

Planning Considerations

Elevate the need for Forest Service participation in all-lands partnerships that target recovery of ESA-listed fish, water quality, aquatic and riparian restoration, and riparian area connectivity across land boundaries and jurisdictions.

32 Designated critical habitat miles were not available for southern Oregon/northern California coho, a coho stock that occurs on the Rogue River-Siskiyou, Klamath, Shasta-Trinity, and Six Rivers National Forests.
Geographic Considerations
Focus on national forests with the greatest overlap of key watersheds, ESA-listed fish, water quality issues, and existing partnerships. Additional analysis is required to highlight priority areas.

Call-out box 21. All-lands aquatic conservation

Since the 1994 Northwest Forest Plan, community-based watershed restoration partnerships, which include national forests, have emerged and multiplied. Many have adopted an “all-lands” stewardship approach by which integrated assessments and restoration are applied across federal, state, and private lands to improve fish habitat and water quality. Organizations that regularly serve in this role include conservation districts, water boards, regional fish enhancement groups, and watershed councils.

The McKenzie Watershed Council in Oregon is comprised of many partners, including watershed residents and landowners, Willamette National Forest, Eugene Water and Electric Board, Oregon Department of Fish and Wildlife, Bureau of Land Management, U.S. Army Corps of Engineers, McKenzie River Trust, Upper Willamette Soil and Water Conservation District, Lane Council of Governments, Weyerhaeuser Company, McKenzie Watershed Stewardship Group, and the McKenzie Collaborative. The Forest Service offers unique technical skills and funding for many of the council’s projects, including stream restoration, fish passage restoration, and outreach efforts throughout the watershed. This engagement demonstrates an all-lands restoration approach, where partners work through land boundary and jurisdictional issues to improve watershed conditions for fish, water, wildlife, and local communities.

Key Change Issue 8—Road Management and Need to Account for Full Suite and Extent of Natural Disturbance Regimes

Currently, about 86,000 miles of roads, the majority of which were constructed before 1994, cross national forests and grasslands in the BioA area and contribute to substantial legacy impacts to terrestrial and aquatic systems. There is a moderate to high probability that road densities in 72 percent of BioA area sub-watersheds have substantially altered local hydrologic regimes (timing, magnitude, duration, and spatial distribution of runoff flows), which increases the likelihood of sediment transport from roads into streams.

Further impacts include creation of passage barriers to aquatic organisms at culverts and restricting watershed process, such as debris flows, downstream large wood movement, and stream channel migration (USDA FS 2011). Even more, increased flood frequency and magnitude, resulting from climate change, is expected to exacerbate hydrology impacts through increased occurrence of landslides and debris flows (Reeves et al. 2018). The NWFP highlighted road impacts to aquatic systems, the need to reverse these impacts, and included standards and guides to direct restoration priorities (USDA FS and USDI BLM 1994: B-31). Existing plans require clarification and revision, given increased understanding of debris-flow dynamics, other ecological processes, and climate change.

Planning Considerations

Refer to BioA Chapter 2 Management Recommendations

Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.

Recommendation 2: Address the dynamic nature of ecosystems to better respond to future environmental uncertainties.

Recommendation 3: Update and consolidate the existing aquatic direction processes and analysis requirements.

Furthermore, consider updating existing plans and direction to guide road management in a manner that aids in the restoration of natural disturbance regimes, such as debris flows (photo 46), and other ecological processes, such as channel migration. Updated plans would have heightened influence in areas important to recovery of ESA-listed aquatic species, water quality, and where climate change is expected to result in more and extensive flooding. Prioritization of actions could be coordinated with federal and state fish and water quality recovery programs.
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Photo 46. Debris flow in a forested stream

This debris flow, located on the Umpqua National Forest, contains large wood that captures boulders, cobbles, and other sediments as it moves downstream. While often seen as a destructive force, this disturbance event is important for formation of aquatic habitat and other watershed functions.

Geographic Considerations

In the BioA area, three groups of national forests and grasslands stand out as having high percentages of sub-watersheds where road densities pose risks to hydrologic regimes.

- **Group 1**—High Urgency for Action: Siuslaw (100 percent), Gifford Pinchot (94 percent), Umpqua (94 percent), Willamette (83 percent), Mt. Hood (76 percent), and Olympic (62 percent) National Forests. These national forests occur in debris-flow-prone areas, contain relatively high numbers ESA-listed fish species, and are situated in areas with moderate to high precipitation levels, a trigger to landslides and subsequent debris flows.

- **Group 2**—High Urgency for Action: Shasta-Trinity (72 percent), Six Rivers (61 percent), Mendocino (60 percent), and Rogue River-Siskiyou (52 percent) National Forests. These national forests occur in debris-flow-prone areas, most contain relatively high numbers ESA-listed fish species and are situated in fire-prone areas that can be more susceptible to landslides and subsequent debris flows.

- **Group 3**—Moderate Urgency for Action: Fremont-Winema (99 percent), Ochoco (99 percent), Okanogan-Wenatchee (84 percent), and Lassen (74 percent) National Forests. This group of national forests have high percentages of watersheds where road densities can impact hydrologic regimes but are less susceptible to landslides and debris flows.
Key Change Issue 9—Climate Change and Need to Account for Impacts to Hydrologic Regimes

The climate is changing and is projected to continue to change hydrologic regimes in the BioA area. For example, precipitation is projected to increase in the winter and decrease in summer, resulting in lower late summer streamflow (streamflows are also projected to be affected by reduced snowpack and water storage), and increased stream temperatures; these effects are expected to particularly apply to stream systems with little groundwater input and shade (Isaak et al. 2017, Luce et al. 2014, Wenger et al. 2010).

Warmer temperatures will result in less snowpack and water storage, more frequent and larger winter floods (also affected by shift in timing of precipitation), and rain-on-snow events in some high-elevation streams where ecosystems are not adapted to these types of floods. Given this, landslides and debris flows are expected to increase (Reeves et al. 2018). Further, lower mean annual streamflows—lower overall water supply—is also projected (Wenger et al. 2010) and will be most prominent in the south and east parts of the BioA area. Conversely, projected increases in water supply are expected in the western Washington Cascades. There is also a predicted increase in channel forming flows known as bankfull flows, which could influence future stream geomorphology (shape and physical function of stream systems) (Wenger et al. 2010). Notable ecological consequences include a reduction in cold-water habitats required to support native trout, steelhead, and salmon, coupled with an expansion of warm-water habitats that accommodate expansion of aquatic invasive species (plants and animals) (Reeves et al. 2018).

Planning Considerations

Create desired conditions and other land management plan components based on projected rather than current hydrologic conditions. Focus on identification of refuge areas for aquatic species, especially in stream types for which groundwater flow can be restored and maintained as well as areas most susceptible to spread of aquatic invasive species. This climate-informed land management plan direction could assist with the selection of key watersheds, development of restoration action plans, and standards and guidelines for road stream crossings and stream restoration projects.

Refer to BioA Chapter 2 Management Recommendations

Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.

Recommendation 2: Address the dynamic nature of ecosystems to better respond to future environmental uncertainties.

Recommendation 3: Update and consolidate the existing aquatic direction processes and analysis requirements.
Geographic Considerations

Climate change impacts in general are projected to be greatest in southern Oregon, northern California, and the east Cascades (see the “Climate Change” section for specifics). It is expected the impacts to hydrology will be most prominent in the same areas. Climate change impacts are associated with several key change issues described in this section, primarily the following: Lack of detailed desired future conditions and disturbance regimes; key watershed alignment with new science; and road management and natural disturbance regimes.

Call-out box 22. Aquatic invasive species

Aquatic invasive species can significantly alter lake, river, stream, and wetland ecosystems. An example of the impacts of an aquatic invasive species is the tui chub infestation and subsequent removal effort at Diamond Lake on the Umpqua National Forest. The tui chub, inadvertently introduced into the lake in the 1950s, interrupted the lake’s food web, resulting in persistent toxic algae blooms, negatively affecting the lakes renowned trout fishery. To address the issue, the lake was partially drained, treated with rotenone, which killed all the fish, and then restocked with trout. The tui chub has continued to plague the fishery throughout the years, and the lake was treated with rotenone again in the mid-2000s.

The key to controlling aquatic invasives is to avoid introduction in the first place. The Forest Service, Oregon Department of Fish and Wildlife, and Washington State Department of Natural Resources have coordinated policy and enforcement efforts to protect National Forest System lands from aquatic invasive species, including plants and animals.

The Northwest Forest Plan Aquatic Conservation Strategy does not explicitly consider invasive species, and a narrow interpretation of this strategy can be an obstacle to efficient treatment of invasive species in riparian and aquatic habitats. Recent plan revision efforts outside of the BioA area have incorporated plan components to address aquatic and riparian invasive species.
Call-out box 23. Potential climate change effects on bull trout

Northwest Forest Plan (NWFP) monitoring has detected decreased summer stream temperatures throughout the NWFP area (Miller et al. 2017) because of increased stream shading and improved stream geomorphology, yet with a changing climate, most streams are projected to increase in temperature over the next 60 years (Isaak et al. 2017). In the Yakima River Basin, for example, this could result in more than a 50-percent loss of available habitat for threatened bull trout because of summer warming above the level at which the species can survive. Depending on the extent of climate change in the area, cases such as this would isolate local bull trout populations, placing them at risk of extirpation. See the figure 48 below.

One area of apparent relief from increasing stream temperatures will be in complex terrain where there is prolonged stream shading combined with subsurface streamflow and cooler groundwater inputs. Maintaining stream buffers for shade and complex channel and floodplain dynamics appear to be important for minimizing the impacts of a changing climate on stream temperatures and fish habitat (Spies et al. 2018). These are examples of climate refugia.

Figure 48. Yakima River Basin critical habitat for threatened bull trout

Projected change in average August stream temperatures in the Yakima River Basin. Comparing baseline (1993–2011) temperature to projected increased stream temperatures in 2080. Highlighted is the threshold of 12 °C.
Wildfire Ecology and Management

Introduction

Fire is an essential and natural part of the BioA area. In appropriate proportion, accounting for frequency, severity and extent, fire is a critical ecosystem driver that helps sustain ecological integrity. When fire is excluded it can lead to uncharacteristic fires that threaten communities and valued resources.

Before Euro-American settlement of the Northwest, fire played a critical role in many of the ecological systems that make up the BioA area. Active fire was promoted by many different American Indian tribes as part of their land stewardship to sustain foods and materials. Because the BioA area encompasses a diverse suite of ecological systems that span vast moisture, temperature, and elevational gradients, fire’s role (both positive and negative) is different for different systems. In designing appropriate management strategies for the BioA area, it is essential to consider the natural role of fire to promote ecological resilience and integrity.

In this report, ecological systems have been differentiated into three primary groups, based on characteristics of fire ecology.

Frequent-Fire Dependent—These are systems in which fire is essential to overall ecosystem functions. Fires were historically (before Euro-American settlement) quite frequent, of low or mixed severity, and served as the primary cornerstone disturbance within these systems. Fire in these systems drove structural and successional dynamics, favoring fire-dependent and fire-adapted species.

Within the BioA analysis area, a total of nine national forests and grasslands have nearly 50 percent or more of their acreage in frequent-fire dependent systems, including the Six Rivers, Klamath, Shasta-Trinity, Rogue River-Siskiyou, Lassen, Mendocino, Ochoco, Modoc, and Fremont-Winema National Forests. The Okanogan-Wenatchee and Deschutes National Forests each have more than a third of their acreage in frequent-fire dependent ecosystems, although both forests contain a broad mix across fire ecology groups.

Fire Diverse (mixed severity)—These are systems in which fire can be important to ecosystem function, but is not the primary driver of successional dynamics, including structure and composition. Historically, fires were moderately frequent, ranging between mixed and high severity in a variety of patch sizes. These highly diverse systems represent a middle gradient between the frequent-fire dependent and fire infrequent ecosystems and are typified by a combination of mixed-severity and stand-replacing fires at medium to long return intervals (approximately 35 to 200 years). The systems generally are associated with forested stands consisting of both fire-tolerant and fire-intolerant species and are most common on the Umpqua, Willamette, Gifford Pinchot, and Mt. Hood National Forests.

Fire Infrequent—These are systems in which fire is not necessarily a part of most ecosystem functions, and when fires do occur, they can be highly impactful. Fires historically were rare or infrequent, of mixed to high severity in large patches, and a rare disturbance within these systems. The infrequent events are an important element associated with the natural range of variation of the long-term successional characteristics.
of the systems. Historically, lightning was uncommon throughout fire infrequent systems, and fuel moistures in these densely canopied forests prevented fire spread. As a result, these areas developed very large, highly complex old-forest structures consisting of species that have little fire tolerance. Fire serves as a stressor in the systems, due to the infrequency and length of time between disturbance events (200+ years), rather than a driver of successional dynamics. These systems are the dominant type on the Siuslaw, Mt. Baker-Snoqualmie, and Olympic National Forests. Also included in this category are systems that have a moderately long fire return interval range (35 to 200 years); these are characterized by stand replacement type events and have a wide, but scattered range throughout the BioA area.

Map 23 shows the three primary fire ecology groups as they occur across the BioA area. No single national forest is represented by just one type or another; however, it is apparent that many forests contain a significant amount of one type or another. While these three primary groups in no way capture the full complexity and pyro-diversity of the assessment area, they represent a useful construct from which to better understand management successes and potential needs for change across the broad landscape.
Map 23. Fire ecology groups and national forests and grasslands in the Bioregional Assessment area

Note that no individual national forest is completely one fire ecology group, but some can be primarily one fire ecology group.
Key Change Issues

**Key Change Issue 1—Reduced Wildfire Risk in the Wildland-Urban Interface**

Within the BioA area, existing land use allocations complicate risk management in the wildland-urban interface; where a mismatch occurs, management direction should be revised to be current with direction that prioritizes protection of populated areas. Existing planning guidance does not adequately address the need for strategic risk mitigation in and around communities and the wildland-urban interface.

Fire impacts to the wildland-urban interface/intermix has continued to increase because of two primary factors: (1) increased urbanization of previously remote locations has increased the proportion of National Forest System lands in the interface and (2) climate change continues to extend and intensify fire seasons. Both complicating factors are expected to continue to increase into the foreseeable future (map 24).

*Map 24. Housing density in the Bioregional Assessment area*

The housing density in and near national forests and grasslands in the BioA area. Where people live (WPL) is measured by housing units (HU) per acre. The graphic on the right highlights the proximity of two communities to forests that are frequent-fire dependent ecosystems with high fire suitability that in 2018 experienced wildfires under conditions that enable extreme fire growth and had severe impacts to these communities and many structures were lost. Basal area loss is an indicator of vegetation fire severity and tree mortality. Dashed line oval in lower right graphic is where about 11,000+ structures were lost in Paradise, California.
Map 25. The 50 most-exposed communities to wildfire risk in Oregon and Washington

Example of post-processing of quantitative wildfire risk assessment (Scott et al. 2018). This depicts the 50 most-exposed communities in Oregon and Washington based on annual burn probability, mapped in dark red. Derived from the Pacific Northwest Quantitative Wildfire Risk Assessment (Gilbertson-Day et al. 2018), to identify community fire threats. Though not currently available for California (in progress), similar data could be displayed to highlight risks to communities and conduct large scale risk assessments.

Planning Considerations
Land use allocations and the associated direction can be adjusted to allow for community protection as a management objective. The 2012 planning rule highlights the need to develop integrated plan components for management areas associated with the wildland-urban interface providing for both ecological objectives and community protection objectives.

Refer to BioA Chapter 2 Management Recommendation
Recommendation 5: Prioritize community and firefighter safety in forested areas near communities at risk from wildfires.
Geographic Considerations

For this key change issue, emphasis should be placed on all forests; forests in primarily frequent-fire dependent ecosystems should be regarded as the highest priority, followed by forests with predominantly fire diverse (mixed severity) ecosystems, and then predominately fire infrequent ecosystem forests.

**Key Change Issue 2—Need for Increased Fire and Vegetation Management in Frequent-Fire Dependent Ecosystems**

In the frequent-fire dependent ecosystems found in the southern, eastern, and northeastern portions of the BioA area, there is an immediate need to increase the pace and scale of work aimed at reducing the risk of undesirable fire effects. This management should be informed by the natural disturbance processes inherent in the development of frequent-fire dependent ecosystems and may involve mechanical treatments paired with extensive prescribed fire and managed wildfire to achieve resource benefits. This management should be targeted where it is strategically and operationally effective both ecologically and from a risk-based framework.

Recent trends in wildfire activity throughout the drier, frequent-fire dependent ecosystems of the plan area demonstrate that current treatments are insufficient in pace and scale to mitigate and restore fire resiliency in this ecosystem (table 5; map 26; figure 49, figure 50). These are systems that evolved with fire and are now highly departed from their historic frequency. Current conditions in these systems are the result of more than a century of fire exclusion combined with grazing and historic timber harvest that often removed the largest, oldest, and most fire-resistant stands of trees. This has resulted in forest conditions that are very different from their natural range of variation, with an overabundance of dense stands of young and immature trees. These highly departed conditions have facilitated a shift in fire effects, resulting in an increase in high-severity fire where historically it was uncommon, which then results in an increased risk of uncharacteristically severe fire types and size of high-severity patches during conditions that are favorable to high-intensity fire and explosive growth.

Despite the highly altered conditions present in many of the frequent-fire dependent ecosystems, there are opportunities to increase the use of managed fire to achieve desired ecological effects. Managing fire for multiple objectives affords opportunities, where and when appropriate, to use fire as a tool to aid in the restoration of departed ecosystems. While some landscape settings currently lend themselves to managing natural ignitions to achieve resource benefits, additional strategic investments in mechanical treatments or prescribed fire applications can make this important and cost-effective tool more widely applicable.

The costs associated with managing and mitigating the direct and indirect impacts from wildfires have increased. This is especially true for large-scale and long-duration high-severity fires that are becoming more common. Opportunities exist to reduce overall suppression costs by using a strategic risk-based decision support process to inform wildfire management response and manage fire as a natural ecological process when and where it can be done safely, taking into account values at risk (Thompson et al. 2015). Managing wildfire based on time and conditions (right place, right time, and right conditions) is an important tool to assist in the reduction of undesirable losses resulting from uncharacteristic
wildfires. In the current and projected climate and environment, using all tools, including managed wildfire to strengthen ecosystem resiliency, will be necessary.

Challenges exist in managing wildfire across agency and multi-ownership lands (Charnley et al. 2018, Zald and Dunn 2018). Upfront recognition and acknowledgment of the conflicts and tradeoffs of managing wildfire can be evaluated through risk-based planning and analysis process (for example, quantitative wildfire risk assessments) (Finney 2005; Calkin et al. 2010, 2011; Scott et al. 2013). Reducing wildfire risk to maintain and protect community and socioeconomic health, alongside other resource values, such as wildlife habitat, watershed integrity, and timber producing markets will have tradeoffs. This may include leveraging risk-based analyses and planning to direct limited funding to wildfire risk reduction and restoration activities near communities, major infrastructure, and other values that are at high risk of being affected by wildfire with great loss and negative consequences. By focusing wildfire risk reduction and restoration activities on and around areas that are highly valued, such as communities at risk, we reduce capability to prioritize and implement funded restoration activities in areas that are away from highly valued areas at risk. This is where a strategic wildland fire management plan approach (including quantitative wildfire risk assessment) helps in identifying areas on the landscape where the potential for managing wildfire for multiple objectives to support for restoration needs may be achieved. This type of pre-decisional assessment and planning primes time-sensitive decision processes by identifying areas where decision-makers can opt to manage a wildfire for multiple objectives or not. Of course, the decision whether to manage a wildfire for multiple resource objectives will continue to be based on a current risk-based decision process for each wildfire incident, where the results could have greater variability of fire effects ranging from undesired to desired.

Planned and identified tradeoffs in managing wildfires through risk-based decision planning, plan components, and agreements would enable and facilitate a timely and appropriate response to implementing a plan to manage wildfire for multiple objectives.

Planning Considerations
Develop plan direction and desired conditions that are informed by ecology and promote natural fire adaptations to improve resilience through the use of prescribed and managed wildfire.

Refer to BioA Chapter 2 Management Recommendations
Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.

Recommendation 2: Address the dynamic nature of ecosystems to better respond to future environmental uncertainties.

Recommendation 5: Prioritize community and firefighter safety in forested areas near communities at risk from wildfires.

Recommendation 6: Recognize that fire is a natural process and plays an important role in reducing the risk of uncharacteristic fire and in promoting ecosystem health.
Recommendation 7: Expand the use of timber harvest as a restoration tool to provide economic and social benefits to communities.

Recommendation 8: Shift from single-species management to maintaining and restoring habitat for multiple species to manage for ecosystem resilience under future uncertainty.

Recommendation 9: Promote active management in plant and animal habitats to restore and encourage ecological resilience.

**Geographic Considerations**

Within the BioA area, a total of nine national forests and grasslands have nearly half or more of their acreage in frequent-fire dependent systems, including the Six Rivers, Klamath, Shasta-Trinity, Rogue River-Siskiyou, Lassen, Mendocino, Ochoco, Modoc, and Fremont-Winema National Forests. The Okanogan-Wenatchee and Deschutes National Forests each have more than a third of their acreage in frequent-fire dependent ecosystems, though both of these forests contain a broad mix across fire ecology groups.

**Table 5. Ecological departure from reference condition for National Forest System lands within the Bioregional Assessment area by fire ecology group**

*Note the greater proportional acreage in high and moderate ecological departure in frequent-fire dependent ecosystems. Values are presented by forest at the end of the “Background Information to Support Key Change Issues” section.*

<table>
<thead>
<tr>
<th>Fire Ecology Group</th>
<th>Ecological Departure “Low (acres)”</th>
<th>Ecological Departure Moderate (acres)</th>
<th>Ecological Departure High (acres)</th>
<th>Ecological Departure Total (acres)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Frequent-fire dependent</td>
<td>1,125,388</td>
<td>7,896,770</td>
<td>2,986,012</td>
<td>12,008,170</td>
</tr>
<tr>
<td>Fire diverse (mixed severity)</td>
<td>4,840,502</td>
<td>3,296,774</td>
<td>180,791</td>
<td>8,318,066</td>
</tr>
<tr>
<td>Fire infrequent</td>
<td>1,844,456</td>
<td>3,090,506</td>
<td>141,397</td>
<td>5,076,359</td>
</tr>
<tr>
<td>Total</td>
<td>7,810,346</td>
<td>14,284,050</td>
<td>3,308,199</td>
<td>25,402,595</td>
</tr>
</tbody>
</table>
Map 26. Ecological departure of fire ecology groups panel map

A geographical representation of the current ecological departure (from reference condition) within the Bioregional Assessment area. Each panel represents a different fire ecology group. Note the concentration of moderate and high ecological departure in frequent-fire dependent ecosystems.
Figure 49. Total annual acres burned by severity class on U.S. Forest Service lands in the Bioregional Assessment area, 1985–2018

Note the widely variable nature of area burned and the apparent increased burn area and regularity of fire on an annual basis.
Figure 50. Mean acres of disturbance for frequent-fire dependent ecosystems by national forest, relative to natural range of variation

Totals are based on 2008–2018 data. The gap between the black bar and the colored bar is the gap between the amount of fire that historically, or naturally, was on the landscape and the amount of fire, wildfire or prescribed, and hazardous fuels treatment that is currently on the landscape. This illustrates the gap that needs to be accomplished to restore ecosystems, and fire's role in those ecosystems.

**Key change issue 3—Develop and Articulate Appropriate Desired Future Conditions**

Management direction should be developed based on a comprehensive understanding of ecological dynamics and fire adaptation. Desired conditions that are developed to promote ecological integrity and resilience should be paired with site potential and underlying ecosystem dynamics. These desired conditions and associated management direction should be applied specifically to where these systems occur on the landscapes.

Fire plays a variable role in different biophysical settings of the BioA area both historically and contemporarily. Land management plan direction should better differentiate between these systems and better align management with the goal of improving sustainable or resilient conditions across the variable landscape.
Additionally, fire-diverse systems are highly variable across the BioA area, and management should be tailored to local conditions to best address this variability, accounting for site potential and natural disturbance regime. This is especially true for desired conditions relating to patch size and patch distribution, where local topography and spatial juxtaposition of vegetative systems have a large influence on natural conditions.

Planning Considerations
Move away from the dry versus moist dichotomy that is prevalent in the NWFP. Replace it with a more comprehensive classification of ecosystems based on their natural disturbance regimes.

Refer to BioA Chapter 2 Management Recommendations
Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.

Recommendation 2: Address the dynamic nature of ecosystems to be better positioned to respond to future environmental uncertainties.

Recommendation 6: Recognize that fire is a natural process and plays an important role in reducing the risk of uncharacteristic fire and in promoting ecosystem health.

Geographic Considerations
Frequent-fire dependent ecosystems are found primarily in northern California, southwestern Oregon, and along the east side of the Cascade Range in both Oregon and Washington (map 23).

Key Change Issue 4—Reduced Risk of Uncharacteristic Wildfire Events
Current vegetation conditions in the BioA include forest structures that are outside of natural ranges of variability primarily in the frequent-fire dependent ecosystems. This has resulted in contemporary fires that are often uncharacteristic and can negatively affect ecological integrity. Climate change is projected to exacerbate the conditions that contribute to uncharacteristic fire effects.

Increased mechanical treatment may, in some settings, mitigate against the negative effects of uncharacteristic stand-replacing fire, especially in drier systems. Investments in reducing and modifying fuel configurations associated with contemporary conditions will require periodic maintenance using a combination of prescribed fire, managed natural fire or repeated fuels reduction treatments at levels higher than those occurring under current management. Without this maintenance, resiliency will be lost and the risk of uncharacteristic stand-replacing fire in the frequent-fire dependent ecosystems will again increase.

Planning Considerations
Management strategies should be developed and applied to reduce the risk of extreme wildfire events across the BioA area.
Chapter 3: Caring for the Land

Refer to BioA Chapter 2 Management Recommendations

Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.

Recommendation 2: Address the dynamic nature of ecosystems to be better positioned to respond to future environmental uncertainties.

Recommendation 5: Prioritize community and firefighter safety in forested areas near communities at risk from wildfires.

Recommendation 6: Recognize that fire is a natural process and plays an important role in reducing the risk of uncharacteristic fire and in promoting ecosystem health.

Recommendation 7: Expand the use of timber harvest as a restoration tool to provide economic and social benefits to communities.

Recommendation 8: Shift from single-species management to maintaining and restoring habitat for multiple species to manage for ecosystem resilience under future uncertainty.

Recommendation 9: Promote active management in plant and animal habitats to restore and encourage ecological resilience.

Geographic Considerations

Frequent-fire dependent ecosystems are located primarily in northern California, southwestern Oregon, and along the east side of the Cascade Range in both Oregon and Washington (map 23).

Key Change Issue 5—Opportunity for Integrated Vegetation Restoration

Land management allocations should promote the potential synergy between noncommercial mechanical treatment, timber harvest, and fire (see also “Sustainable Timber”). The original NWFP land use allocations were broadly designed to separate the land into conservation- versus production-focused areas. The majority of the timber volume produced on National Forest System lands today comes as byproduct of vegetation management that focuses on fuel reduction, restoration, and resiliency. An opportunity exists to better align these activities across land use allocations. Doing so will require aligning objectives with science-based, ecologically appropriate treatments and obtaining the public trust and approval to manage in such a fashion. This will require shifting some timber harvest away from a volume-production focus and toward an ecologically sound, outcome-driven approach.

Prescribed fire and managed wildfire to promote resource benefits are important tools for achieving desired conditions and can be paired with timber harvest where practicable to obtain optimal results. The ability of mechanical treatments alone to mimic fire on the landscape is varied across the BioA area and falls short of producing the full effect of fire as a natural and important disturbance in these landscapes. These treatments are an important tool to use in concert with fire, where appropriate, to reintroduce healthy disturbance in many northwestern systems.
This synergy is strongest in the dry, frequent-fire dependent systems, where mechanical thinning is often essential before first entry with fire to mitigate the potential for high-severity fire and unwanted effects associated with fire occurrence in the current, dense fuel configurations.

Outside the dry, frequent-fire dependent regime systems, mechanical treatments can be used to create settings where fire can play a more natural role on the landscape.

In places where managing fire is difficult due to the occurrence of communities, homes, and highly valued resources and assets in settings where they are prone to fire degradation, mechanical treatment should be used to mitigate risk and exposure. In some places, this has the potential to produce timber volume as hazardous fuels are removed. This synergy can be used to help defray and support the cost of initial fuel reduction treatments where it applies. It should be recognized that long-term maintenance of these strategic fuel treatments likely will not produce significant volume. Future maintenance of these critical fuel reductions will need to be supported without this cost defrayment.

With regards to complex early-seral conditions, mechanical treatments tend to do a poor job of mimicking the effects of fire for creating this important habitat type. The need for complex early-seral conditions is a larger and more relevant topic in fire diverse (mixed severity) and fire infrequent systems, though contemporary fire rates in these systems are currently adding complex early-seral conditions naturally. Large patches of complex early-seral were not typical of the frequent-fire dependent regimes. Early-seral conditions in these regimes occurred as small patches (less than 1 to 2 acres) imbedded in a mosaic dominated by relatively open older patches of trees.

**Planning Considerations**
Increase the use of prescribed and managed wildfire as a tool to achieve ecological objectives by pairing it with mechanical treatments and timber production objectives where possible.

*Refer to BioA Chapter 2 Management Recommendations*

Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.

Recommendation 2: Address the dynamic nature of ecosystems to be better positioned to respond to future environmental uncertainties.

Recommendation 5: Prioritize community and firefighter safety in forested areas near communities at risk from wildfires.

Recommendation 6: Recognize that fire is a natural process and plays an important role in reducing the risk of uncharacteristic fire and in promoting ecosystem health.

Recommendation 7: Expand the use of timber harvest as a restoration tool to provide economic and social benefits to communities.

Recommendation 8: Shift from single-species management to maintaining and restoring habitat for multiple species to manage for ecosystem resilience under future uncertainty.
Chapter 3: Caring for the Land

Recommendation 9: Promote active management in plant and animal habitats to restore and encourage ecological resilience.

Geographic Considerations
This potential synergy of activity types is predominantly applicable in the frequent-fire dependent ecosystems that are located primarily in northern California, southwestern Oregon, and along the east side of the Cascade Range in both Oregon and Washington (map 23).

Background Information to Support Key Change Issues
The fire regime classification background, crosswalk, and methods section identifies how the existing data were used to simplify and define new terms to address wildfire across a very diverse landscape to help clarify urgency and need for change issues in current plan components. Defining the wildfire ecology groups was paramount to this process, and this section defines what they are, and in what context they relate to the key points.

In the fire suppression and risk management section, fire management and technological advances that contribute to increased suppression capability provide a recent example of extreme fire events and identified wildfire risk assessment processes are described. Reducing adverse effects from wildfire and smoke provides socio-economic context to the impacts that wildfire smoke have on communities and how growing urbanization of wildlands and climate change are contributing.

The fire policy section provides historical context to how wildfire policy has contributed to the legacy of suppression and the contemporary effects of wildfire on communities, ecosystems, and current policies and direction. This section also identifies and provides context for how wildfire suppression has contributed to and changed forest health and ecosystem resilience to insects, disease, and drought affecting a majority of the BioA area.

Fire Regime Classification
The BioA area spans an enormous geographic area that is home to a diverse suite of ecological systems (see map 12 “vegetation zones” in chapter 3). These ecological systems exist along climatic gradients, including moisture and temperature. Fire is an essential and natural part of the BioA area, and before the contemporary period, fire played a critical role in many of these ecological systems. Understanding the role of natural fire along the climatic gradients as a driver of ecological trajectories and as a moderating, maintaining force is critical to designing appropriate management strategies and supporting ecological integrity and resiliency.

Due to the inherent complexity in ecological systems, fire ecology groups were created for the BioA; these classify fire regimes and their connection to ecosystem function and forest management. The fire ecology groups represent elements of current and reference conditions and the ecology of how fire interacts with vegetation, topography, climate, and fire variables (Agee 1996, Barrett et al. 2010, Schmidt et al. 2002). The fire ecology groups helped characterize the fire departure for the fire regime type, characterize where on the landscape the need for management for restoration and resiliency is most urgent, and use terminology that clearly integrates these concepts for planning purposes.
Chapter 3: Caring for the Land

These groups include:

- Frequent-fire dependent
- Fire diverse (mixed severity)
- Fire infrequent

Figure 51 illustrates how the different classification methods used in the *Synthesis of Science to Inform Land Management Within the Northwest Forest Plan Area* (NWFP Science Synthesis) and NWFP BioA crosswalk to the LANDFIRE fire regime groups were established after approval of the National Fire Plan (2000), which led to a charter to provide a consistent and comprehensive classification for the nation (Rollins 2009).

![Figure 51. Fire regime classification crosswalk for LANDFIRE (Barrett et al. 2010), volume 1 synthesis of science to inform land management within the Northwest Forest Plan area (Spies et al. 2018), and the NWFP Bioregional Assessment fire ecological groupings.](image)

This information is provided for readers who are familiar with LANDFIRE regime groups or the fire regime terminology used in the NWFP Science Synthesis, to facilitate understanding of how those terminologies relate to the simplified fire ecology groups described in the BioA.

Table 6 provides a crosswalk for the terminology between fire ecology groups and fire regimes (Barrett et al. 2010, Hann et al. 2004), and table 7 includes a brief description of each system type and their subtypes. It also shows how each fire ecology group relates to traditional fire regime groups and terminology used in LANDFIRE (based on Hardy et al. 1998, 2001).

Map 23 shows the spatial extent and configuration of these broad fire regime types across the BioA area. Figure 52 depicts the proportional representation of these fire systems, and figure 53 illustrates subgroups on each forest within the BioA area.
Table 6. Relationship between fire regime groups and the fire ecology groupings

This table is another crosswalk to relate some terminology you may be familiar with to the fire ecology terminology used in the Bioregional Assessment.

<table>
<thead>
<tr>
<th>Terminology</th>
<th>Fire Ecology Group</th>
<th>Fire Regime Group</th>
<th>Frequency (data range)</th>
<th>Severity</th>
</tr>
</thead>
<tbody>
<tr>
<td>Frequent-fire dependent</td>
<td>Frequent-fire</td>
<td>Fire regime groups I and II</td>
<td>0–35 years (6–35 years)</td>
<td>Low/mixed and replacement</td>
</tr>
<tr>
<td>Fire diverse (mixed severity)</td>
<td>Fire diverse</td>
<td>Fire regime group III</td>
<td>35–200 years (27–185 years)</td>
<td>Mixed/low</td>
</tr>
<tr>
<td>Fire infrequent</td>
<td>Fire infrequent²</td>
<td>Fire regime groups IV and V</td>
<td>35–200 and 200+ years (36–unknown years)</td>
<td>Replacement and any severity</td>
</tr>
</tbody>
</table>

1. System in fire regime group III but with long mean fire return intervals and limited fuels to carry fire were classified into fire infrequent (for example, biophysical settings [BpS] 10190).
2. Represents primarily stand replacement fire for fire regime group IV, and long mean fire return intervals for any BpS.

Methods

To depict the fire ecology groups for the BioA footprint, the use of existing datasets were tiered from the datasets used in the ecological vegetation zones and subzones and crosswalks to the biophysical settings (BpS). Percentage of seral stage, reference percentage of fire severity type, and reference mean fire return interval are spatially depicted on the landscape at 30m pixel level. The results were used to depict levels of departure, based on mean fire departure and type of fire/frequency for fire regime type and adaptation, using the three-categories (frequent-fire dependent; fire diverse (mixed severity); fire infrequent).

This process follows the same linkage of vegetation subzone mapping and BpS as was conducted in support of the developed ecological departure scoring and associated restoration/succession needs analysis (Haugo et al. 2015, Ringo et al. 2019). Using the same base underpinning for the two products is essential in linking the necessary evaluation of ecological departure to the underlying disturbance regimes and balance/imbalance that contribute to the departure (table 5).
Table 7. Fire ecology groupings for the Bioregional Assessment area
This table provides a more thorough description of the fire ecology groups as split by subgroups for those new to fire regimes and similar terminology.

<table>
<thead>
<tr>
<th>Fire Ecology Group</th>
<th>Fire Ecology Subgroup</th>
<th>Description</th>
</tr>
</thead>
<tbody>
<tr>
<td>Frequent-fire dependent</td>
<td>Driven by fire</td>
<td>These systems naturally develop with frequent, low-intensity fire, and that repeated disturbance is important in modifying vegetative structure and composition through time. This natural disturbance favors fire tolerant species, with structures conducive to low-intensity fire.</td>
</tr>
<tr>
<td>Frequent-fire dependent</td>
<td>Regenerated by fire</td>
<td>These systems are generally fire dependent, but fire naturally regenerates vegetation conditions. Dominant vegetation tends to be fire intolerant.</td>
</tr>
<tr>
<td>Frequent-fire dependent</td>
<td>Tolerant of fire</td>
<td>Systems where frequent, predominantly low-intensity fire was naturally common, but is not critical to maintain and promote natural stand development. Dominant species in these systems tend to be highly tolerant of fire and are long lived.</td>
</tr>
<tr>
<td>Fire diverse (mixed severity)</td>
<td>not applicable</td>
<td>These systems naturally supported predominantly mixed-severity fires. They historically burned relatively infrequently and are comprised of a combination of fire tolerant and fire-intolerant species.</td>
</tr>
<tr>
<td>Fire infrequent</td>
<td>Fire prone – Regeneration</td>
<td>These systems naturally had infrequent, but high-severity fire. These fires were often episodic and correlated with extreme climate and weather events.</td>
</tr>
<tr>
<td>Fire infrequent</td>
<td>Limited fire</td>
<td>These systems naturally had very little fire of any type in them because of fuel configurations and/or climate.</td>
</tr>
<tr>
<td>Fire infrequent</td>
<td>not applicable</td>
<td>No subclass identified.</td>
</tr>
</tbody>
</table>
Figure 52. Percentage of fire ecological group based on historical fire regime by national forest. Note that no forest is all one fire ecology group.

Figure 53. Acres of fire ecology subgroup by national forest within the Bioregional Assessment area

This chart includes the overall acreage of fire ecology subgroup by forest so that the size of some of these areas plays a role in overall impact across the BioA area.
**Frequent-Fire Dependent Ecosystems**

Frequent-fire dependent ecosystems consist of landscapes where wildfire is essential to overall ecosystem functions. Before Euro-American settlement, fires were more frequent, of low or mixed severity, and served as the primary cornerstone disturbance within these systems. Wildfire in these systems drove structural and successional dynamics, favoring frequent-fire dependent and fire-adapted species, which typically were the most fire resistant.

Current conditions are the result of more than a century of fire suppression combined with historic timber harvest that often removed the largest, and oldest stands of trees. This resulted in forest conditions that are very different from their natural range of variation with an overabundance of young and immature, dense stands. These highly departed conditions have facilitated a shift in fire effects, resulting in an increase in high-severity fire where historically it was uncommon.

Within the BioA area, nine national forests and grasslands have nearly half or more of their acreage in frequent-fire dependent systems: Six Rivers, Klamath, Shasta-Trinity, Rogue River-Siskiyou, Lassen, Mendocino, Ochoco, Modoc, and Fremont-Winema National Forests. The Okanogan-Wenatchee and Deschutes National Forests each have more than a third of their acreage in frequent-fire dependent ecosystems, with both of these forests containing a broad mix across fire ecology groups. Table 8 illustrates the relative proportions of each of these forests based on predominant fire regime characteristics.

<table>
<thead>
<tr>
<th>National Forest</th>
<th>Frequent-Fire Dependent</th>
<th>Fire Diverse (mixed severity)</th>
<th>Fire Infrequent</th>
</tr>
</thead>
<tbody>
<tr>
<td>Fremont-Winema National Forest</td>
<td>48%</td>
<td>37%</td>
<td>15%</td>
</tr>
<tr>
<td>Klamath National Forest</td>
<td>82%</td>
<td>15%</td>
<td>4%</td>
</tr>
<tr>
<td>Lassen National Forest</td>
<td>91%</td>
<td>5%</td>
<td>4%</td>
</tr>
<tr>
<td>Mendocino National Forest</td>
<td>93%</td>
<td>6%</td>
<td>0%</td>
</tr>
<tr>
<td>Modoc National Forest</td>
<td>49%</td>
<td>34%</td>
<td>16%</td>
</tr>
<tr>
<td>Ochoco National Forest</td>
<td>71%</td>
<td>24%</td>
<td>5%</td>
</tr>
<tr>
<td>Rogue River-Siskiyou National Forest</td>
<td>65%</td>
<td>31%</td>
<td>5%</td>
</tr>
<tr>
<td>Shasta-Trinity National Forest</td>
<td>88%</td>
<td>10%</td>
<td>3%</td>
</tr>
<tr>
<td>Six Rivers National Forest</td>
<td>86%</td>
<td>13%</td>
<td>1%</td>
</tr>
<tr>
<td>Okanogan-Wenatchee National Forest</td>
<td>36%</td>
<td>23%</td>
<td>41%</td>
</tr>
<tr>
<td>Deschutes National Forest</td>
<td>34%</td>
<td>48%</td>
<td>17%</td>
</tr>
</tbody>
</table>
While the overall amount of wildfire in these systems is significantly less than the amounts that historically drove these systems (table 9), there has been a pronounced shift from low-severity fire to high-severity fire. Figure 54 shows the contemporary percentage of all burned acres by severity class. It also illustrates the relatively high percentage of mixed-severity and stand-replacing fire that has occurred in the frequent-fire dependent ecosystems which historically would have burned predominantly with low-severity fire.

Table 9. Calculated contemporary mean fire return interval in years for frequent-fire dependent ecosystems (all fire severity classes 1993 to 2018)

For those forests with a high proportion of frequent-fire dependent ecosystems, fires are currently burning less frequently than they would have.

<table>
<thead>
<tr>
<th>Forest</th>
<th>Mean Fire Return Interval (MFRI) in Years</th>
<th>Frequent-Fire Dependent</th>
<th>Percent of forest in Frequent-Fire Dependent:</th>
</tr>
</thead>
<tbody>
<tr>
<td>Deschutes National Forest</td>
<td>123</td>
<td>190</td>
<td>124</td>
</tr>
<tr>
<td>Fremont-Winema National Forest</td>
<td>267</td>
<td>866</td>
<td>270</td>
</tr>
<tr>
<td>Gifford Pinchot National Forest</td>
<td>213</td>
<td>29,263</td>
<td>218</td>
</tr>
<tr>
<td>Klamath National Forest</td>
<td>57</td>
<td>513</td>
<td>58</td>
</tr>
<tr>
<td>Lassen National Forest</td>
<td>116</td>
<td>409</td>
<td>312</td>
</tr>
<tr>
<td>Mendocino National Forest</td>
<td>44</td>
<td>55</td>
<td>44</td>
</tr>
<tr>
<td>Modoc National Forest</td>
<td>157</td>
<td>82</td>
<td>135</td>
</tr>
<tr>
<td>Mt. Baker-Snoqualmie National Forest</td>
<td>2,310</td>
<td>15,500</td>
<td>2,964</td>
</tr>
<tr>
<td>Mt. Hood National Forest</td>
<td>1,124</td>
<td>5,980</td>
<td>1,125</td>
</tr>
<tr>
<td>Ochoco National Forest</td>
<td>182</td>
<td>170</td>
<td>181</td>
</tr>
<tr>
<td>Okanagan-Wenatchee National Forest</td>
<td>53</td>
<td>35</td>
<td>53</td>
</tr>
<tr>
<td>Olympic National Forest</td>
<td>588</td>
<td></td>
<td>624</td>
</tr>
<tr>
<td>Rogue River-Siskiyou National Forest</td>
<td>37</td>
<td>123</td>
<td>37</td>
</tr>
<tr>
<td>Shasta-Trinity National Forest</td>
<td>68</td>
<td>136</td>
<td>69</td>
</tr>
<tr>
<td>Siuuslaw National Forest</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Six Rivers National Forest</td>
<td>67</td>
<td>184</td>
<td>513</td>
</tr>
<tr>
<td>Umpqua National Forest</td>
<td>67</td>
<td>106</td>
<td>67</td>
</tr>
<tr>
<td>Willamette National Forest</td>
<td>317</td>
<td>169</td>
<td>316</td>
</tr>
</tbody>
</table>

MFRI (1993-2018) and Percent of BioA Forests

Supplemental Report to the Bioregional Assessment of Northwest Forests
Chapter 3: Caring for the Land

Figure 54. Percentage of severity class for all wildfire in frequent-fire dependent ecosystems by national forest

*The Siuslaw National Forest did not have a wildfire of minimum threshold size to be evaluated by Monitoring Trends in Burn Severity database from 1993 to 2018.*

Figure 55. Fire severity class trend for frequent-fire dependent ecosystems of the Bioregional Assessment area, 1993–2018

*Burn area is increasing and low-severity fire is increasing faster than both high- and moderate-severity fire.*
The frequent-fire dependent ecosystems also generally align with the highest overall burn probabilities. This is a result of increased ignition sources as well as flammability of the systems. Figure 55 illustrates the trend in fire occurrence by severity type for the fire-driven subcategory of the frequent-fire dependent ecosystems. There is an overall increase in acres burned across all severity classes, though the steepest increase has been observed in the unburned-low-severity class. While the low-severity fire is generally beneficial to this ecosystem type, the overall percentage as well as the increase in mixed and stand-replacing fire in these systems illustrates the imbalance in these systems. The environmental condition to host large wildfires is expected to increase with projected climate change (Davis et al. 2017). The effects of these large fires as they move north and upslope are likely to change from historic effects, especially because forest conditions are outside of the natural range of variation (Ringo et al. 2019). Larger fires in more dense, multi-strata forests would reasonably have more severe effects than expected based on historical fire effects. This is especially important to consider, given climate change and the associated changes in periods of critical fire weather.

Fire Diverse Systems (Mixed Severity)

Fire diverse (mixed severity) ecosystems occur on landscapes where fire can be important to ecosystem function, but is not the primary driver of successional dynamics, including structure and composition. Fires historically were moderately frequent, primarily ranging between mixed and high severity in a variety of patch sizes. These systems represent a middle gradient between the frequent-fire dependent and fire infrequent ecosystems and are typified by a combination of mixed-severity and stand-replacing fires at medium to long return intervals (about 35 to 200 years). These systems are highly diverse and are typified by stands consisting of both fire-tolerant and fire-intolerant species. Fire diverse (mixed severity) ecosystems are most common on the Umpqua, Willamette, Gifford Pinchot and Mt. Hood National Forests (table 10).

Table 10. National forests dominated by fire diverse (mixed severity) ecosystems within the Bioregional Assessment area

<table>
<thead>
<tr>
<th>National Forest</th>
<th>Frequent-Fire Dependent</th>
<th>Fire Diverse (mixed severity)</th>
<th>Fire Infrequent</th>
</tr>
</thead>
<tbody>
<tr>
<td>Umpqua National Forest</td>
<td>11%</td>
<td>80%</td>
<td>9%</td>
</tr>
<tr>
<td>Willamette National Forest</td>
<td>1%</td>
<td>75%</td>
<td>23%</td>
</tr>
<tr>
<td>Gifford Pinchot National Forest</td>
<td>2%</td>
<td>64%</td>
<td>35%</td>
</tr>
<tr>
<td>Mt. Hood National Forest</td>
<td>11%</td>
<td>60%</td>
<td>29%</td>
</tr>
</tbody>
</table>

Unlike the frequent-fire dependent systems, these systems show less pronounced impacts from fire suppression (Reilly et al. 2017). Historically, these areas supported closed-canopied and multi-storied systems, and naturally burned with mixed to high severity. This pattern remains, though overall fire in these systems remains below historical averages (table 11). The fire diverse (mixed severity) ecosystems exhibit departure from reference conditions, and the mechanisms for departure are less driven by fire exclusion than past management and harvest. While some loss of system heterogeneity may be a result of fire suppression in these types, legacy timber harvest is a much bigger driver of these ecological departures.
Table 11. Calculated contemporary mean fire return interval in years for fire diverse ecosystems (all fire severity classes, 1993–2018)

<table>
<thead>
<tr>
<th>Forest</th>
<th>Fire Diverse Total</th>
<th>Mean Fire Return Interval (MFRI) in Years</th>
<th>Percent of forest in Fire Diverse:</th>
</tr>
</thead>
<tbody>
<tr>
<td>Deschutes National Forest</td>
<td>333</td>
<td>48%</td>
<td></td>
</tr>
<tr>
<td>Fremont-Winema National Forest</td>
<td>368</td>
<td>37%</td>
<td></td>
</tr>
<tr>
<td>Gifford Pinchot National Forest</td>
<td>1,463</td>
<td>64%</td>
<td></td>
</tr>
<tr>
<td>Klamath National Forest</td>
<td>92</td>
<td>15%</td>
<td></td>
</tr>
<tr>
<td>Lassen National Forest</td>
<td>128</td>
<td>5%</td>
<td></td>
</tr>
<tr>
<td>Mendocino National Forest</td>
<td>45</td>
<td>6%</td>
<td></td>
</tr>
<tr>
<td>Modoc National Forest</td>
<td>154</td>
<td>34%</td>
<td></td>
</tr>
<tr>
<td>Mt. Baker-Snoqualmie National Forest</td>
<td>820</td>
<td>21%</td>
<td></td>
</tr>
<tr>
<td>Mt. Hood National Forest</td>
<td>1,834</td>
<td>60%</td>
<td></td>
</tr>
<tr>
<td>Ochoco National Forest</td>
<td>163</td>
<td>24%</td>
<td></td>
</tr>
<tr>
<td>Okanogan-Wenatchee National Forest</td>
<td>74</td>
<td>23%</td>
<td></td>
</tr>
<tr>
<td>Olympic National Forest</td>
<td>7,871</td>
<td>47%</td>
<td></td>
</tr>
<tr>
<td>Rogue River-Siskiyou National Forest</td>
<td>104</td>
<td>31%</td>
<td></td>
</tr>
<tr>
<td>Shasta-Trinity National Forest</td>
<td>72</td>
<td>10%</td>
<td></td>
</tr>
<tr>
<td>Siuslaw National Forest</td>
<td></td>
<td>18%</td>
<td></td>
</tr>
<tr>
<td>Six Rivers National Forest</td>
<td>69</td>
<td>13%</td>
<td></td>
</tr>
<tr>
<td>Umpqua National Forest</td>
<td>365</td>
<td>80%</td>
<td></td>
</tr>
<tr>
<td>Willamette National Forest</td>
<td>912</td>
<td>75%</td>
<td></td>
</tr>
</tbody>
</table>

*MFRI (1993-2018) and Percent of BioA Forests* 217 33%

Fire exclusion, and to a much larger degree past timber harvests, have changed patch size and continuity as well as reduced the prevalence of complex early-seral conditions on these landscapes. Current landscapes have high proportions in early-seral conditions, especially on private timber lands, but conditions on these post regeneration harvest sites lack many of the important ecological functions of complex early-seral forests created by fire (Reilly and Spies 2015).

Fire as a disturbance agent enhances nutrient cycling and provides fire defects in trees and snags, which provides important ecological benefits and habitats for wildlife species. Mixed and stand-replacing fire can aid in the creation of complex early-seral conditions in these ecosystems since fire, in the right proportions, is the primary driver in this important habitat type.
Impacts of fire exclusion may be more limited in these systems, and the increased risk of large-scale, stand-replacing fire presents unique problems. Because these systems produce significant biomass and support multistoried stand structures, fuel arrangements are favorable for large and intense fires. Compounding this is the contemporary and predicted shift in climatic conditions to favor more extreme fire weather days and longer fire seasons. This combination of factors makes this system at risk for loss due to wildfire. Trends in fire occurrence show an increase in area burned across all severity types for the fire diverse (mixed severity) ecosystems, as shown in figure 56.

![Fire by Severity Class for Fire Diverse (mixed severity) Ecosystems](image)

**Figure 56. Fire severity class trends for fire diverse (mixed severity) ecosystems of the Bioregional Assessment area, 1993–2018**

*Burn area is increasing in fire diverse (mixed severity) ecosystems, but in this case high severity is increasing at a faster rate than low- and moderate-severity burn area.*

By examining projected changes in climate as it relates to the fire environment, it is evident that some portions of the BioA area, including portions of the fire diverse (mixed severity) landscapes, will be subjected to increased fire likelihood (Davis et al. 2017). Portions of the broader landscape where fire diverse (mixed severity) ecosystems are projected to become more suitable to large fire likelihood in the future, are more susceptible to uncharacteristic loss. These areas of convergence are candidates for more intensive management to reduce the potential effects of large-scale fire events (map 27).
Map 27. Projected future large-fire suitability within the fire diverse (mixed severity) areas in 2040

*Large-fire suitability model results for mid-century climate projections based on methods by Davis et al. (2017).*
Fire Infrequent Systems (Infrequent, Stand Replacing, and Limited Fire)

Fire infrequent ecosystems represent systems where fire is not necessarily a frequent component of most ecosystem functions; these include historically wetter ecosystems, high elevation, rocky areas, and environments with sparse fuels. When fires do occur, they can be highly influential and important for system dynamics. Historically, fires were rare or infrequent, of mixed to high severity, and occurred in large patches. These infrequent events are an important element in the natural range of variation in the long-term successional characteristics of these systems. Fire also serves as a stressor in these systems and can be uncharacteristic due to the infrequency and length of time between disturbance events (typically 200 plus years) for their natural range of variation, rather than a driver of successional dynamics.

These areas have not missed natural fire cycles and there is not a widespread need to reintroduce natural fire into these systems, especially because these fires can be hard to control and threaten human values. These systems are highly departed from reference conditions in some places, primarily due to historic and contemporary regeneration harvest and large-patch clear-cuts. As contemporary and projected climate adds system stressors to these forests and given the predicted increase in extreme fire weather events, there is a risk of large-scale stand-replacing fire in these systems.

Concentrated primarily in Oregon’s north-central coastal mountains, the Olympic Peninsula and on the western slope of the north Cascades, are a suite of systems that historically developed without much natural fire. These systems are the dominant types on the Siuslaw, Mt. Baker-Snoqualmie, and Olympic National Forests (table 12).

Table 12. National forests dominated by fire infrequent ecosystems within the Bioregional Assessment area

<table>
<thead>
<tr>
<th>National Forest</th>
<th>Frequent-Fire Dependent</th>
<th>Fire Diverse (mixed severity)</th>
<th>Fire Infrequent</th>
</tr>
</thead>
<tbody>
<tr>
<td>Olympic National Forest</td>
<td>0%</td>
<td>47%</td>
<td>53%</td>
</tr>
<tr>
<td>Mt. Baker-Snoqualmie National Forest</td>
<td>0%</td>
<td>21%</td>
<td>79%</td>
</tr>
<tr>
<td>Siuslaw National Forest</td>
<td>0%</td>
<td>18%</td>
<td>81%</td>
</tr>
</tbody>
</table>

Lightning was uncommon throughout this area, and fuel moistures in these densely canopied forests prevented fire spread. As a result, these areas developed very large, highly complex old-forest structures consisting of species that have little fire tolerance.

Pre-settlement fires, which did occur in these systems, would have been extremely infrequent, occurring only in periods of highly uncharacteristic fire weather conditions and coinciding with infrequent ignition sources. Such fires burned large areas with high severity due to these conditions. Despite the lack of historic fire as a primary driver in these systems, there remains some risk of stand-replacing fire. The conditions under which these systems are most likely to burn (extreme fire weather) are the same conditions that hinder suppression effectiveness.
Areas in productive forests had naturally high fuel loads, and widespread fuel reduction treatments are not an effective way to reduce fire risk in these infrequent-fire forest systems. Some strategic landscape-scale fuel breaks can help with fire management activities, ranging from suppression to resource benefit (or for multiple objectives) in the regimes where fire was less frequent. Additionally, managed natural fires that occur in wilderness areas and that do not threaten other values can maintain patch/stand dynamic mosaics, such as complex early seral-conditions and meadows.

As contemporary and projected climate adds system stressors to these forests and given the predicted increase in extreme fire weather events, one challenge for management will be protecting these areas from large-scale, stand-replacing fire long enough for them to age.

Because these systems represent less fire-prone areas, there is less urgency to resolve the potential management needs for change in these fire infrequent areas.

Table 13. Calculated contemporary mean fire return interval in years (all fire severity classes, 1993–2018) for fire infrequent ecosystems

<table>
<thead>
<tr>
<th>Forest</th>
<th>Fire Prone - Regeneration</th>
<th>Limited Fire</th>
<th>Fire Infrequent Total</th>
<th>Mean Fire Return Interval (MFRI) in Years</th>
<th>Percent of forest in Fire Infrequent:</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>N/A</td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Deschutes National Forest</td>
<td>80</td>
<td>297</td>
<td>187</td>
<td>131</td>
<td>14%</td>
</tr>
<tr>
<td>Fremont-Winema National Forest</td>
<td>274</td>
<td>160</td>
<td>440</td>
<td>189</td>
<td>14%</td>
</tr>
<tr>
<td>Gifford Pinchot National Forest</td>
<td>267</td>
<td>465</td>
<td>1,674</td>
<td>816</td>
<td>15%</td>
</tr>
<tr>
<td>Klamath National Forest</td>
<td>64</td>
<td>930</td>
<td>180</td>
<td>114</td>
<td>4%</td>
</tr>
<tr>
<td>Lassen National Forest</td>
<td>476</td>
<td>395</td>
<td>197</td>
<td>377</td>
<td>4%</td>
</tr>
<tr>
<td>Mendocino National Forest</td>
<td>113</td>
<td>60</td>
<td>46</td>
<td>73</td>
<td>0%</td>
</tr>
<tr>
<td>Modoc National Forest</td>
<td>8,956</td>
<td>424</td>
<td>205</td>
<td>213</td>
<td>16%</td>
</tr>
<tr>
<td>Mt. Baker-Snoqualmie National Forest</td>
<td>699</td>
<td>1,585</td>
<td>2,506</td>
<td>2,069</td>
<td>16%</td>
</tr>
<tr>
<td>Mt. Hood National Forest</td>
<td>324</td>
<td>1,439</td>
<td>846</td>
<td>476</td>
<td>29%</td>
</tr>
<tr>
<td>Ochoco National Forest</td>
<td>28</td>
<td>177</td>
<td>272</td>
<td>200</td>
<td>5%</td>
</tr>
<tr>
<td>Okanogan-Wenatchee National Forest</td>
<td>94</td>
<td>47</td>
<td>181</td>
<td>80</td>
<td>41%</td>
</tr>
<tr>
<td>Olympic National Forest</td>
<td>5,066</td>
<td>3,007</td>
<td>12,531</td>
<td>11,609</td>
<td>53%</td>
</tr>
<tr>
<td>Rogue River-Siskiyou National Forest</td>
<td>80</td>
<td>222</td>
<td>90</td>
<td>88</td>
<td>5%</td>
</tr>
<tr>
<td>Shasta-Trinity National Forest</td>
<td>304</td>
<td>4,715</td>
<td>690</td>
<td>954</td>
<td>3%</td>
</tr>
<tr>
<td>Siuslaw National Forest</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td>81%</td>
</tr>
<tr>
<td>Six Rivers National Forest</td>
<td>76</td>
<td>259</td>
<td>186</td>
<td>99</td>
<td>1%</td>
</tr>
<tr>
<td>Umpqua National Forest</td>
<td>189</td>
<td>9,974</td>
<td>740</td>
<td>237</td>
<td>9%</td>
</tr>
<tr>
<td>Willamette National Forest</td>
<td>135</td>
<td>237</td>
<td>418</td>
<td>159</td>
<td>23%</td>
</tr>
</tbody>
</table>

MFRI (1993-2018) and Percent of BioA Forests

| MFRI (1993-2018) and Percent of BioA Forests | 126 | 83 | 636 | 202 | 23% |

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Fire Suppression and Risk Management, Effects on Communities and Timber

Increasingly, wildfires have directly or indirectly affected communities in and around the Pacific Northwest and northern California, from single homesteads to major cities. Communities that are in or near fire-adapted ecosystems will always have risks from wildfires. Such risks can range from short to long term and from low to high cost, and may be related to smoke, resource changes, economic losses and gains, changes in recreation opportunities, and personal health and well-being. It is well established that since fire suppression efforts began in the early part of the 20th century in the Western United States, suppression efficiency and effectiveness have improved tremendously. This is associated with an increase in the number of suppression forces being staffed, access to improved training and equipment, and continuing advances in technology (Nelson 1979). There has been improvement in the ability to effectively respond to and manage wildfire, which has helped to reduce impacts in some communities and has enabled land managers to be able to look at a range of alternatives when managing fires.

Relatively recent events—such as the 2018 Carr Fire near Redding, California, (229,651 acres burned; 1,604 structures destroyed; 3 lives lost) and the 2018 Camp Fire in Paradise, California, (153,336 acres burned; 18,793 structures destroyed; 86 lives lost)—have illustrated that even with advances in our suppression ability, wildfires can still impact communities directly, albeit under rare or uncharacteristic circumstances. Figure 58 illustrates the fire severity, represented by basal area loss, for these two fires, where initial wind-driven fire growth caused great losses to structures and life. There are many elements that have contributed to the recent change in wildfire size and severity in fire-adapted ecosystems. Some of these elements are climate change, drought, reduction of fire in fire-adapted ecosystems (Haugo et al. 2019), forest densification, differing land management agency (local, state, federal) fire management plans and policies, and continued encroachment and growth of communities into areas susceptible to wildfires.

Figure 57. Fire severity class trends for fire infrequent ecosystems within the Bioregional Assessment area, 1993–2018
Over time, the effects of fire as a natural and human-caused disturbance process in frequent-fire dependent and fire diverse (mixed severity) ecosystems have shifted from beneficial to having increasing impacts to communities and natural resources.

Wildfire risk assessments and analyses have become much more sophisticated and comprehensive, from preplanning quantitative wildfire risk assessments to near real-time wildfire incident decision support (for example, Wildland Fire Decision Support System), because of technological and research advancements and development. Working together with cooperators and neighbors to review and use the results in risk assessments provides a way to collaborate, prioritize, and address the risk from wildfires. There is an increasing number of quantitative wildfire risk assessments completed or in progress, varying in complexity and extent. In the BioA footprint, for example, there is the recently completed Pacific Northwest Quantitative Wildfire Risk Assessment (Gilbertson-Day et al. 2018), which covers all lands for the states of Washington and Oregon. Quantitative Wildfire Risk Assessments serve many purposes and can be used to identify relative exposure to communities. An example application of Quantitative Wildfire Risk Assessment to identify community exposure across Oregon and Washington is presented in map 25.

Reducing Adverse Effects from Wildfire and Smoke
Adverse fire impacts to communities in the wildland-urban interface/intermix areas have continued to increase because of two primary factors: increased urbanization of previously remote locations has increased the proportion of forest lands located within this area, and climate change effects continue to extend and intensify fire seasons. Both of these complicating factors are expected to continue to increase into the foreseeable future, causing concern for communities.

Smoke from wildfires also adversely affects forest communities, resulting in social and economic impacts, including impacts on the health and quality of life of local residents and economic effects to communities that are dependent on recreation visitation when visitors choose to avoid smoke-filled areas.
Prescribed fire can also negatively impact the health and quality of life of local residents. However, these fires are burned under specific “prescriptions” designed to limit smoke impacts on communities, while achieving desired conditions on the ground. The prescriptions include season of burning, wind direction, and the ability to loft smoke high into the atmosphere. Additionally, prescribed burns can result in less smoke during subsequent wildfires in the same area.

**Call-out box 24. Economic and health impacts of smoke**

*There seems to be no good time for smoke. Prescribed fire in the spring and fall can reduce future increases of more intense, longer duration smoke from wildfires during the summer fire season. The effects of smoke during wildfires can be wide reaching. In 2018, many communities in the Bioregional Assessment area experienced long periods of heavy smoke from wildfires. Ashland in southwest Oregon is home to the Oregon Shakespeare Festival, which canceled or moved more than 26 outdoor performances, resulting in estimated losses of $2 million to the local community. This does not include any additional trickle-down impacts on ancillary businesses. Visits to nearby Crater Lake National Park dropped by 22 percent; and uses of other area outdoor recreation businesses declined as much as 45 percent. Ashland Chamber of Commerce noted some member’s sales were 20 to 60 percent lower during this smoke of the summer of 2018.*

**Fire Policy and Management Changes**

The federal wildland fire management policy provides guidance for fire response and management on federal lands and has evolved substantially over the past few decades since around the time the NWFP went into effect. After the establishment of the Forest Service in 1905, fire suppression/exclusion became a primary agency goal for the better part of the 20th century. In 1926, Forest Service objectives were to restrict wildfires to 10 acres, and in 1935 the agency added a temporal element known as the 10 am policy, under which all fires were to be suppressed by 10 a.m. the next day (Husari and McKelvey 1996). After World War II, with a surplus of personnel and a unified nation, advances in suppression continued and firefighting capability improved with the addition of smokejumpers, chainsaws, and bulldozers; fire science and technology and an understanding of fire ecology also advanced (Nelson 1979).

The passing of the Wilderness Act of 1964 (Husari and McKelvey 1996) began to bring about change on how the agency manages fire by recommending fire be allowed to play a more natural role in designated wilderness areas. However, it was not until the early 1970s, that a decisive split in fire management objectives and policy began between suppression and reintroduction of fire (Husari and McKelvey 1996). In 1972, the Chief of the Forest Service signed an exemption letter for the Selway-Bitterroot Wilderness from the 10 a.m. fire policy (Smith 2014), representing the first deviation from the agency’s full suppression emphasis that had been in effect for three quarters of a century.
Chapter 3: Caring for the Land

The National Forest Management Act of 1976 provided additional measures that have assisted in mitigating the fire control paradox\(^{33}\) (Arno and Brown 1991, Boisrame et al. 2017, Calkin et al. 2015, Thompson et al. 2018). When land and resource management plans were first being developed as a result of the 1976 National Forest Management Act, the Forest Service outlined specific management goals and objectives where the role of fire could be described for both natural and human-caused fires. These initial first forest plans underemphasized the important role natural fire ignitions play and failed to adequately address what type of prescribed fire activity would be appropriate.

A clear path forward for managing fire for resource benefits began to materialize, and a national program was initiated for the use of what was then described as “prescribed natural fire” (later “wildfire use”). However, the Yellowstone fires if 1988 paused this program until a review could be completed. In 1989, the Final Report on Fire Management Policy was completed with recommendations to strengthen interagency cooperation in applying fire management programs, review fire management plans, and reaffirm that agency policy describe fire as either a prescribed fire or wildfire, and that natural fires should not be allowed to burn free of prescriptions or appropriate suppression actions.

Modern wildfire management policy began with recognition of wildland fire as a natural ecosystem process. The 1995 Federal Wildland Fire Management Policy and Program Review is the first fire management direction that integrates the broad spectrum of fire responses that are possible, based on sound risk management and land management plan direction. The 1995 policy looks at a range of fire responses, from reintroducing wildland fire to full suppression, as appropriate. Even under this new policy, not much progress was made for several reasons: for example, most forest plans were completed before the 1995 national fire policy and did not include managed wildfire or associated decision space. Also, it takes time to change the culture of fire suppression and management that emphasized fire exclusion or restricted use of fire in the Western United States.

Increases in fire frequency and losses from wildfires during the late 1990s and early 2000s prompted policy milestones, including the National Fire Plan in 2000, identification of wildland-urban interface communities at high risk from wildfire (USDA FS and USDI BLM 2001), and the Healthy Forests Restoration Act of 2003 (PL 108-148). Together, these policies provided land managers planning direction and funding opportunities that helped improve acres treated and protect communities. This is also when the National Cohesive Wildland Fire Management Strategy\(^{34}\) was developed in response to a report by the General Accounting Office regarding catastrophic fires (US GAO 1999).

The 1995 national fire management policy was updated in 2001 with minor changes, to provide greater emphasis on fire program management and implementation. Implementation strategy for the fire plan update came out in 2003, describing three types of fire (suppression, appropriate management response, prescribed fire). Then in 2009, further

\(^{33}\) Wildfire paradox: when fires are suppressed quickly and at a small size to protect ecosystem resources in a fire-adapted ecosystem, the result is greater vegetation fuel loads and connectivity, increasing difficulty in controlling future fires, and the creation of conditions that may be outside the range of historical variation for fire departure for ecosystem fire regime.

\(^{34}\) https://www.forestsandrangelands.gov/strategy/thestrategy.shtml.
clarification is provided in the Updated Guidance for Implementation of Federal Wildland Fire Management Policy, where types of fire were revised from three to two, wildland fire: wildfire and prescribed fire. Also, wildland fire may be concurrently managed for more than one objective.

The Federal Land Assistance, Management and Enhancement Act of 2009 (FLAME Act) required the secretaries of the U.S. Departments of the Interior and Agriculture to submit a cohesive wildland fire management strategy to Congress. The National Cohesive Wildland Fire Management Strategy identified three primary factors that would provide the best opportunity to respond to the increasing size and severity of wildfires: restoring and maintaining resilient landscapes, creating fire-adapted communities, and responding to wildfires. It also required the development of a national wildfire risk assessment process to be able to evaluate wildfire risk at multiple scales, from national to local. The risk assessment process provides a way to be able to prioritize risk reduction to communities, national resources, and fire management, in order to improve decisions in ecosystem resilience, fire management response and risk management, and create fire adapted communities as we learn to live with fire.

The 1994 NWFP is a result of conflicts that occurred between highly contrasting social, economic and environmental values, and represents a long-term cohesive strategy that has successfully moved the Forest Service forward. A more recent example of a cohesive strategy is the Rogue Basin Cohesive Forest Restoration Strategy, developed by The Nature Conservancy and the Southern Oregon Forest Restoration Collaborative (Metlen et al. 2017). This strategy uses a wildfire risk assessment process to quantify the wildfire risk to highly valued resources and assets. It illustrates conflicts that exist spatially between managing fire, reducing community wildfire risk, producing timber, while conserving wildlife habitat, and provides a way to prioritize activities based on high-risk areas. It also incorporates planning scenarios for fires that have the potential to occur before mitigation actions can be accomplished.

**Synopsis of Current Management**

Although wildfire was considered during development of the NWFP, the effort focused primarily on sustaining economic viability of communities, retaining old-growth forest systems, and protecting the northern spotted owl and other old-growth-associated species. The degree to which wildfire planning and management was incorporated in the associated land management direction was somewhat limited. Fire was generally characterized at that time as a stressor on otherwise functioning ecosystems. The role of fire as a critical ecosystem driver from an ecological process perspective, especially in historically fire-prone ecosystems, was not well addressed.

Contemporary land management recognizes the importance of fire as a system driver, as well as a system stressor with associated resource loss. Existing plan direction makes interweaving these two concepts difficult, and the result is a continued emphasis on fire suppression and risk mitigation. While fire suppression and risk mitigation remain agency priorities, an added emphasis on ecological restoration requires better incorporation of managed fire (both naturally and management ignited) to support functional and resilient ecosystems.
In project-level planning, the role of fire is normally incorporated. For example, a recent objective in a project record of decision included “Protect forest ecosystems from high-intensity, stand-destroying wildfires and provide safe locations for fire-suppression personnel.” Few projects are currently including restoring fire as a management objective to improve resiliency and reduce departure. More often, the associated objectives are to mitigate fire impacts, such as reduction of wildfire risk, improve forest resiliency to disturbance, and reduce surface and ladder fuels. These are desirable objectives, but inclusion of fire restoration would be more comprehensive in recognizing the role of fire.

During the project development phase, when alternatives are being developed for management activities, these two concepts often intersect: wildfire as a risk to highly valued resources and assets, and prescribed fire as a natural disturbance process and fuels management tool. Consequently, during the project development stage different layers of land management direction from the forest’s standards and guidelines are identified, and then used in identifying risks and barriers to implementation and where opportunities exist. This is more often a process of paring down areas to treat that removes many areas that should be a high priority due to risk from unplanned wildfire and the need to reduce fire departure but are dropped due to real and perceived constraints in implementing the project. Risk, both of unintended impacts and litigation, is also considered during project planning.

An additional compounding factor for wildfire management has been the multiple, and at times conflicting, management direction resulting from decades of overlapping policy and plans.

**Fire Management and Forest Resilience**

It is important to note that fire exclusion impacts forest resilience in many ways, not just in changes to fire behavior. Wildfire exclusion, grazing, and historic logging activities have created overly dense stands, a loss of species and age diversity, and an altered mix of vegetation across many areas; this results in effects to ecosystem integrity, including increases in susceptibility to insects, pathogens, and weather-induced stresses such as drought (Cochrane 1998, Egan et al. 2010). When tree stress and mortality occur, there can be effects to fire behavior through altered fuel patterns that are different from the fuel patterns left by mortality from fire. Fire exclusion and its legacy have led to negative effects for ecological resilience and integrity throughout the BioA area.

*Fire and the Land are Inextricably Linked—So Too Should Be Our Understanding and Response:*

Fire is an essential ecological process that to differing degrees shapes the ecology and land management responses for many of our ecosystems. Fire also is responsive to other ecological processes and land-use practices. In practice, wildland fire cannot be separated from other aspects of land management. Fire inevitably, and on its own terms if need be, will occur on the landscape. Our ability to prevent it will be temporally and spatially limited; however, by acknowledging its role in a broader ecological and social context, we may be able to shape its occurrence and effects and thereby better live with fire on the landscape. Given the history of fire suppression, we will be challenged to determine how to use all the tools of vegetation management,
including prescribed fire, to not only reduce the risk of catastrophic wildfire but also to maintain, and increase if needed, an ecosystem's resiliency under changing environmental conditions and its capacity to provide the variety of ecosystem services that society demands (Hall et al. 2018: 16–17).

Climate Change

Introduction
Since enactment of the NWFP in 1994, climate change has emerged as an overarching theme in natural resource science and management. Climate change has the potential to affect all ecological and socioeconomic components of the BioA area as well as other objectives for federal forest and grassland managers in this region.

Current land management plans do not directly address climate refugia and mitigation strategies, but there is the opportunity to include these as the Forest Service updates the plans. Incorporating the role of natural processes into our ecological desired conditions will also be important.

For the past few decades, unanticipated conditions associated with invasive species, wildfire, and climate change have begun affecting the sustainability of our forests and grasslands and their ability to provide numerous benefits. This includes benefits to rural and urban communities, as well as to American Indian tribes who rely on national forests and grasslands for maintaining their culture and way of life. It is a challenge under existing land management plans to maintain and restore natural processes, such as fire, which promote ecological resilience. Altered conditions due to a changing climate will impact ecosystems, biodiversity, and the delivery of benefits to people.

Within the BioA area, anticipated effects of climate change will be the greatest in northern California, southern Oregon, and the eastern Cascades. Increases in greenhouse gases and temperature, as well as altered precipitation and disturbance regimes (for example, fire, insects, pathogens, and windstorms), may profoundly affect biodiversity, socioeconomics, and the delivery of ecosystem services within the BioA area over the next century (Dale et al. 2001, Franklin et al. 1991).

The BioA area is projected to enter a novel climate regime during the next century, and conditions are projected to exceed the 20th century range of variability by the 2040s for some portions of the area. Significant warming may occur across the BioA area, although the magnitude of the warming may differ. Increased frequency and intensity of extreme temperature and precipitation events may occur (Davis et al. 2019, Gutzler and Robbins 2011, Williams et al. 2012), and climate extremes, such as acute drought, may have disproportionate effects on vegetation and result in rapid vegetation change.

Water balance deficits are also projected throughout the BioA area. A water-balance deficit is the difference between the atmospheric demand for water from vegetation and the amount of water that is actually available for use. Changes in the magnitude and seasonality of temperature and precipitation patterns will most likely affect vegetation by altering the availability of water in the soil. Water balance deficit indicates there will be drought impacts on trees and is an early sign of potential future mortality.
What is Working Well

**What is Working Well 1—Carbon Sequestration**

Forests and grasslands in the BioA area have great potential to mitigate the effects of climate change through the storage of large amounts of carbon in both live and dead biomass (Smithwick et al. 2002). At current rates, harvest and disturbance have little overall impact on carbon sequestration on federal lands in Oregon and Washington as a whole (Gray and Whittier 2014).

Adaptation and mitigation are essential to strategic planning for the effects of climate change (Millar et al. 2007).

Adaptive management actions at both the stand and landscape scales can reduce vulnerabilities to climate change. Mitigation includes efforts to increase carbon sequestration in forest ecosystems and provide new energy-efficient products and technologies for society (see "Key Change Issues 3—Carbon Sequestration and Need for Increased Emphasis" below).

**What is Working Well 2—Aquatic Refugia**

Restoring watersheds has resulted in improved water quality and streamflow conditions (USDA FS 2017), as well as improved stream temperature and macroinvertebrate diversity (Miller et al. 2017).

**What is Working Well 3—Water Supply from National Forests and Grasslands**

Most of the water supply for drinking water, irrigation, habitat, recreation, and more comes from National Forest System land, and that has been largely maintained or improved through better streamflow and water quality conditions.

Key Change Issues

**Key Change Issue 1—Changing Landscapes and Determining Suitability of Use**

Climate change will continue to alter the composition, structure, and function of forested and non-forested ecosystems in the BioA area (Vose et al. 2012). Climate change is expected to alter vegetation through both direct and indirect effects. Direct effects may lead to changes in mortality, growth, and reproduction, all of which may be sensitive to altered phenology and biotic interactions within and among species (Peterson et al. 2014). Indirect effects are expected to be expressed through increases in the frequency, severity, and extent of disturbances, including drought, fire, insects, and pathogens.

**Planning Considerations**

Consider the impacts of climate change when determining suitability of uses, such as timber, range, and recreation. Develop standards and guidelines or other plan components that would encourage suitable uses based on best available science and projections. Include plan components that encourage the use of best available tools, processes, and science, including adaptive management.
Refer to BioA Chapter 2 Management Recommendations

Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.

Recommendation 2: Address the dynamic nature of ecosystems to better respond to future environmental uncertainties.

Geographic Considerations

The effects of climate change will likely be most pronounced in the southern portion of the BioA area (northern California and southern Oregon) and in the drier forested and non-forested types (eastern Cascades).

Key Change Issue 2—Longer and Warmer Fire Seasons and Need for More Active Management

Fire is an important factor in disturbance regimes in the BioA area. Increases in the frequency and extent of fire are related to longer fire seasons, which are associated with earlier snowmelt, warmer spring and summer temperatures and drought. Since the mid-1980s, there has been an increase in annual area, however, there is growing consensus that fire suppression has led to dry vegetation zones experiencing less fire during this period than they would have historically (Steel et al. 2015).

Planning Considerations

Desired conditions should be developed for these systems based on their natural disturbance regimes and disturbance-succession dynamics that promote and encourage the use of prescribed and managed fire to achieve desired outcomes where appropriate.

Refer to BioA Chapter 2 Management Recommendation

Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.

Additional planning considerations include increasing stand and landscape resilience, which may be accomplished by reducing stand densities, shifting toward fire-adapted species’ compositions through mechanical treatment, using wildfire and prescribed fire, increasing heterogeneity and diversity of patch sizes, using topography to guide treatments, and favoring fire-tolerant species in natural regeneration and planting (table 14).

Table 14. Summary of adaptation and mitigation options for climate change vulnerabilities in the Bioregional Assessment area (adapted from Halofsky and Peterson 2016)

<table>
<thead>
<tr>
<th>Vulnerability</th>
<th>Strategy</th>
<th>Tactics</th>
</tr>
</thead>
<tbody>
<tr>
<td>Increased drought stress</td>
<td>Increase resilience</td>
<td>Reduce forest/stand densities</td>
</tr>
<tr>
<td></td>
<td>Foster genetic and phenotypic</td>
<td>Favor drought-resistant species/genotypes</td>
</tr>
<tr>
<td></td>
<td>diversity</td>
<td>Protect trees adapted to water stress</td>
</tr>
<tr>
<td></td>
<td></td>
<td>Collect seed for future</td>
</tr>
<tr>
<td></td>
<td></td>
<td>Maintain connectivity for natural species</td>
</tr>
<tr>
<td></td>
<td></td>
<td>migration</td>
</tr>
</tbody>
</table>

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### Geographic Considerations

The greatest fire increases are expected in drier forest types of northern California, southern Oregon, and the eastern Cascades (frequent-fire dependent). Most studies project little increase in fire activity in the moist maritime forests, for example Sitka spruce, redwood, and western hemlock forests (fire infrequent).

### Key Change Issue 3—Carbon Sequestration and Need for Increased Emphasis

Although they are not actively managed for carbon storage, the forests within the BioA area store carbon at some of the highest rates and levels in the United States. Annual temperatures are generally highest in areas with the least amount of annual precipitation in the eastern and southern portions of the BioA area and are linked to both increased drought and wildfire risk (map 28 and map 29), which reduce carbon storage stability over the long term. Summer environmental variables such as these illustrate the difference between the hot and dry Mediterranean climate of the southern BioA area and the cooler and wetter summer conditions in the Pacific Northwest. These cooler and wetter forests are better suited for long-term carbon storage. Warming is projected to occur across all seasons, with the greatest temperature increases occurring during summer months. Along the coast, decreases in summer fog may substantially reduce suitable climate for redwood and other coastal species.

Changes in the magnitude and seasonality of temperature and precipitation patterns will most likely affect vegetation by altering the availability of water in the soil. These effects will be most dramatic in the southern and eastern portions of the BioA area, and most of the region is projected to experience increased summer (June through September) water balance deficits by the middle of this century (map 30). Carbon fluxes, or changes from...
one form to another, can be complex but it is important to note that management actions, including wildfire, shift where carbon is located but rarely result in near 100-percent emissions or loss to the atmosphere (figure 59).

Map 28. Range of environmental and climatic variables across the Bioregional Assessment area from 1981 to 2010 (PRISM)

*EPA = Environmental Protection Agency. Adopted from Reilly et al. (2018) with updated date range.*
Map 29. Mean summer temperature, total summer precipitation, summer moisture stress, and summer fog in the Bioregional Assessment area from 1981 to 2010 (PRISM)

EPA = Environmental Protection Agency. Adopted from Reilly et al. (2018) with updated date range.

Planning Considerations

Refer to BioA Chapter 2 Management Recommendations

Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.
Chapter 3: Caring for the Land

Recommendation 2: Address the dynamic nature of ecosystems to better respond to future environmental uncertainties.

Planning components that focus on carbon sequestration should be developed for the wetter, west-side forests of Oregon and Washington; these should include guidelines for wood retention and minimizing soil disturbance, along with an objective of creating stable carbon.

Map 30. Water balance deficits in the Bioregional Assessment area
The water deficit in inches indicate changes in the magnitude and seasonality of temperature, and precipitation patterns will most likely affect vegetation by altering the availability of water in the soil. EPA = Environmental Protection Agency.

**Figure 59. Forest carbon sequestration**

This figure illustrates the importance of carbon sequestration by showing carbon stored in forests as live trees, dead wood, and soil and how these pools change after fire. Carbon amounts and the time since fire are relative and only shown for illustration. (Adapted from Ryan et al. 2010).

**Geographic Considerations**

Carbon is still being stored in forests throughout the BioA area but at a slower rate in forests to the east and south. This is due to slower plant growth rates. but more importantly, due also to higher carbon emissions from wildfires. As the historic fire regimes are restored and continue to shift with climate change it is likely these forests will store less carbon overall, and they will have fewer episodes of high carbon emissions.

**Key Change Issue 4—Soil Productivity and Need for Increased Attention on Soil Carbon Sequestration**

Soils can provide long-term carbon storage. Soil organic carbon is linked with soil productivity, yet little attention has been paid to soil organic carbon in land management planning.

**Planning Considerations**

For long-term soil productivity and carbon storage, objectives for down wood and slash retention in areas with infrequent fire and rapid decomposition can be considered in planning. In areas with more frequent fire, direction to use tools such as broadcast burning or creation of biochar, a form of charcoal, at landings and redistributing biomass across the stand should focus on the outcomes of higher carbon storage/retention on site after
treatments. Focusing on outcomes and objectives in land management plans, rather than specific methodology, allows for management flexibility as additional research into carbon storage informs development of tools and techniques.

Refer to BioA Chapter 2 Management Recommendations
Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.

Recommendation 2: Address the dynamic nature of ecosystems to better respond to future environmental uncertainties.

Recommendation 6: Recognize that fire is a natural process and plays an important role in reducing the risk of uncharacteristic fire and in promoting ecosystem health.

Geographic Considerations
Managing for carbon retention is best focused on areas with infrequent wildfire. Retention of stable carbon through mitigation techniques such as the creation of biochar is possible throughout the BioA area, but most urgent in frequent-fire dependent and fire diverse (mixed severity) areas.

Key Change Issue 5—Species Viability and Need to Identify Refugia to Buffer Climate Change Impacts
Linking isolated habitats to nearby climate refugia, increasing colonization capacity of sustainable reserve networks, and optimizing reserve networks can all help to mitigate projected changes in climate. Climate refugia may enable species persistence during unfavorable climatic conditions and serve as sources for future recolonization, provided that suitable conditions return in the future. Identifying these areas can be challenging, and climate refugia will most likely be found in topographically complex landscapes where microclimates differ because of differences in aspect, shading, insolation, and cold-air drainages (Dobrowski et al. 2011). Therefore, identifying areas that may serve as climate refugia on forests is important (figure 60).

Planning Considerations

Refer to BioA Chapter 2 Management Recommendations
Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.

Recommendation 2: Address the dynamic nature of ecosystems to better respond to future environmental uncertainties.

Recommendation 3: Update and consolidate the existing aquatic direction processes and analysis requirements.
Figure 60. Landscapes and climate change refugia

Illustrated graphic of climate refugia principles and examples of the physiographic and vegetation-based refugia that may experience reduced rates of climate change. Source: Morell et al. (2016). After Spies et al. 2018.

Geographic Considerations
During the past century, average annual temperatures in western Oregon and Washington increased more than in northern California and in the Cascade Range. Overall, the entire BioA area has experienced warming on some level. In Oregon and Washington, precipitation increased the most during spring. Trends in precipitation in northern California are more variable with some areas experiencing decreases in precipitation. The recent drought in northern California (2012-2016) was the hottest and driest period on record, followed by extreme precipitation events and severe flooding.

Future climate projections for Washington and Oregon include warming across all seasons and the possibility of wetter winters and drier summers. Projections for northern California and the eastern Cascades depict drier futures and greater wet and dry extremes during the wet season (October to March). The entire BioA area may see reduced snowpack with more precipitation falling as rain at higher elevations. Heavy precipitation events from warming and shifts in seasonal precipitation patterns may also increase flooding in most of Oregon and Washington (Tohver et al. 2014) and the northern California Coast Range (Kim 2005). The projected future climate in the Klamath Mountains represents conditions of temperature and precipitation not experienced in the recent past (Saxon et al. 2005). Fragmented populations at their range margins, narrowly distributed species, and species with poor dispersal are all vulnerable to declines from losses of climate-suited habitat.
Water-balance deficits are projected throughout the NWFP area, with the greatest deficits in the eastern Cascades, Klamath Mountains, southeastern portion of the Oregon Coast Range, northern portion of the California Coast Range, and southern portion of the western Cascades in Oregon (Littell et al. 2016). The least amount of change is projected in the northern portions of the Coast Range along the Pacific Ocean, and at higher elevations of the Olympic Peninsula and the northern portion of the western Cascades.

Fog frequency in coastal northern California declined by 33 percent during the 20th century (Johnstone and Dawson 2010), and the area experienced lower summertime cloudiness (Schwartz et al. 2014).

Decreases in summer fog along the coast may substantially reduce suitable climate for redwood and other coastal species that depend on it to mitigate summer drought.

**Key Change Issue 6—Drought Stress and Ecosystem Resiliency**

Tree mortality from higher temperatures and drought stress is already occurring over much of the Western United States and is expected to increase (Allen et al. 2010; 2015). Warmer temperatures and increased frequency and length of droughts are likely to increase climate-induced physiological stress on plants (Adams et al. 2009). Old-growth forests may be vulnerable to periods of elevated mortality rates associated with increases in insects and pathogens during drought (Reilly and Spies 2016, van Mantgem et al. 2009). Tree growth and viability will also be affected by warmer winters, earlier snowmelt, and changing water availability.

Native insects and pathogen activity are likely to increase as trees experience more growing season drought, and the magnitude of their effects will likely vary geographically as well as among species (Chmura et al. 2011, Kolb et al. 2016, Sturrock et al. 2011).

The timing of seasons may also change, interacting with both biological and social processes tied to phenology.

**Planning Considerations**

Include adaptive measures to increase stand and landscape resilience in the face of increasing drought stress, insects, and pathogens.

*Refer to BioA Chapter 2 Management Recommendation*

Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.

**Geographic Considerations**

Climate change will likely lead to the loss of some high-elevation species (especially subalpine forests) where warmer winters and earlier snowmelt may increase the potential for drought and water stress, especially toward the southern portion of the BioA area. Wetter forests in coastal Washington in particular may be vulnerable to a continued northward shift of high-wind events and windthrow.
**Key Change Issue 7—Invasive Species and Need to Increase Control Efforts**

The effect of invasive species is one of the primary concerns associated with maintaining ecological integrity across the BioA area. Agency understanding of the ecological and economic impacts of invasive species has greatly increased over the last few decades. Invasive species can have widespread social, economic, and ecological impacts, including negative impacts to native species, reductions in water quality, altered fire regimes, degraded forage quality, adverse effects on human health and well-being, and economic losses. Increasing temperatures may favor spread and introduction of some invasive species, especially grasses in California (Sandel and Dangremond 2012).

**Planning Considerations**

Existing land management plans are quite limited in addressing potential impacts of invasive species and focus primarily on invasive plants. However, the term “invasive species” also includes native terrestrial and aquatic insects, animals, and pathogens that have moved into habitats or areas where they previously did not exist. Land management plans can address the need to manage habitats to reduce and prevent introduction of invasive species. Land management plans can also address proactive invasive species management.

*Refer to BioA Chapter 2 Management Recommendation*

Recommendation 4: Reduce the introduction and spread of plant, animal, and other invasive species.

**Geographic Considerations**

Invasive species, and insect and pathogen activity will occur throughout the BioA area but may be most pronounced in the southern and drier portions.

**Key Change Issue 8—Changing Hydrologic Regimes and Need to Reevaluate and Adjust Riparian Management**

Increases in winter temperatures are linked with decreases in snowpack (Mote 2006), and earlier snowmelt has altered streamflow timing (Hamlet et al. 2005; Jung and Chang 2011; Stewart et al. 2004, 2005). There are also decreases in the proportion of annual precipitation falling as snow (Klos et al. 2014) and decreases in the amount of water contained in spring snowpack (Hamlet et al. 2005). Lower late summer streamflow, affected by reduced snowpack and water storage, creates a potential for warmer stream temperatures in streams with little groundwater input. In the future there may be more frequent and larger winter floods and rain-on-snow events higher in elevation in streams that do not have the capacity to handle intense floods. Lower mean annual streamflows are also projected and may be most prominent in south and east parts of the BioA with projected increases in water supply in the western Washington Cascades. There is also a predicted increase in channel-forming flows, known as bankfull flows, which could influence future stream geomorphology (Wenger et al. 2010).
Planning Considerations
Future land management plan direction can focus on adapting to new and projected future hydrologic conditions. Future land management plans could include monitoring to assess changed hydrologic and riparian habitat conditions. Riparian management area vegetation management to address departed vegetation conditions would mirror departed forest conditions across the BioA area. These are most prevalent in the southern and eastern forests of the BioA area.

Refer to BioA Chapter 2 Management Recommendation
Recommendation 6: Recognize that fire is a natural process and plays an important role in reducing the risk of uncharacteristic fire and in promoting ecosystem health.

Geographic Considerations
The greatest climate change impacts are projected in northwestern California and east Cascades, and, the impacts to hydrology are expected to be most prominent in the same areas. Groundwater drawdown (using more than is replenished every year) is most evident in drier areas of the BioA (south and east); potential impacts to groundwater-fed ecosystems like fens are mostly of concern on the high desert plateau (Modoc and Lassen National Forests). Mean annual streamflow is projected to decrease in the south and east portions of the BioA area and somewhat increase in the western Washington Cascades. Trends in eastern and southern portion of the BioA area indicate at least some drawdown of groundwater levels.

These trends may impact sustainable and available drinking and irrigation water supplies coming from national forests and grasslands in western communities and nationally important agricultural areas like California’s Central Valley, Oregon’s Willamette Valley and the Columbia River valley.
Chapter 4: Northwest Forest Plan Land Use Allocations and Management Direction

Introduction

This chapter addresses broad-scale planning concepts to consider in the Bioregional Assessment of Northwest Forests, including the land allocations described in the Northwest Forest Plan (NWFP) and relevant management direction. Since enactment of the 1994 NWFP, the landscape within the BioA area has changed and many layers of overlapping and adjacent management direction were developed. The system of reserved areas within the NWFP has been successful in reducing the region-wide loss of old forests, improving watershed health and meeting the objective of protecting old-forest aquatic and wildlife habitats on national forest and grasslands lands. However, the following are issues and needs for change where current land management plan direction within the BioA area does not allow managers to address the ecological needs of the varied and changing landscape or promote the opportunities for restoration-based economic stability.

The basic structure of land use allocation (map 31 and figure 61) has worked well to accomplish the intended goals of the NWFP. Owing to changing environmental conditions in the BioA area—for example, increased fire scope and severity, threats to adjacent lands, and climate change—structural adjustments are needed to enable more responsive and sustainable management across the landscape.

35 Other management direction within the BioA area includes four region-wide forest plan amendments (three in Oregon/Washington and one in California). These and their relevance to the NWFP modernization are discussed below.
Map 31. Northwest Forest Plan (NWFP) land use allocations

For land use allocation descriptions, see figure 61.
Figure 61. Northwest Forest Plan land use allocations.

Each allocation has specific direction to help ensure consistent management wherever that allocation occurs.

What is Working Well

**What is Working Well 1—Late-Successional Reserves**

Late-successional reserves were established to protect and enhance conditions of late-successional and old-forest ecosystems and habitat for associated species such as the northern spotted owl and the marbled murrelet. Late-successional reserves overlay a variety of forest vegetation types and fire regimes and incorporate frequent-fire dependent, fire diverse (mixed severity), and fire infrequent ecosystems. Similar management direction applies across all lands within these reserves. Managed late-successional areas are similar to late-successional reserves but are generally located in dry forests. The intent of the NWFP was to allow additional silvicultural and fuels reduction strategies in managed late-successional areas.
Overall, late-successional reserves and managed late-successional areas have been effective in protecting existing dense multi-layered, old-forest habitats across the NWFP area and protecting against further loss of old forests on federal lands. Late-successional reserves have largely met expectations for contributing to the conservation of dense multi-layered old forests and the species that depend on them. Late-successional reserves have worked well in combination with other land management designations such as wilderness, wild and scenic rivers, and riparian reserves to provide a network of habitats for fish and wildlife species.

**What is Working Well 2—Riparian Reserves**

Riparian Reserves implement the Aquatic Conservation Strategy (ACS) of the NWFP by overlaying all riparian areas throughout the NWFP area and providing management direction within those riparian corridors; they protect aquatic systems and provide travel corridors to terrestrial species and generally limit management activities within the reserved acreage.

Riparian reserves provide habitat for aquatic and riparian-dependent fish and wildlife species. They protect the integrity of water, providing clean water sources for downstream uses, and serve as connective corridors for multiple species of wildlife.

Riparian management areas have been very effective as a component of the ACS in preventing timber harvest-related degradation of riparian and aquatic habitats. They have frequently served as an effective filter between upland sedimentation and other pollutants and water bodies, protecting aquatic habitat and downstream water resources. Riparian management areas have also effectively protected water and associated riparian areas from solar radiation, moderating water temperatures.

**What is Working Well 3—Matrix**

The intent of the matrix land use allocation under the NWFP was primarily to support desired and probable timber production goals set forth in the plan, while also interfacing with other land use allocations to sustain old forest and riparian areas. Most of the area where timber has been harvested since 1994 has been in matrix.

Matrix lands have contributed to timber production, old-forest conditions, overall watershed improvement, aesthetic and recreation scenery quality, and other valuable ecosystem services.

**What is Working Well 4—Survey and Manage**

The survey and manage standards and guidelines have contributed to knowledge of rare and uncommon late-successional and old-forest-dependent species in the NWFP area, despite the shortcomings described under “Key Change Issue 4” of this section. A focus on survey and manage species has contributed to a management shift toward ecological forestry, such as leaving more dead trees, downed wood, and refugia habitat.  

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36 In this case, “refugia habitat” refers to habitat that offers protection to a segment of a species’ population in the face of changes on the landscape.
**What is Working Well 5—Administratively Withdrawn Areas**

Administratively withdrawn areas include areas that were withdrawn from programmed timber harvest by individual forest plans and subsequent administrative decisions; examples include research natural areas, scenic areas, backcountry motorized or non-motorized areas, and inventoried roadless areas. Timber harvest may be allowed in some areas if determined to be an appropriate management tool to meet the purposes and desired conditions of administratively withdrawn areas.

Inventoried roadless areas are a substantial portion of the late-successional old-forest habitats that benefit many fish and wildlife species. Wide-ranging species, such as mesocarnivores, benefit substantially from inventoried roadless areas, as do many aquatic species.

**What is Working Well 6—Congressionally Reserved Areas**

Congressionally reserved areas are areas that are reserved by congressional designation, including national monuments, wilderness areas, and wild and scenic rivers. Congress has the authority to designate new areas, and significant additions have occurred within the BioA area since the development of the NWFP and other multi-national forest and grassland land management plan amendments.

Congressionally reserved areas continue to provide for a diverse suite of ecosystem services and contribute to the overall recovery of old-forest ecosystems throughout the BioA area. Congressionally reserved areas, such as designated wilderness and wild and scenic rivers, also provide important recreational opportunities for people seeking solitude or adventure.

**What is Working Well 7—Land Management Outside the NWFP Boundary and Other Management Considerations in the Bioregional Assessment area**

PacFish and InFish have resulted in successful protection of water bodies and adjacent riparian habitat. A broad-scale monitoring program established under the PacFish and InFish Endangered Species Act (ESA) process, commonly referred to as “PIBO” (PacFish/InFish biological opinion), continues to provide valuable monitoring feedback to managers throughout the basins.

The concept of historic range of variability (HRV) as a reference for desired management outcomes is contributing to managing for diverse and resilient landscapes, early-seral forests and open stands of large, old trees in dry forest landscapes. The rate of harvest of late structure and old forests on National Forest System lands has slowed significantly. Habitat and connectivity between habitats for species that prefer dense, multi-structured forest has generally increased under the Eastside Screens and Sierra Nevada Framework.

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37 Historic range of variability is the terminology used in the Eastside Screens. The BioA uses the term natural range of variation to describe essentially the same concept to align with the 2012 Planning Rule. Important considerations regarding the application of historic conditions under changing climate are discussed in Ecological Integrity.
Key Change Issues

Key Change Issue 1—Late-Successional Reserves

Critical habitat that has been designated subsequent to the delineation of late-successional reserves, especially critical habitat for northern spotted owl, does not fully align with designated late-successional reserves; this creates management challenges. As late-successional reserves were initially designated to serve as habitat for northern spotted owl, it would make the most sense if northern spotted owl-designated critical habitat and reserve land use allocations (including congressionally reserved areas such as wilderness, riparian reserve and late-successional reserve) were aligned.

Current late-successional reserve management objectives emphasize dense multi-layered forest structural conditions that in some areas are out of balance with the natural range of variation, especially in frequent-fire dependent forests. Current management constraints of late-successional reserves have limited restoration of natural disturbance regimes to promote ecological integrity and ecological resilience in frequent-fire dependent forests. Direction to protect dense multi-layered, old-forest habitat and retain high levels of dead and down wood in late-successional reserves may conflict with other values at risk or management objectives, on adjacent lands, in developed recreation areas, or for permitted special use infrastructure. Despite being designated differently from late-successional reserves, managed late-successional areas also have primarily been managed like late-successional reserves. Conducting additional silvicultural treatments, including fuel reduction strategies, has generally not been accomplished in managed late-successional areas. Current management in late-successional reserves, except for specific exemptions, requires time-consuming regional and inter-agency review, which can hamper efficiency.

Planning Considerations

Alignment of late-successional reserve with designated critical habitat could simplify management direction, while late-successional reserve management direction that is compatible with the diverse landscapes is needed.

Refer to BioA Chapter 2 Management Recommendations

Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.

Recommendation 2: Address the dynamic nature of ecosystems to be better positioned to respond to future environmental uncertainties.

Recommendation 5: Prioritize community and firefighter safety in forested areas near communities at risk from wildfires.

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38 The biological and physical environment that can affect the diversity of plant and animal communities, the persistence of native species, and the productive capacity of ecological systems (36 CFR 219.19).

39 The quality or condition of an ecosystem when its dominant ecological characteristics occur within the natural range of variation and can withstand and recover from most perturbations imposed by natural environmental dynamics or human influence (36 CFR 219.19).
Recommendation 6: Recognize that fire is a natural process and plays an important role in reducing the risk of uncharacteristic fire and in promoting ecosystem health.

Recommendation 7: Expand the use of timber harvest as a restoration tool to provide economic and social benefits to communities.

Recommendation 8: Shift from single-species management to maintaining and restoring habitat for multiple species to help ensure we are managing our ecosystem to be resilient in the face of future change.

Recommendation 9: Promote active management in plant and animal habitats to restore and promote ecological resilience.

**Geographic Considerations**

National forests and grasslands with the most fire-dependent plant communities in the eastern Cascades, southern Oregon, northern and northeastern California frequently face conflicts when designing projects to sustain or improve resilient forest conditions because dense, multi-storied stands have more risk of high-severity wildfire on fire-dependent landscapes. Maintenance of existing old forest and maturation of mid-development forest into old forest may continue with the largest gains in the western Oregon Coast Range.

Losses of old forest within late-successional reserves are most evident in fire-dependent ecosystems due to uncharacteristically severe wildfire. The Rogue River-Siskiyou, Okanogan-Wenatchee, Klamath, Shasta-Trinity, Mendocino, and Umpqua National Forests have experienced significant amounts of high-severity fire within late-successional reserves (Davis et al., in progress).

![Figure 62. Total acres of moderate- and high-severity fire in late-successional reserve, 1993–2018](image)

*In general, national forests and grasslands with the most area in frequent-fire dependent ecosystems are highlighted here as having the most area in moderate- and high-severity fire within late-successional reserves.*
Key Change Issue 2—Riparian Reserves and Key Watersheds

Riparian reserves implement the ACS\(^{40}\) of the NWFP by overlaying all riparian areas in the NWFP area and providing management direction within those riparian corridors. By limiting management activities to those benefitting aquatic and riparian-dependent species, riparian reserves were intended to protect aquatic systems, provide habitat for aquatic and riparian-dependent fish and wildlife species, protect water resources for downstream uses, and provide travel corridors to terrestrial species. Key watersheds provide optimal fish habitat and protect high-quality water sources. Several issues have arisen since the NWFP was developed that need to be addressed:

- The process to complete a watershed analysis, which is required to determine if proposed management activities within riparian reserves and key watersheds are consistent with ACS objectives or for adjustments to riparian reserve boundaries, has been inefficient and lengthy.
- Active management to mimic natural conditions and fire regimes is often avoided in riparian reserves because of stringent standards and guidelines; this may leave these areas vulnerable to uncharacteristic stand-replacement fire.\(^{41}\)
- Fire is an integral component of functioning aquatic systems, and its absence in some areas has resulted in less than fully functioning systems.

Planning Considerations

Refer to BioA Chapter 2 Management Recommendations

Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.

Recommendation 3: Update and consolidate the existing aquatic direction processes and analysis requirements.

Furthermore, consider simplifying the requirements for a watershed analysis to increase efficiency and make it easier to keep them up to date.

Adjust the distribution of key watersheds considering designated critical habitat for ESA-listed fish, high intrinsic potential assessments, bull trout core areas, climate change vulnerability assessments, presence of vulnerable non-ESA-listed aquatic species, and refuge areas for ESA-listed aquatic species identified in federal recovery plans.

\(^{40}\) The ACS-designated buffer zones along in areas where management must focus on maintaining aquatic values, and where roads or logging should receive increased scrutiny to avoid impacts to aquatic species.

\(^{41}\) It should be noted that the need for active management in riparian areas can be found throughout the NWFP area (Reeves et al. 2018). In this example, active management in riparian areas is most urgent in the frequent-fire dependent ecosystems, but also noted as a need throughout the BioA area.
Add geographic-based descriptions of reference conditions\textsuperscript{42} and disturbance regimes\textsuperscript{43} and use of mechanical and fire treatments to help guide land management direction.

**Geographic Distribution and Regional Trends**

Due to the relative density of aquatic and riparian resources, higher percentages of coastal and western Cascade national forests and grasslands are designated as riparian reserves than in eastern Cascades national forests and grasslands or most northern California national forests and grasslands.

National forests and grasslands along the eastern Cascades of Oregon and Washington, southwestern Oregon, and northwest California depend on frequent fire and have high levels of departure from reference conditions. While wildfire is a part of these aquatic systems, the scale of the fires could result in short-term negative consequences on habitat for ESA-listed fish. Native aquatic species in the BioA area have adapted to the effects of fire, particularly if populations are connected, and long-term effects of wildfires are most often beneficial. Benefits include addition of large wood and spawning gravel and renewed, reinvigorated riparian vegetation.

**Key Change Issue 3—Matrix**

Matrix designation includes all acres (about 4 million) in the NWFP that are not within any other land use allocation. If matrix lands are not in a riparian reserve or affected by other NWFP direction, such as survey and manage or matrix and all land standards and guidelines, then they are managed under guidance in individual land management plans. Timber harvest in matrix lands has resulted in production below the estimated probable sale quantity.\textsuperscript{44} One reason for this decline in timber production is that the designation of critical habitat for northern spotted owl incorporated some matrix lands that resulted in conflicting management direction.

**Planning Considerations**

*Refer to BioA Chapter 2 Management Recommendations*

Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.

Recommendation 7: Expand the use of timber harvest as a restoration tool to provide economic and social benefits to communities.

Recommendation 9: Promote active management in plant and animal habitats to restore and promote ecological resilience.

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\textsuperscript{42} A standard or benchmark against which current conditions are compared.

\textsuperscript{43} Cumulative effects of disturbance events across space and time.

\textsuperscript{44} Probable sale quantity reflects the acres available for harvest and expected acre yields as well as land management standards and guidelines.
Additional planning considerations include recalculating appropriate timber output values during forest plan revision to better account for contemporary harvest practices, modern desired conditions, and up-to-date harvest restrictions.

**Geographic Distribution and Regional Trends**
The land management plans on the Gifford Pinchot, Mt. Hood, Willamette, and Umpqua National Forests projected the highest timber outputs under the NWFP, and these forests have experienced significant gaps between projected production and actual output. These are highly productive forests and the timber output does not reflect that potential productivity.

Providing a predictable and sustainable timber supply is a goal of the Forest Service. Implementing restoration and resiliency projects often have timber output as a coproduct.

**Key Change Issue 4—Survey and Manage**
The survey and manage standards and guidelines apply across all land allocations in the NWFP area. The standards and guidelines, adopted to mitigate the impacts of continued timber harvesting in old forests, requires the Forest Service look for rare and uncommon species before timber harvest or other specific project activities identified by the courts as Pechman exemptions.45 If these plants or animals are found, they must be protected/buffered or potential impacts must be otherwise avoided or mitigated.

There is a need to update the NWFP’s survey and manage program to reflect current science and lessons learned through implementation. Survey and manage is out of step with the multiple coarse-filter species conservation approach, as directed by the Forest Service 2012 Planning Rule. Complexity of direction has made it difficult to conduct the annual species review process, which is required as part of the adaptive management requirement of the survey and manage program. Therefore, changing the status of species has not occurred since 2003. Complex and lengthy survey protocols and requirements for managing known sites have been obstacles to forest management and species protection.

Additionally, the survey and manage program requires region-wide periodic strategic surveys to update the program’s species list and make changes to management recommendations. These survey protocols are not implemented as intended and are costly. Even with the positive outcomes of survey and manage stated earlier, several important issues need to be addressed:

- The annual species review is part of the adaptive management design of the survey and manage standards and guidelines, and staffing shortages and complexity of direction have made it difficult to apply. When species are no longer considered rare or uncommon, removing or moving species between categories does not occur or takes a long time.
- Complex and lengthy survey protocols and requirements for managing known sites result in uncertainties and difficulties in designing and implementing forest management activities.

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Region-wide, periodic strategic surveys to fill information gaps in order to update the survey and manage species list and make changes to management recommendations or survey protocols are costly or not implemented as intended.

Photo 47. The red tree vole is a “survey and manage” species

The red tree vole (Arborimus longicaudus) is an arboreal rodent that lives in the canopies of old-growth Douglas-fir forests. Its range extends south from the Columbia River, west of the Cascades; down to northwestern California and serves as a good indicator of old-forest canopy dynamics.

Planning Considerations
Taking a coarse-filter approach can help us manage for the habitat needs of multiple species, including species that are considered imperiled or vulnerable, while some mitigations for specific rare and uncommon species where limited status information exists and or there are persistence concerns are still needed.

Refer to BioA Chapter 2 Management Recommendations
Recommendation 8: Shift from single-species management to maintaining and restoring habitat for multiple species to help ensure we are managing our ecosystem to be resilient in the face of future change.

Furthermore, additional Planning Considerations include that in forest plan revision the 2012 planning rule and Forest Service Handbook 1909.12 direct the application of species of conservation concern.46

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46 A species, other than Federally recognized threatened, endangered, proposed, or candidate species, that is known to occur in the plan area and for which the regional forester has determined that the best available scientific information indicates substantial concern about the species’ capability to persist over the long term in the plan area (36 CFR 219.9(c)).
Geographic Distribution and Regional Trends
Within the range of the red tree vole, particularly toward the center of the range (Umpqua and Willamette National Forests), managers often avoid projects in potential habitat due to management constraints, cost and timing of surveys, and potential public controversy.

On national forests and grasslands in the Klamath province where there is a likelihood of survey and manage species, project activities are substantially reduced owing to timing and cost of intensive requirements.

Key Change Issue 5 – Adaptive Management Areas
The NWFP established ten adaptive management areas (about 1.5 million acres), each with its own specific objectives. The intent of the adaptive management areas was to develop and test new management strategies related to ecological and economic management goals. However, adaptive management areas are subject to the same substantive and process regulations and policies as the rest of NWFP lands, constraining experimental design and implementation. Risk aversion and lack of social acceptance by regulatory agencies, environmental interest groups, and the public in general about implementation of experimental design are among several additional reasons adaptive management areas were underused (Stankey et al. 2006). The goal of adaptive management areas was not fully realized.

Work is needed to create more effective processes and to build trust and collaboration that could more successfully integrate adaptive management into land management plans.

Planning Considerations
While our knowledge about national forests and grasslands and the communities that we serve has grown, uncertainties remain. By improving how we integrate future uncertainty into our land management planning direction, we will be better positioned to manage ecosystems in the face of anticipated change.

Refer to BioA Chapter 2 Management Recommendations
Recommendation 2: Address the dynamic nature of ecosystems to be better positioned to respond to future environmental uncertainties.

Furthermore, universities, other agency research branches, and Forest Service research stations should be involved in land management modernization to determine whether to continue specific adaptive management area objectives.

Key Change Issue 6—Administratively Withdrawn Areas
Administratively withdrawn areas are those that were withdrawn from timber harvest by individual forest plans or subsequent administrative decisions. Examples include research natural areas, scenic areas, backcountry motorized or nonmotorized areas, and inventoried roadless areas. While management activities are not precluded in administratively withdrawn areas or inventoried roadless areas, many management activities may not be analyzed or implemented without regional or national review processes, and the inability to build roads or harvest timber in inventoried roadless areas substantially limits management options in these areas.
Planning Considerations
Designate land use allocations more compatible with inventoried roadless area direction to enable management that is more effective in achieving ecological desired conditions.

Make adjustments to late-successional reserves, matrix, and in some cases, key watersheds to better align with the values and management constraints on inventoried roadless areas.

**Key Change Issue 7—Congressionally Reserved Areas**
Congressionally reserved areas are reserved by congressional designation, and include national monuments, wilderness areas, and wild and scenic rivers. Congress has the authority to designate new areas, and significant additions have occurred within the BioA area since the NWFP was first implemented. Even with the positive results of congressionally reserved areas described earlier in this section, some things are not working well.

Fire management in congressionally reserved areas is complicated due to restrictions of, and additional requirements for, the use of firefighting tools and techniques. Using prescribed and unplanned fires to meet resource objectives within designated wilderness can be difficult due to collateral impacts to recreation resources and air quality, and limited opportunities to control these fires outside the reserved lands.

Wildfire has burned a larger proportion of the congressionally reserved areas within the NWFP in comparison to matrix, riparian reserves, or administratively withdrawn areas. This is likely due to land management choices to not aggressively suppress wildfire that start in wilderness areas, whereas wildfires in other land use allocations are managed with more aggressive suppression tactics.

Some congressionally reserved areas, especially some popular designated wilderness areas, have become overcrowded with visitors, affecting both the natural environment and opportunities for solitude.

Planning Considerations
Develop management direction to enable the use of prescribed or wildland fire in congressionally reserved areas where the natural role of fire is appropriate.

*Refer to BioA Chapter 2 Management Recommendation*
Recommendation 6: Recognize that fire is a natural process and plays an important role in reducing the risk of uncharacteristic fire and in promoting ecosystem health.

**Geographic Distribution and Regional Trends**
Since the establishment of NWFP direction, congressional designations have occurred in Washington, Oregon, and California, with more on some national forests and grasslands or geographic areas than others.

Public engagement in the periodic enactment of congressionally designated areas is evidence of continued commitment and involvement in public land management.
National forests and grasslands closest to population centers, such as the Mt. Hood, Mt. Baker-Snoqualmie, Deschutes, Okanogan-Wenatchee, and Shasta-Trinity National Forests, have the greatest challenges regarding the ability to respond to growing and more diverse recreation use and meeting NWFP goals and direction.

**Key Change Issue 8—Land Management Outside the Northwest Forest Plan Boundary and Other Management Considerations in the Bioregional Assessment**

Approximately 25 percent of the area included in the BioA is not within the footprint of the NWFP. Land management plans have been amended on these acres by several pieces of management direction:

- The Interim Strategies for Managing Anadromous Fish-Producing Watersheds in Eastern Oregon and Washington, Idaho, and Portions of California (also known as “PacFish”), August 18, 1994.
- The regional forester’s Amendment #2 for the Revised Continuation of Interim Management Direction Establishing Riparian, Ecosystem and Wildlife Standards for Timber Sales (also known as “Eastside Screens”), June 5, 1995.
- The Inland Native Fish Strategy Environmental Assessment, decision notice, and finding of no significant impact (also known as “InFish”), dated July 28, 1995.
- The Sierra Nevada Framework Amendment (also known as “Sierra Nevada Framework”) originally signed on January 12, 2001 and revised in October 2004.

These strategies and amendments along with the NWFP and the underlying land management plans often create conflicting management direction that is complex and inefficient to implement. In the bigger picture, the landscape trends are larger than the boundaries of individual land management plans and regional amendments. Therefore, revisions of the land management plans within the BioA area should recognize the successes and challenges of all management direction within that greater landscape area, with an eye toward integration to facilitate consistency and landscape-level management across and outside of the BioA area. Detailed issues include the following:

- Complicated management of terrestrial and aquatic resources on eastern Cascades national forests and grasslands\(^{47}\) and northeastern California national forests and grasslands\(^{48}\) due to the different management direction that applies to different portions of these national forests and grasslands; landscape-level management strategies are difficult to plan and implement.

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\(^{47}\) Okanogan-Wenatchee, Deschutes, Fremont-Winema, and parts of Mt. Hood and Gifford Pinchot National Forests.

\(^{48}\) Lassen and Modoc National Forests.
Elements of the Eastside Screens and the Sierra Nevada Framework, including diameter limits for timber harvest (21 inches in the Eastside Screens, 20 inches in the Sierra Nevada Framework) can hamper the ability of managers to meet objectives for natural range of variation, and are often at odds with desired conditions for resilient landscape conditions that were developed by federal land managers, community collaborative groups, and forest researchers using the best available science.

The mismatch between existing land management plan direction and current landscape conditions resulted in many project-specific plan amendments to meet objectives; amendments can be lengthy and inefficient processes.

Planning Considerations

Refer to BioA Chapter 2 Management Recommendations

Recommendation 1: Maintain and restore ecosystem characteristics and processes by working toward desired conditions that are compatible with the diverse landscapes across the BioA area.

Recommendation 3: Update and consolidate the existing aquatic direction processes and analysis requirements.

The intent of PacFish, InFish and the Eastside Screens was to be interim guidance until clear overarching direction was developed or affected land management plans were revised. Revision efforts should build upon these strategies, while considering landscape-level trends and desired outcomes relative to ecosystem integrity, natural range of variation, fire regimes, and watershed health. Furthermore, integration of management direction to create consistency between the NWFP, PacFish, InFish, the Eastside Screens, and the Sierra Nevada Framework is needed to meet ecological and social objectives.

Modernization Options

Modernization options are outlined at the end of chapter 2 of the BioA. These options are a starting point as the Forest Service engages the public and develops a strategy for updating land management plans across the BioA area. We want to keep and enhance management direction that is working well, but make changes where necessary to meet today’s social, economic, and ecological conditions and challenges on our dynamic landscapes. Large-scale management challenges, such as climate change, affect all the national forests and grasslands across the BioA area. Other challenges, including maintenance of spotted owl habitat and maintaining the role of wildfire in frequent-fire dependent ecosystems, are unique to, or more urgent on, individual or several national forests and grasslands. To increase efficiency in land management, it is important for modernization efforts to create consistent direction for universal challenges, but also to develop direction that recognizes and is compatible with diverse ecosystems and communities.

49 Management recommendations in dry pine and Douglas-fir forests often focus on creation and maintenance of open stands of large trees without the dense understory of shade-tolerant trees that are characteristic of the denser stands preferred by spotted owls.
Potential strategies for modernizing the land management plans in the BioA area include simultaneous plan revision, incremental plan revision, amendments, or some combination of these.

Next Steps
Our 2012 planning rule supports and encourages productive working relationships between the Forest Service and diverse communities, including youth, low-income, and minority populations, and stakeholders, American Indian tribes, and other governments by providing “for a transparent, collaborative process that allows for effective participation” throughout the entire land management planning process.

For several years after finalization of the current planning rule, a small committee of public, state- and local-elected officials, tribes, and youth provided advice and recommendation on implementation of the planning rule. This same committee developed the Citizens Guide to National Forest Planning. In this spirit, we are committed to a participatory collaborative process to accomplish modernization of the existing land management plans. The collaborative effort will bring diverse interests together to explore critical issues and provide meaningful input to our decision process.

Before any formal planning begins, a road map will be provided to inform, engage, and collaborate during each phase of the land management planning process. The Forest Service will also provide tools for gathering and sharing information to keep the public and stakeholders informed and to ensure robust collaboration.

The next step is to reengage with and understand what is important to our customers: the public. Effective engagement will help complete the picture of what we need to be aware of for future planning opportunities. With your participation, we will work together to determine what in the NWFP should be carried forward and what can be improved upon based on new information, today’s issues, and what best meets the needs of our diverse communities and stakeholders.

The Forest Service will stay connected and provide the latest information through various channels of communication to make that as easy as possible. Please visit the NWFP Modernization webpage\(^50\) for more information on the meetings and channels of communication and to learn how you can stay engaged.

Glossary

This glossary is adapted from the *Synthesis of Science to Inform Land Management Within the Northwest Forest Plan (NWFP) Area* (Science Synthesis 2018) to ensure consistency of language between the Science Synthesis, the *Bioregional Assessment of Northwest Forests* (BioA), and this report and to help readers understand various terms used in the documents.

Sources include the Forest Service Handbook (FSH), the Code of Federal Regulations (CFR), executive orders, the Federal Register (FR), and various scientific publications. The authors have added working definitions of terms used in the Science Synthesis and its source materials, especially when formal definitions may be lacking or when they differ across sources.

**active management**—Direct interventions to achieve desired outcomes, which may include harvesting and planting of vegetation and the intentional use of fire, among other activities.

**adaptive management**—A structured, cyclical process for planning and decision-making in the face of uncertainty and changing conditions with feedback from monitoring, which includes using the planning process to actively test assumptions, track relevant conditions over time, and measure management effectiveness (FSH 1909.12.5). Additionally, adaptive management includes iterative decision-making, through which results are evaluated and actions are adjusted based on what has been learned.

**adaptive management area (AMA)**—A portion of the federal land area within the NWFP area that was specifically allocated for scientific monitoring and research to explore new forestry methods and other activities related to meeting the goals and objectives of the plan. Ten AMAs were established in the NWFP area, covering about 1.5 million ac (600 000 ha), or 6 percent of the planning area (Stankey et al. 2003).

**ancestral lands (of American Indian tribes)**—Lands that historically were inhabited by the ancestors of American Indian tribes.

**annual species review**—A procedure established under the NWFP in which panels of managers and biologists evaluate new scientific and monitoring information on species to potentially support the recommendation of changes in their conservation status.

**Aquatic Conservation Strategy (ACS)**—A regional strategy that uses an ecosystem approach to manage and protect riparian and aquatic habitats across the broad landscapes of lands in the NWFP area.

**biodiversity**—In general, the variety of life forms and their processes and ecological functions, at all levels of biological organization from genes to populations, species, assemblages, communities, and ecosystems.

**climate adaptation**—Management actions to reduce vulnerabilities to climate change and related disturbances.

**climate change**—Changes in average weather conditions (including temperature, precipitation, and risk of certain types of severe weather events) that persist over multiple decades or longer, and that result from both natural factors and human activities such as increased emissions of greenhouse gases (U.S. Global Change Research Program 2017).
climate change refugia—Areas that remain relatively buffered from contemporary climate change across time and enable persistence of valued physical, ecological, and sociocultural resources.

collaboration or collaborative process—A structured manner in which a collection of people with diverse interests share knowledge, ideas, and resources, while working together in an inclusive and cooperative manner toward a common purpose (FSH 1909.12.05).

commercial thin—An intermediate harvest with the objective of reducing stand density primarily to improve growth, enhance forest health, and other resources objectives. Treatment can recover potential mortality, while producing merchantable material. Thinning includes the following: chemical (killing of unwanted trees by herbicide application); crown (removal of trees from dominant and co-dominant strata); free (no consideration to crown position); low (removal of trees from lower crown classes); mechanical or row (removal of trees either in row, strips by using a fixed spacing interval); selection (removal of the crown class to favor those in the lower crown classes) (Forest Service Activity Tracking System, app. B).

community (plant and animal)—A naturally occurring assemblage of plant and animal species living within a defined area or habitat (36 CFR 219.19).

community resilience—The capacity of a community to return to its initial function and structure when initially altered under disturbance.

community resistance—The capacity of a community to withstand a disturbance without changing its function and structure.

compatible—Capable of existing together in harmony.

composition—The biological elements within the various levels of biological organization, from genes and species to communities and ecosystems (FSM 2020).

consistent—Marked by harmony, regularity, or steady continuity: free from variation or contradiction.

congressionally reserved land—Land use allocations that have been designated by United States Congress. These include wilderness areas, wild and scenic rivers and some national monuments.

connectivity (of habitats)—Environmental conditions that exist at several spatial and temporal scales that provide landscape linkages that permit (1) the exchange of flow, sediments, and nutrients; (2) genetic interchange of genes among individuals between populations; and (3) the long-distance range shifts of species, such as in response to climate change (36 CFR 219.19).

desired conditions—A description of specific social, economic, or ecological characteristics toward which management of the land and resources should be directed.

disturbance regime—A description of the characteristic types of disturbance on a given landscape; the frequency, severity, and size distribution of these characteristic disturbance types and their interactions (36 CFR 219.19).
**disturbance restoration need**—The area departed from the natural range of variability where a disruption is needed to move existing conditions closer to natural range of variation. These disruption processes include fire, wind, and insects and disease. Disturbance can also be achieved through management tools of thinning and/or prescribed burning (Haugo et al. 2015, DeMeo et al. 2018).

**disturbance**—Any relatively discrete event in time that disrupts ecosystem, watershed, community, or species population structure or function, and that changes resources, substrate availability, or the physical environment (36 CFR 219.19).

**dynamic reserves**—A conservation approach in which protected areas are relocated following changes in environmental conditions, especially owing to disturbance.

**early-seral vegetation**—Forest conditions in the early stages of succession following an event that removes the forest canopy (for example, timber harvest, wildfire, windstorm), on sites that are capable of developing a closed canopy (Swanson et al. 2014). A non-forest or “pre-forest” condition occurs first, followed by an “early-seral forest” as young shade-intolerant trees form a closed canopy.

**complex early-seral forest**—A forest comprised of early-seral vegetation that differs from more simplified early-seral forest in a few key ways. Complex early-seral forest is often naturally occurring. It has high species diversity and is made up of survivors and legacies, including organic structures including live and dead trees that provide habitat for surviving and colonizing organisms. Traditional forestry practices like clear-cutting, salvage logging, and tree planting can reduce species richness and key ecological processes associated with complex early-seral habitat (Swanson et al. 2011).


**ecocultural resources**—Valued elements of the biophysical environment, including plants, fungi, wildlife, water, and places, and the social and cultural relationships of people with those elements.

**ecological conditions**—The biological and physical environment that can affect the diversity of plant and animal communities, the persistence of native species, invasibility, and productive capacity of ecological systems. Ecological conditions include habitat and other influences on species and the environment. Examples of ecological conditions include the abundance and distribution of aquatic and terrestrial habitats, connectivity, roads and other structural developments, human uses, and occurrence of other species (36 CFR 219.19).

**ecological forestry**—An ecosystem management approach designed to achieve multiple objectives that may include conservation goals and sustainable forest management, and which emphasizes disturbance-based management and retention of “legacy” elements such as old trees and dead wood (Franklin et al. 2007).

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**ecological integrity**—The quality or condition of an ecosystem when its dominant ecological characteristics (for example, composition, structure, function, connectivity, and species composition and diversity) occur within the natural range of variation and can withstand and recover from most perturbations imposed by natural environmental dynamics or human influence (36 CFR 219.19).

**ecological sustainability**—The capability of ecosystems to maintain ecological integrity (36 CFR 219.19).

**economic sustainability**—The capability of society to produce and consume or otherwise benefit from goods and services, including contributions to jobs and market and nonmarket benefits (36 CFR 219.19).

**ecoregion**—A geographic area containing distinctive ecological assemblages, topographic and climatic gradients, and historical land uses.

**ecosystem**—A spatially explicit, relatively homogeneous unit of the Earth that includes all interacting organisms and elements of the abiotic environment within its boundaries (36 CFR 219.19).

**ecosystem diversity**—The variety and relative extent of ecosystems (36 CFR 219.19).

**ecosystem integrity**—See “ecological integrity.”

**ecosystem services**—Benefits that people obtain from ecosystems (see also “provisioning services,” “regulating services,” “supporting services,” and “cultural services”).

**endangered species**—Any species or subspecies that the secretary of the interior or the secretary of commerce has deemed in danger of extinction throughout all or a significant portion of its range (16 U.S.C. Section 1532).

**environmental justice populations**—Groups of people who have low incomes or who identify themselves as African American, Asian or Pacific Islander, American Indian or Alaskan Native, or of Hispanic origin.

**environmental justice**—an executive order designation requiring that federal land managers identify any disproportionately high and adverse human health and environmental effects of agency programs, policies, and actions on minority and low-income populations. (Grinspoon et al. 2014). An environmental justice population is a group of people that meets the criteria for low-income or minority status under Executive Order 12898. An environmental justice population may be low income and/or minority.

**environmental suitability**—environmental suitability is the conditions (here predicted by fire season precipitation, maximum temperature, slope and elevation) where large wildfires have manifested in the past and therefore could reasonably be predicted to occur in the future.

**federally recognized American Indian tribe**—An American Indian tribe or Alaska Native Corporation, band, nation, pueblo, village, or community that the secretary of the interior acknowledges to exist as an Indian tribe under the Federally Recognized Indian Tribe List Act of 1994, 25 U.S.C. 479a (36 CFR 219.19).
**fire-diverse ecosystems (mixed severity) (fire ecology group)**—Fire can be important to ecosystem function, but it is not the primary driver of successional dynamics, including structure and composition. Fires were historically moderately frequent, ranging primarily between mixed and high severity in a variety of patch sizes.

**fire exclusion**—Curtailment of wildland fire because of deliberate suppression of ignitions, as well as unintentional effects of human activities such as intensive grazing that removes grasses and other fuels that carry fire (Keane et al. 2002).

**Fire infrequent ecosystems (fire ecology group)**—Fire is not necessarily a part of most ecosystem functions, although when fires do occur, they can be highly impactful. Fires were historically rare or infrequent, of mixed to high severity, in large patches, and were a rare disturbance within these systems.

**fire regime**—A characterization of long-term patterns of fire in a given ecosystem over a specified and relatively long period of time, based on multiple attributes, including frequency, severity, extent, spatial complexity, and seasonality of fire occurrence.

**fire refugia**—Landscape elements that remain unburned or minimally affected by fire, thereby supporting postfire ecosystem function, biodiversity, and resilience to disturbances.

**fire severity**—The magnitude of the effects of fire on ecosystem components, in this document specifically effects of fire on vegetation.

**fire suitability**—The environmental conditions as measured by fire season precipitation, maximum temperature, slope and elevation that, based on past fire occurrence and size, would potentially host a similar fire in the future. In the Bioregional Assessment we discuss suitability for large wildfires.

**fire suppression**—The human act of extinguishing wildfires (Keane et al. 2002).

**forest assessment**—A report available to the public that must be completed for the development of a new plan or for a plan revision. An assessment is the identification and evaluation of existing information to support land management planning. Assessments are not decision-making documents but provide current information on select topics relevant to the plan area, in the context of the broader landscape. (36 CFR 219.19).

**frequent-fire-dependent ecosystems (fire ecology group)**—Fire is essential to overall ecosystem functions. Before Euro-American settlement, fires were quite frequent, of low or mixed severity, and were the primary driver of disturbance. Fire in these systems drives structural and successional dynamics, favoring fire-dependent and fire-adapted species.

**fuels (wildland)**—Combustible material in wildland areas, including live and dead plant biomass such as trees, shrub, grass, leaves, litter, snags, and logs.

**fuels management**—Manipulation of wildland fuels through mechanical, chemical, biological, or manual means, or by fire, in support of land management objectives to control or mitigate the effects of future wildland fire.

**function (ecological)**—Ecological processes, such as energy flow; nutrient cycling and retention; soil development and retention; predation and herbivory; and natural disturbances such as wind, fire, and floods that sustain composition and structure (FSM 2020). See also “key ecological function.”
goals (in land management plans)—Broad statements of intent, other than desired conditions, which do not include expected completion dates (36 CFR part 219.7(e)(2)).

habitat—An area with the environmental conditions and resources that are necessary for occupancy by a species and for individuals of that species to survive and reproduce.

invasive species—An alien species (or subspecies) whose deliberate, accidental, or self-introduction is likely to cause economic or environmental harm or harm to human health (Executive Order 13112).

key watersheds—Watersheds that are expected to serve as refugia for aquatic organisms, particularly in the short term, for at-risk fish populations that have the greatest potential for restoration, or to provide sources of high-quality water.

land management direction—guides and directs management through a combination of aspirations and projections (desired conditions and objectives) and constraints (standards and guidelines). Land management direction also specifies what activities are acceptable or suitable on what parts of a national forest.

land management plan (U.S. Forest Service)—A document or set of documents that provides management direction for an administrative unit of the National Forest System (FSH 1909.12.5).

land use allocation—A process of allocating different activities or uses to specific units of area within a geospatial context, to maximize a spectrum of social, economic, and ecological benefits.

landscape—A defined area irrespective of ownership or other artificial boundaries, such as a spatial mosaic of terrestrial and aquatic ecosystems, landforms, and plant communities, repeated in similar form throughout such a defined area (36 CFR 219.19).

late-successional forest—Forests that have developed after long periods of time (typically at least 100 to 200 years) following major disturbances, and that contain a major component of shade-tolerant tree species that can regenerate beneath a canopy and eventually grow into the canopy in which small canopy gaps occur. Note that FEMAT (1993) and the NWFP also applied this term to older (at least 80 years) forest types, including both old-growth and mature forests, regardless of the shade tolerance of the dominant tree species (for example, 90-year-old forests dominated by Douglas-fir were termed late successional).

late-successional reserve (LSR)—Lands reserved for the protection and restoration of late-successional and old-growth forest ecosystems and habitat for associated species.

managing wildfire for resource objectives—Managing wildfires to promote multiple objectives such as reducing fire danger or restoring forest health and ecological processes rather than attempting full suppression. The terms “managed wildfire” or “resource objective wildfire” have also been used to describe such events (Long et al. 2017). However, fire managers note that many unplanned ignitions are managed using a combination of tactics, including direct suppression, indirect containment, monitoring of fire spread, and even accelerating fire spread, across their perimeters and over their full duration. Therefore, terms that separate “managed” wildfires from fully “suppressed” wildfires do not convey that complexity. (See “Use of wildland fire,” which also includes prescribed burning).
**matrix**—Federal and other lands outside of specifically designated reserve areas, particularly the late-successional reserves under the NWFP, that are managed for timber production and other objectives.

**minority population**—A readily identifiable group of people living in geographic proximity with a population that is at least 50 percent minority; or, an identifiable group that has a meaningfully greater minority population than the adjacent geographic areas, or may also be a geographically dispersed/transient set of individuals such as migrant workers or Americans Indians (CEQ 1997).

**mitigation (climate change)**—Efforts to reduce anthropogenic alteration of climate, in particular by increasing carbon sequestration.

**monitoring**—A systematic process of collecting information to track implementation (implementation monitoring), to evaluate effects of actions or changes in conditions or relationships (effectiveness monitoring), or to test underlying assumptions (validation monitoring) (see 36 CFR 219.19).

**native species**—A species historically or currently present in a particular ecosystem as a result of natural migratory or evolutionary processes and not as a result of an accidental or deliberate introduction or invasion into that ecosystem (see 36 CFR 219.19).

**natural range of variation**—The variation of ecological characteristics and processes over specified scales of time and space that are appropriate for a given management application (FSH 1909.12.5).

**nontimber forest products (also known as “special forest products”)**—Various products from forests that do not include logs from trees but do include bark, berries, boughs, bryophytes, bulbs, burls, Christmas trees, cones, ferns, firewood, forbs, fungi (including mushrooms), grasses, mosses, nuts, pine straw, roots, sedges, seeds, transplants, tree sap, wildflowers, fence material, mine props, posts and poles, shingle and shake bolts, and rails (36 CFR part 223 Subpart G).

**old-growth forest**—A forest distinguished by old trees (older than 200 years) and related structural attributes that often (but not always) include large trees, high biomass of dead wood (for instance, snags, down coarse wood), multiple canopy layers, distinctive species composition and functions, and vertical and horizontal diversity in the tree canopy. In dry, fire-frequent forests, old-growth is characterized by large, old fire-resistant trees and relatively open stands without canopy layering.

**passive management**—A management approach in which natural processes are allowed to occur without human intervention to reach desired outcomes.

**patch**—A relatively small area with similar environmental conditions, such as vegetative structure and composition. Sometimes used interchangeably with vegetation or forest stand.

**prescribed fire**—A wildland fire originating from a planned ignition to meet specific objectives identified in a written and approved prescribed fire plan for which U.S. Environmental Policy Act requirements (where applicable) have been met before ignition (synonymous with controlled burn).

**probable sale quantity (PSQ)**—An estimate of the average amount of timber likely to be awarded for sale for a given area (such as the NWFP area) during a specified period.
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**recreation opportunity**—An opportunity to participate in a specific recreation activity in a particular recreation setting to enjoy desired recreation experiences and other benefits that accrue. Recreation opportunities include nonmotorized, motorized, developed, and dispersed recreation on land, water, and in the air (36 CFR 219.19).

**reference conditions**—Vegetation or forest metrics that represent resilient conditions. For the Bioregional Assessment, either natural range of variation, historic range of variation, or conditions that incorporate future environmental change. Historic range of variation is often based on pre-European settlement conditions.

**refugia**—An area that remains less altered by climatic and environmental change (including disturbances such as wind and fire) affecting surrounding regions and that therefore forms a haven for plants and wildlife.

**reserve**—An area of land designated and managed for a special purpose, often to conserve or protect ecosystems, species, or other natural and cultural resources from particular human activities that are detrimental to achieving the goals of the area.

**resilience**—The ability of an ecosystem and its component parts to absorb, or recover from the effects of disturbances through preservation, restoration, or improvement of its essential structures and functions and redundancy of ecological patterns across the landscape.

**restoration need**—The difference between existing conditions and natural range of variation. In terms of forest structure, this is the area departed from the natural range of variation. It can need treatment (thinning and/or prescribed fire) to change or maintain structure, or in need of succession to develop into older structural conditions. See disturbance and succession restoration need (Haugo et al. 2015, DeMeo et al. 2018).

**riparian areas**—Three-dimensional ecotones (the transition zone between two adjoining communities) of interaction that include terrestrial and aquatic ecosystems that extend down into the groundwater, up above the canopy, outward across the floodplain, up the near slopes that drain to the water, laterally into the terrestrial ecosystem, and along the water course at variable widths (36 CFR 219.19).

**riparian reserves**—Reserves established along streams and rivers to protect riparian ecological functions and processes necessary to create and maintain habitat for aquatic and riparian-dependent organisms over time and ensure connectivity within and between watersheds. The Aquatic Conservation Strategy in the NWFP record of decision included standards and guidelines that delineated riparian reserves.

**risk**—A combination of the probability that a negative outcome will occur and the severity of the subsequent negative consequences (36 CFR 219.19).

**salvage cut**—An intermediate harvest removing trees which are dead or dying because of injurious agents other than competition, to recover economic value that would otherwise be lost (Forest Service Activity Tracking System app. B).

**sanitation cut**—An intermediate harvest removing trees to improve stand health by stopping or reducing the actual or anticipated spread of insects and disease (Forest Service Activity Tracking System app. B).
scale—In ecological terms, the extent and resolution in spatial and temporal terms of a phenomenon or analysis, which differs from the definition in cartography regarding the ratio of map distance to Earth surface distance (Jenerette and Wu 2000).

science synthesis—A narrative review of scientific information from a defined pool of sources that compiles and integrates and interprets findings and describes uncertainty, including the boundaries of what is known and what is not known.

sensitive species—Plant or animal species that receive special conservation attention because of threats to their populations or habitats, but which do not have special status as listed or candidates for listing under the Endangered Species Act.


special forest products—See “nontimber forest products.”

species of conservation concern—A species, other than federally recognized as a threatened, endangered, proposed, or candidate species, which is known to occur in the NWFP area and for which the regional forester has determined that the best available scientific information indicates substantial concern about the species’ capability to persist over the long term in the Plan area (36 CFR 219.9(c)).

stand—A descriptor of a land management unit consisting of a contiguous group of trees sufficiently uniform in age-class distribution, composition, and structure, and growing on a site of sufficiently uniform quality, to be a distinguishable unit.

stand clear-cut—An even-aged regeneration or harvest method that removes all trees in the stand producing a fully exposed microclimate for the development of a new age class in one entry (Forest Service Activity Tracking System app. B).

standard—A mandatory constraint on project and activity decision-making, established to help achieve or maintain the desired condition or conditions, to avoid or mitigate undesirable effects, or to meet applicable legal requirements.

single-tree selection cut—An uneven-aged regeneration method where individual trees of all size classes are removed more or less uniformly throughout the stand creating or maintaining a multiage structure to promote growth of remaining trees and to provide space for regeneration. Multiple entries of this activity ultimately results in an uneven-aged stand of 3 or more age classes (Forest Service Activity Tracking System app. B).

strategic surveys—One type of field survey, specified under the NWFP, designed to fill key information gaps on species distributions and ecologies by which to determine if species should be included under the plan’s survey and manage species list.

stressors—Factors that may directly or indirectly degrade or impair ecosystem composition, structure, or ecological process in a manner that may impair its ecological integrity, such as an invasive species, loss of connectivity, or the disruption of a natural disturbance regime (36 CFR 219.19).
structure (ecosystem)—The organization and physical arrangement of biological elements such as snags and down woody debris, vertical and horizontal distribution of vegetation, stream habitat complexity, landscape pattern, and connectivity (FSM 2020).

succession restoration need—The area departed from the natural range of variation where natural ecological processes are needed to move existing conditions closer to natural range of variation. Succession processes inherently require time and include plant growth, decomposition, and regeneration (Haugo et al. 2015, DeMeo et al. 2018).

Survey and Manage Program—A formal part of the NWFP that established protocols for conducting various types of species surveys, identified old-forest-associated species warranting additional consideration for monitoring and protection (see “survey and manage species”), and instituted an annual species review procedure that evaluated new scientific and monitoring information on species for potentially recommending changes in their conservation status, including potential removal from the survey and manage species list.

survey and manage species—A list of species, compiled under the Survey and Manage Program of the NWFP, that were deemed to warrant particular attention for monitoring and protection beyond the guidelines for establishing late-successional forest reserves.

sustainability—The capability to meet the needs of the present generation without compromising the ability of future generations to meet their needs (36 CFR 219.19).

sustainable recreation—The set of recreation settings and opportunities in the National Forest System that is ecologically, economically, and socially sustainable for present and future generations (36 CFR 219.19).

threatened species—Any species that the secretary of the interior or the secretary of commerce has determined is likely to become an endangered species within the foreseeable future throughout all or a significant portion of its range. Threatened species are listed at 50 CFR sections 17.11, 17.12, and 223.102.

timber harvest—The removal of trees for wood fiber use and other multiple-use purposes (36 CFR 219.19).

timber production—The purposeful growing, tending, harvesting, and regeneration of regulated crops of trees to be cut into logs, bolts, or other round sections for industrial or consumer use (36 CFR 219.19).

uncertainty—Amount or degree of confidence as a result of imperfect or incomplete information.

use of wildland fire—Management of either wildfire or prescribed fire to meet resource objectives specified in land or resource management plans (see “managing wildfire for resource objectives” and “prescribed fire”).

watershed—A region or land area drained by a single stream, river, or drainage network; a drainage basin (36 CFR 219.19).
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**watershed analysis**—An analytical process that characterizes watersheds and identifies potential actions for addressing problems and concerns, along with possible management options. It assembles information necessary to determine the ecological characteristics and behavior of the watershed and to develop options to guide management in the watershed, including adjusting riparian reserve boundaries.

**watershed condition**—The state of a watershed based on physical and biogeochemical characteristics and processes (36 CFR 219.19).

**watershed restoration**—Restoration activities that focus on restoring the key ecological processes required to create and maintain favorable environmental conditions for aquatic and riparian-dependent organisms.

**well-being**—The condition of an individual or group in social, economic, psychological, spiritual, or medical terms.

**wilderness**—Any area of land designated by Congress as part of the National Wilderness Preservation System that was established by the Wilderness Act of 1964 (16 U.S.C. 1131–1136) (36 CFR 219.19).

**wildlife**—Undomesticated animal species, including amphibians, reptiles, birds, mammals, fish, and invertebrates that live wild in an area without being introduced by humans.

**wildfire**—Unplanned ignition of a wildland fire (such as a fire caused by lightning, volcanoes, unauthorized and accidental human-caused fires), and escaped prescribed fires.

**wildland-urban interface (WUI)**—The line, area, or zone where structures and other human development meet or intermingle with undeveloped wildland or vegetation fuels.
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National Forest System Land Management Planning; New plan development or plan revision; 36 CFR part 219.7(e)(2)
National Forest System Land Management Planning; Diversity of plant and animal communities; Species of Conservation Concern; 36 CFR 219.9(c)

National Forest System Land Management Planning; Definitions; 36 CFR 219.19


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References


References


## Fire Ecology Group Classification Crosswalk Tables

The tables below display fire ecology group and subgroup classification crosswalks from vegetation zone and subzone, to biophysical setting (BpS) and fire regime group (FRG).

**Table 15. Fire ecology group classification: frequent-fire dependent (driven by fire subgroup), by vegetation zone and subzone**

<table>
<thead>
<tr>
<th>Vegzone Name</th>
<th>Subzone Name</th>
<th>BpS</th>
<th>BpS Name</th>
<th>FRG</th>
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</tr>
<tr>
<td>White Fir-Grand Fir</td>
<td>Moist White Fir-</td>
<td>0710280</td>
<td>Mediterranean California Mesic Mixed Conifer Forest and Woodland</td>
<td>I</td>
</tr>
<tr>
<td>Hardwoods</td>
<td>Oak Woodlands</td>
<td>0310290</td>
<td>Mediterranean California Mixed Oak Woodland</td>
<td>I</td>
</tr>
<tr>
<td>Hardwoods</td>
<td>Other Hardwoods</td>
<td>0311130</td>
<td>California Coastal Live Oak Woodland and Savanna</td>
<td>I</td>
</tr>
</tbody>
</table>
Table 16. Fire ecology group classification: frequent-fire dependent (regenerated by fire subgroup), by vegetation zone and subzone

<table>
<thead>
<tr>
<th>Vegzone Name</th>
<th>Subzone Name</th>
<th>BpS</th>
<th>BpS Name</th>
<th>FRG</th>
</tr>
</thead>
<tbody>
<tr>
<td>Pinyon-Juniper-Cypress</td>
<td>Cypress Woodlands</td>
<td>0311770</td>
<td>California Coastal Closed-Cone Conifer Forest and Woodland</td>
<td>I</td>
</tr>
<tr>
<td>Shrublands</td>
<td>Montane Shrub</td>
<td>911060</td>
<td>Northern Rocky Mountain Montane-Foothill Deciduous Shrubland</td>
<td>II</td>
</tr>
<tr>
<td>Shrublands</td>
<td>Upland Shrub</td>
<td>611260</td>
<td>Inter-Mountain Basins Montane Sagebrush Steppe</td>
<td>II</td>
</tr>
</tbody>
</table>

Table 17. Fire ecology group classification: frequent-fire dependent (tolerant of fire subgroup), by vegetation zone and subzone

<table>
<thead>
<tr>
<th>Vegzone Name</th>
<th>Subzone Name</th>
<th>BpS</th>
<th>BpS Name</th>
<th>FRG</th>
</tr>
</thead>
<tbody>
<tr>
<td>Giant Sequoia</td>
<td>Giant Sequoia Moist</td>
<td>0710280</td>
<td>Mediterranean California Mesic Mixed Conifer Forest and Woodland</td>
<td>I</td>
</tr>
<tr>
<td>Redwood</td>
<td>Redwood Moist</td>
<td>0410150</td>
<td>California Coastal Redwood Forest</td>
<td>I</td>
</tr>
<tr>
<td>Redwood</td>
<td>Redwood Wetlands</td>
<td>0310150</td>
<td>California Coastal Redwood Forest</td>
<td>I</td>
</tr>
</tbody>
</table>

Table 18. Fire ecology group classification: fire diverse (mixed severity), by vegetation zone and subzone

<table>
<thead>
<tr>
<th>Vegzone Name</th>
<th>Subzone Name</th>
<th>BpS</th>
<th>BpS Name</th>
<th>FRG</th>
</tr>
</thead>
<tbody>
<tr>
<td>Mountain Hemlock</td>
<td>Dry Mountain Hemlock</td>
<td>0710550</td>
<td>Rocky Mountain Subalpine Dry-Mesic Spruce-Fir Forest and Woodland</td>
<td>III</td>
</tr>
<tr>
<td>Pinyon-Juniper-Cypress</td>
<td>Juniper Woodlands</td>
<td>0710170</td>
<td>Columbia Plateau Western Juniper Woodland and Savanna</td>
<td>III</td>
</tr>
<tr>
<td>California Red Fir-Shasta Red Fir</td>
<td>Moist Red Fir</td>
<td>0610322</td>
<td>Mediterranean California Red Fir Forest - Southern Sierra</td>
<td>III</td>
</tr>
<tr>
<td>California Red Fir-Shasta Red Fir</td>
<td>Moist Red Fir</td>
<td>0610322</td>
<td>Red Fir</td>
<td>III</td>
</tr>
<tr>
<td>Silver Fir</td>
<td>Moist Silver Fir</td>
<td>0711740</td>
<td>North Pacific Dry-Mesic Silver Fir-Western Hemlock-Douglas-fir Forest</td>
<td>III</td>
</tr>
<tr>
<td>Western Hemlock</td>
<td>Moist Western Hemlock</td>
<td>0710370</td>
<td>North Pacific Maritime Dry-Mesic Douglas-fir-Western Hemlock Forest</td>
<td>III</td>
</tr>
<tr>
<td>Vegzone Name</td>
<td>Subzone Name</td>
<td>BpS</td>
<td>BpS Name</td>
<td>FRG</td>
</tr>
<tr>
<td>------------------------</td>
<td>-------------------------------</td>
<td>---------</td>
<td>--------------------------------------------------------------------------</td>
<td>-----</td>
</tr>
<tr>
<td>Western Red Cedar</td>
<td>Moist Western Red Cedar</td>
<td>1010471</td>
<td>Northern Rocky Mountain Mesic Montane Mixed Conifer Forest</td>
<td>III</td>
</tr>
<tr>
<td>Ponderosa Pine</td>
<td>Ponderosa Pine-Lodgepole Pine</td>
<td>0710532</td>
<td>Northern Rocky Mountain Ponderosa Pine Woodland and Savanna - Xeric</td>
<td>III</td>
</tr>
<tr>
<td>Hardwoods</td>
<td>Riparian Hardwood Forest</td>
<td>0911590</td>
<td>Rocky Mountain Montane Riparian Systems</td>
<td>III</td>
</tr>
<tr>
<td>Hardwoods</td>
<td>Riparian Hardwood Forest</td>
<td>0511520</td>
<td>California Montane Riparian Systems</td>
<td>III</td>
</tr>
<tr>
<td>Hardwoods</td>
<td>Riparian Shrub</td>
<td>0711582</td>
<td>North Pacific Montane Riparian Woodland and Shrubland - Dry</td>
<td>III</td>
</tr>
<tr>
<td>Hardwoods</td>
<td>Riparian Shrub</td>
<td>0611520</td>
<td>California Montane Riparian Systems</td>
<td>III</td>
</tr>
<tr>
<td>Parklands</td>
<td>Sierra Lodgepole Pine Parklands</td>
<td>0610582</td>
<td>Sierra Nevada Subalpine Lodgepole Pine Forest and Woodland</td>
<td>III</td>
</tr>
<tr>
<td>Parklands</td>
<td>Subalpine Fir Parklands</td>
<td>1010460</td>
<td>Northern Rocky Mountain Subalpine Woodland and Parkland</td>
<td>III</td>
</tr>
<tr>
<td>Parklands</td>
<td>Subalpine Larch Parklands</td>
<td>1010460</td>
<td>Northern Rocky Mountain Subalpine Woodland and Parkland</td>
<td>III</td>
</tr>
<tr>
<td>Douglas-Fir</td>
<td>Wet Douglas-Fir</td>
<td>0110350</td>
<td>North Pacific Dry Douglas-fir Forest and Woodland</td>
<td>III</td>
</tr>
<tr>
<td>Western Hemlock</td>
<td>Wet Western Hemlock</td>
<td>0710370</td>
<td>North Pacific Maritime Dry-Mesic Douglas-fir-Western Hemlock Forest</td>
<td>III</td>
</tr>
<tr>
<td>Western Red Cedar</td>
<td>Wet Western Red Cedar</td>
<td>0110180</td>
<td>East Cascades Mesic Montane Mixed Conifer Forest and Woodland</td>
<td>III</td>
</tr>
<tr>
<td>White Fir - Grand Fir</td>
<td>Wet White Fir - Grand Fir</td>
<td>0910470</td>
<td>Northern Rocky Mountain Mesic Montane Mixed Conifer Forest</td>
<td>III</td>
</tr>
<tr>
<td>Parklands</td>
<td>WhiteBark Pine Parklands</td>
<td>1010460</td>
<td>Northern Rocky Mountain Subalpine Woodland and Parkland</td>
<td>III</td>
</tr>
<tr>
<td>Ponderosa Pine</td>
<td>Xeric Pine</td>
<td>0710532</td>
<td>Northern Rocky Mountain Ponderosa Pine Woodland and Savanna - Xeric</td>
<td>III</td>
</tr>
</tbody>
</table>

Table 19. Fire ecology group classification: fire infrequent (fire prone – regeneration subgroup), by vegetation zone and subzone

<table>
<thead>
<tr>
<th>Vegzone Name</th>
<th>Subzone Name</th>
<th>BpS</th>
<th>BpS Name</th>
<th>FRG</th>
</tr>
</thead>
<tbody>
<tr>
<td>White Fir - Grand Fir</td>
<td>Cold Dry White Fir - Grand Fir</td>
<td>0410580</td>
<td>Sierra Nevada Subalpine Lodgepole Pine Forest and Woodland</td>
<td>III</td>
</tr>
<tr>
<td>Lodgepole Pine</td>
<td>Dry Lodgepole Pine</td>
<td>0711670</td>
<td>Rocky Mountain Poor-Site Lodgepole Pine Forest</td>
<td>IV</td>
</tr>
<tr>
<td>Subalpine Fir - Engelmann Spruce</td>
<td>Dry Subalpine Fir</td>
<td>1010550</td>
<td>Rocky Mountain Subalpine Dry-Mesic Spruce-Fir Forest and Woodland</td>
<td>IV</td>
</tr>
<tr>
<td>Lodgepole Pine</td>
<td>Moist Lodgepole Pine</td>
<td>0110500</td>
<td>Rocky Mountain Lodgepole Pine Forest</td>
<td>IV</td>
</tr>
<tr>
<td>Subalpine Fir - Engelmann Spruce</td>
<td>Moist Subalpine Fir</td>
<td>0710560</td>
<td>Rocky Mountain Subalpine Mesic-Wet Spruce-Fir Forest and Woodland</td>
<td>III</td>
</tr>
</tbody>
</table>
### Table 20. Fire ecology group classification: fire infrequent (limited fire subgroup), by vegetation zone and subzone

<table>
<thead>
<tr>
<th>Vegzone Name</th>
<th>Subzone Name</th>
<th>BpS</th>
<th>BpS Name</th>
<th>FRG</th>
</tr>
</thead>
<tbody>
<tr>
<td>Parklands</td>
<td>Bristlecone-Foxtail Pine</td>
<td>0610200</td>
<td>Inter-Mountain Basins Subalpine Limber-Bristlecone Pine Woodland</td>
<td>III</td>
</tr>
<tr>
<td>Pinyon-Juniper-Cypress</td>
<td>Juniper Steppe</td>
<td>0911150</td>
<td>Inter-Mountain Basins Juniper Savanna</td>
<td>III</td>
</tr>
<tr>
<td>Parklands</td>
<td>Limber Pine</td>
<td>0610200</td>
<td>Inter-Mountain Basins Subalpine Limber-Bristlecone Pine Woodland</td>
<td>III</td>
</tr>
<tr>
<td>Parklands</td>
<td>Mountain Hemlock Parklands</td>
<td>0110380</td>
<td>North Pacific Maritime Mesic Subalpine Parkland</td>
<td>V</td>
</tr>
<tr>
<td>Parklands</td>
<td>Mountain Hemlock Parklands</td>
<td>0610330</td>
<td>Mediterranean California Subalpine Woodland</td>
<td>V</td>
</tr>
<tr>
<td>Mountain Hemlock</td>
<td>Mountain Hemlock Wetlands</td>
<td>0710411</td>
<td>North Pacific Mountain Hemlock Forest - Wet</td>
<td>V</td>
</tr>
<tr>
<td>Pinyon-Juniper-Cypress</td>
<td>Pinyon Woodlands</td>
<td>0610190</td>
<td>Great Basin Pinyon-Juniper Woodland</td>
<td>III</td>
</tr>
<tr>
<td>Hardwoods</td>
<td>Riparian Hardwood Forest</td>
<td>0211560</td>
<td>North Pacific Lowland Riparian Forest and Shrubland</td>
<td>V</td>
</tr>
<tr>
<td>Shrublands</td>
<td>Salt Desert Shrub</td>
<td>911530</td>
<td>Inter-Mountain Basins Greasewood Flat</td>
<td>V</td>
</tr>
<tr>
<td>Grasslands - Meadows</td>
<td>Scabland grass</td>
<td>910650</td>
<td>Columbia Plateau Scabland Shrubland</td>
<td>V</td>
</tr>
<tr>
<td>Shrublands</td>
<td>Scabland Shrub</td>
<td>910650</td>
<td>Columbia Plateau Scabland Shrubland</td>
<td>V</td>
</tr>
<tr>
<td>Silver Fir</td>
<td>Silver Fir Parklands</td>
<td>0710420</td>
<td>North Pacific Mesic Western Hemlock-Silver Fir Forest</td>
<td>V</td>
</tr>
<tr>
<td>Sitka Spruce</td>
<td>Sitka Spruce Wetlands</td>
<td>0210360</td>
<td>North Pacific Hypermaritime Sitka Spruce Forest</td>
<td>V</td>
</tr>
<tr>
<td>Parklands</td>
<td>Subalpine Grassland-Forbland</td>
<td>711710</td>
<td>North Pacific Alpine and Subalpine Dry Grassland</td>
<td>IV</td>
</tr>
<tr>
<td>Parklands</td>
<td>Subalpine Grassland-Forbland</td>
<td>610670</td>
<td>Mediterranean California Alpine Fell-Field</td>
<td>V</td>
</tr>
<tr>
<td>Parklands</td>
<td>Subalpine Grassland-Forbland</td>
<td>911430</td>
<td>Rocky Mountain Alpine Fell-Field</td>
<td>V</td>
</tr>
<tr>
<td>Parklands</td>
<td>Subalpine Shrub</td>
<td>110680</td>
<td>North Pacific Dry and Mesic Alpine Dwarf-Shrubland or Fell-field or Meadow</td>
<td>V</td>
</tr>
<tr>
<td>Parklands</td>
<td>Subalpine Shrub</td>
<td>910700</td>
<td>Rocky Mountain Alpine Dwarf-Shrubland</td>
<td>V</td>
</tr>
<tr>
<td>Parklands</td>
<td>Subalpine Shrub</td>
<td>610710</td>
<td>Sierra Nevada Alpine Dwarf-Shrubland</td>
<td>V</td>
</tr>
</tbody>
</table>
### Table 21. Fire ecology group classification: fire infrequent, by vegetation zone and subzone

<table>
<thead>
<tr>
<th>Vegzone Name</th>
<th>Subzone Name</th>
<th>BpS</th>
<th>BpS Name</th>
<th>FRG</th>
</tr>
</thead>
<tbody>
<tr>
<td>Foothill Pine - Coulter Pine</td>
<td>Coulter Pine-Oak</td>
<td>0411770</td>
<td>California Coastal Closed-Cone Conifer Forest and Woodland</td>
<td>IV</td>
</tr>
<tr>
<td>Lodgepole Pine</td>
<td>Lodgepole Pine Wetlands</td>
<td>0711570</td>
<td>North Pacific Swamp Systems</td>
<td>IV</td>
</tr>
<tr>
<td>Mountain Hemlock</td>
<td>Wet Mountain Hemlock</td>
<td>0710412</td>
<td>North Pacific Mountain Hemlock Forest - Xeric</td>
<td>V</td>
</tr>
<tr>
<td>Hardwoods Riparian Shrub</td>
<td></td>
<td>0911540</td>
<td>Inter-Mountain Basins Montane Riparian Systems</td>
<td>IV</td>
</tr>
<tr>
<td>Western Red Cedar</td>
<td>Western Red Cedar Wetlands</td>
<td>1010472</td>
<td>Northern Rocky Mountain Mesic Montane Mixed Conifer Forest - Cedar Groves</td>
<td>V</td>
</tr>
<tr>
<td>Mountain Hemlock</td>
<td>Wet Mountain Hemlock</td>
<td>0710412</td>
<td>North Pacific Mountain Hemlock Forest - Xeric</td>
<td>V</td>
</tr>
<tr>
<td>Subalpine Fir-Engelmann Spruce</td>
<td>Wet Subalpine Fir</td>
<td>0710412</td>
<td>North Pacific Mountain Hemlock Forest - Xeric</td>
<td>V</td>
</tr>
<tr>
<td>Foothill Pine - Coulter Pine</td>
<td>Coulter Pine-Oak</td>
<td>0411770</td>
<td>California Coastal Closed-Cone Conifer Forest and Woodland</td>
<td>IV</td>
</tr>
<tr>
<td>Lodgepole Pine</td>
<td>Lodgepole Pine Wetlands</td>
<td>0711570</td>
<td>North Pacific Swamp Systems</td>
<td>IV</td>
</tr>
<tr>
<td>Mountain Hemlock</td>
<td>Moist Mountain Hemlock</td>
<td>0710412</td>
<td>North Pacific Mountain Hemlock Forest - Xeric</td>
<td>V</td>
</tr>
<tr>
<td>Hardwoods Riparian Shrub</td>
<td></td>
<td>0911540</td>
<td>Inter-Mountain Basins Montane Riparian Systems</td>
<td>IV</td>
</tr>
<tr>
<td>Western Red Cedar</td>
<td>Western Red Cedar Wetlands</td>
<td>1010472</td>
<td>Northern Rocky Mountain Mesic Montane Mixed Conifer Forest - Cedar Groves</td>
<td>V</td>
</tr>
<tr>
<td>Mountain Hemlock</td>
<td>Wet Mountain Hemlock</td>
<td>0710412</td>
<td>North Pacific Mountain Hemlock Forest - Xeric</td>
<td>V</td>
</tr>
<tr>
<td>Subalpine Fir-Engelmann Spruce</td>
<td>Wet Subalpine Fir</td>
<td>0710412</td>
<td>North Pacific Mountain Hemlock Forest - Xeric</td>
<td>V</td>
</tr>
</tbody>
</table>
Table 22. Fire ecology group classification: other, by vegetation zone and subzone

<table>
<thead>
<tr>
<th>Vegzone Name</th>
<th>Subzone Name</th>
<th>BpS</th>
<th>BpS Name</th>
<th>FRG</th>
</tr>
</thead>
<tbody>
<tr>
<td>Developed</td>
<td>Developed</td>
<td>na</td>
<td>No data</td>
<td>na</td>
</tr>
<tr>
<td>Rock</td>
<td>Rock</td>
<td>na</td>
<td>Barren-Rock/Sand/Clay</td>
<td>na</td>
</tr>
</tbody>
</table>

na = not applicable.

**Ecological Departure by Fire Ecology Group Classification: Northwest Forests**

The tables below display ecological departure by fire ecology group for national forests in the Bioregional Assessment of Northwest Forests (BioA).

Table 23. Deschutes National Forest ecological departure by fire ecology group

<table>
<thead>
<tr>
<th>Fire Ecology Group</th>
<th>Low (acres)</th>
<th>Moderate (acres)</th>
<th>High (acres)</th>
<th>Grand Total (acres)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Frequent-fire dependent</td>
<td>123,772</td>
<td>447,748</td>
<td>16,176</td>
<td>587,695</td>
</tr>
<tr>
<td>Fire diverse (mixed severity)</td>
<td>19,737</td>
<td>783,113</td>
<td>7,983</td>
<td>810,833</td>
</tr>
<tr>
<td>Fire infrequent</td>
<td>21,969</td>
<td>204,354</td>
<td>10,976</td>
<td>237,299</td>
</tr>
<tr>
<td>Deschutes National Forest total</td>
<td>165,478</td>
<td>1,435,215</td>
<td>35,135</td>
<td>1,635,827</td>
</tr>
</tbody>
</table>

Table 24. Fremont-Winema National Forest ecological departure by fire ecology group

<table>
<thead>
<tr>
<th>Fire Ecology Group</th>
<th>Low (acres)</th>
<th>Moderate (acres)</th>
<th>High (acres)</th>
<th>Grand Total (acres)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Frequent-fire dependent</td>
<td>74,700</td>
<td>948,883</td>
<td>7,547</td>
<td>1,031,129</td>
</tr>
<tr>
<td>Fire diverse (mixed severity)</td>
<td>23,841</td>
<td>719,284</td>
<td>4,299</td>
<td>747,425</td>
</tr>
<tr>
<td>Fire infrequent</td>
<td>21,319</td>
<td>221,942</td>
<td>45,988</td>
<td>289,249</td>
</tr>
<tr>
<td>Fremont-Winema National Forest total</td>
<td>119,860</td>
<td>1,890,109</td>
<td>57,833</td>
<td>2,067,802</td>
</tr>
</tbody>
</table>

Table 25. Gifford Pinchot National Forest ecological departure by fire ecology group

<table>
<thead>
<tr>
<th>Fire Ecology Group</th>
<th>Low (acres)</th>
<th>Moderate (acres)</th>
<th>High (acres)</th>
<th>Grand Total (acres)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Frequent-fire dependent</td>
<td>1,708</td>
<td>8,547</td>
<td>7,327</td>
<td>17,582</td>
</tr>
<tr>
<td>Fire diverse (mixed severity)</td>
<td>714,185</td>
<td>91,481</td>
<td>1,472</td>
<td>807,138</td>
</tr>
<tr>
<td>Fire infrequent</td>
<td>80,456</td>
<td>342,224</td>
<td>1,403</td>
<td>424,083</td>
</tr>
<tr>
<td>Gifford Pinchot National Forest total</td>
<td>796,349</td>
<td>442,252</td>
<td>10,202</td>
<td>1,248,803</td>
</tr>
</tbody>
</table>
### Table 26. Klamath National Forest ecological departure by fire ecology group

<table>
<thead>
<tr>
<th>Fire Ecology Group</th>
<th>Low (acres)</th>
<th>Moderate (acres)</th>
<th>High (acres)</th>
<th>Grand Total (acres)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Frequent-fire dependent</td>
<td>120,278</td>
<td>767,376</td>
<td>398,518</td>
<td>1,286,172</td>
</tr>
<tr>
<td>Fire diverse (mixed severity)</td>
<td>77,527</td>
<td>127,974</td>
<td>14,587</td>
<td>220,088</td>
</tr>
<tr>
<td>Fire infrequent</td>
<td>11,886</td>
<td>27,666</td>
<td>2,942</td>
<td>42,494</td>
</tr>
<tr>
<td>Klamath National Forest total</td>
<td>209,691</td>
<td>923,016</td>
<td>416,047</td>
<td>1,548,754</td>
</tr>
</tbody>
</table>

### Table 27. Lassen National Forest ecological departure by fire ecology group

<table>
<thead>
<tr>
<th>Fire Ecology Group</th>
<th>Low (acres)</th>
<th>Moderate (acres)</th>
<th>High (acres)</th>
<th>Grand Total (acres)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Frequent-fire dependent</td>
<td>201,525</td>
<td>856,449</td>
<td>174,373</td>
<td>1,232,348</td>
</tr>
<tr>
<td>Fire diverse (mixed severity)</td>
<td>2,326</td>
<td>29,916</td>
<td>23,957</td>
<td>56,199</td>
</tr>
<tr>
<td>Fire infrequent</td>
<td>2,652</td>
<td>31,346</td>
<td>15,986</td>
<td>49,984</td>
</tr>
<tr>
<td>Lassen National Forest total</td>
<td>206,504</td>
<td>917,711</td>
<td>214,317</td>
<td>1,338,531</td>
</tr>
</tbody>
</table>

### Table 28. Mendocino National Forest ecological departure by fire ecology group

<table>
<thead>
<tr>
<th>Fire Ecology Group</th>
<th>Low (acres)</th>
<th>Moderate (acres)</th>
<th>High (acres)</th>
<th>Grand Total (acres)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Frequent-fire dependent</td>
<td>79,532</td>
<td>362,475</td>
<td>375,850</td>
<td>817,856</td>
</tr>
<tr>
<td>Fire diverse (mixed severity)</td>
<td>9,977</td>
<td>35,266</td>
<td>6,584</td>
<td>51,827</td>
</tr>
<tr>
<td>Fire infrequent</td>
<td>696</td>
<td>1,445</td>
<td>295</td>
<td>2,437</td>
</tr>
<tr>
<td>Mendocino National Forest total</td>
<td>90,205</td>
<td>399,186</td>
<td>382,729</td>
<td>872,120</td>
</tr>
</tbody>
</table>

### Table 29. Modoc National Forest ecological departure by fire ecology group

<table>
<thead>
<tr>
<th>Fire Ecology Group</th>
<th>Low (acres)</th>
<th>Moderate (acres)</th>
<th>High (acres)</th>
<th>Grand Total (acres)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Frequent-fire dependent</td>
<td>89,702</td>
<td>672,956</td>
<td>23,568</td>
<td>786,226</td>
</tr>
<tr>
<td>Fire diverse (mixed severity)</td>
<td>37,949</td>
<td>241,226</td>
<td>989</td>
<td>280,165</td>
</tr>
<tr>
<td>Fire infrequent</td>
<td>3,831</td>
<td>46,461</td>
<td>3,077</td>
<td>53,369</td>
</tr>
<tr>
<td>Modoc National Forest total</td>
<td>131,482</td>
<td>960,644</td>
<td>27,634</td>
<td>1,119,760</td>
</tr>
</tbody>
</table>
### Appendix

#### Table 30. Mt. Baker-Snoqualmie National Forest ecological departure by fire ecology group

<table>
<thead>
<tr>
<th>Fire Ecology Group</th>
<th>Low (acres)</th>
<th>Moderate (acres)</th>
<th>High (acres)</th>
<th>Grand Total (acres)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Frequent-fire dependent</td>
<td>300</td>
<td>625</td>
<td>342</td>
<td>1,267</td>
</tr>
<tr>
<td>Fire diverse (mixed severity)</td>
<td>281,508</td>
<td>92,561</td>
<td>1,295</td>
<td>375,364</td>
</tr>
<tr>
<td>Fire infrequent</td>
<td>518,356</td>
<td>749,189</td>
<td>7,454</td>
<td>1,274,998</td>
</tr>
<tr>
<td>Mt. Baker-Snoqualmie National Forest total</td>
<td>800,163</td>
<td>842,375</td>
<td>9,091</td>
<td>1,651,629</td>
</tr>
</tbody>
</table>

#### Table 31. Mt. Hood National Forest ecological departure by fire ecology group

<table>
<thead>
<tr>
<th>Fire Ecology Group</th>
<th>Low (acres)</th>
<th>Moderate (acres)</th>
<th>High (acres)</th>
<th>Grand Total (acres)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Frequent-fire dependent</td>
<td>10,620</td>
<td>83,537</td>
<td>9,576</td>
<td>103,733</td>
</tr>
<tr>
<td>Fire diverse (mixed severity)</td>
<td>566,059</td>
<td>18,656</td>
<td>638</td>
<td>585,353</td>
</tr>
<tr>
<td>Fire infrequent</td>
<td>136,887</td>
<td>142,320</td>
<td>1,659</td>
<td>280,867</td>
</tr>
<tr>
<td>Mt. Hood National Forest total</td>
<td>713,566</td>
<td>244,513</td>
<td>11,874</td>
<td>969,953</td>
</tr>
</tbody>
</table>

#### Table 32. Ochoco National Forest ecological departure by fire ecology group

<table>
<thead>
<tr>
<th>Fire Ecology Group</th>
<th>Low (acres)</th>
<th>Moderate (acres)</th>
<th>High (acres)</th>
<th>Grand Total (acres)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Frequent-fire dependent</td>
<td>52,408</td>
<td>424,183</td>
<td>3,162</td>
<td>479,753</td>
</tr>
<tr>
<td>Fire diverse (mixed severity)</td>
<td>52,632</td>
<td>167,225</td>
<td>21,549</td>
<td>241,406</td>
</tr>
<tr>
<td>Fire infrequent</td>
<td>4,259</td>
<td>26,095</td>
<td>170</td>
<td>30,525</td>
</tr>
<tr>
<td>Ochoco National Forest total</td>
<td>109,299</td>
<td>617,503</td>
<td>24,881</td>
<td>751,683</td>
</tr>
</tbody>
</table>

#### Table 33. Okanogan-Wenatchee National Forest ecological departure by fire ecology group

<table>
<thead>
<tr>
<th>Fire Ecology Group</th>
<th>Low (acres)</th>
<th>Moderate (acres)</th>
<th>High (acres)</th>
<th>Grand Total (acres)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Frequent-fire dependent</td>
<td>74,156</td>
<td>523,571</td>
<td>514,563</td>
<td>1,112,291</td>
</tr>
<tr>
<td>Fire diverse (mixed severity)</td>
<td>553,289</td>
<td>193,720</td>
<td>46,796</td>
<td>793,804</td>
</tr>
<tr>
<td>Fire infrequent</td>
<td>444,098</td>
<td>462,108</td>
<td>25,864</td>
<td>932,070</td>
</tr>
<tr>
<td>Fire Ecology Group</td>
<td>Low (acres)</td>
<td>Moderate (acres)</td>
<td>High (acres)</td>
<td>Grand Total (acres)</td>
</tr>
<tr>
<td>-----------------------------------------</td>
<td>-------------</td>
<td>-----------------</td>
<td>--------------</td>
<td>---------------------</td>
</tr>
<tr>
<td>Okanogan-Wenatchee National Forest total</td>
<td>1,071,542</td>
<td>1,179,399</td>
<td>587,224</td>
<td>2,838,165</td>
</tr>
</tbody>
</table>

**Table 34. Olympic National Forest ecological departure by fire ecology group**

<table>
<thead>
<tr>
<th>Fire Ecology Group</th>
<th>Low (acres)</th>
<th>Moderate (acres)</th>
<th>High (acres)</th>
<th>Grand Total (acres)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Frequent-fire dependent</td>
<td>94</td>
<td>70</td>
<td>181</td>
<td>345</td>
</tr>
<tr>
<td>Fire diverse (mixed severity)</td>
<td>227,968</td>
<td>76,735</td>
<td>291</td>
<td>304,995</td>
</tr>
<tr>
<td>Fire infrequent</td>
<td>161,066</td>
<td>190,448</td>
<td>857</td>
<td>352,371</td>
</tr>
<tr>
<td>Olympic National Forest total</td>
<td>389,128</td>
<td>267,254</td>
<td>1,329</td>
<td>657,711</td>
</tr>
</tbody>
</table>

**Table 35. Rogue River-Siskiyou National Forest ecological departure by fire ecology group**

<table>
<thead>
<tr>
<th>Fire Ecology Group</th>
<th>Low (acres)</th>
<th>Moderate (acres)</th>
<th>High (acres)</th>
<th>Grand Total (acres)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Frequent-fire dependent</td>
<td>89,616</td>
<td>815,509</td>
<td>265,231</td>
<td>1,170,356</td>
</tr>
<tr>
<td>Fire diverse (mixed severity)</td>
<td>310,995</td>
<td>235,684</td>
<td>14,337</td>
<td>561,015</td>
</tr>
<tr>
<td>Fire infrequent</td>
<td>6,702</td>
<td>72,049</td>
<td>7,550</td>
<td>86,301</td>
</tr>
<tr>
<td>Rogue River-Siskiyou National Forest total</td>
<td>407,314</td>
<td>1,123,241</td>
<td>287,118</td>
<td>1,817,673</td>
</tr>
</tbody>
</table>

**Table 36. Shasta-Trinity National Forest ecological departure by fire ecology group**

<table>
<thead>
<tr>
<th>Fire Ecology Group</th>
<th>Low (acres)</th>
<th>Moderate (acres)</th>
<th>High (acres)</th>
<th>Grand Total (acres)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Frequent-fire dependent</td>
<td>159,759</td>
<td>1,150,427</td>
<td>863,871</td>
<td>2,174,056</td>
</tr>
<tr>
<td>Fire diverse (mixed severity)</td>
<td>60,582</td>
<td>138,568</td>
<td>21,844</td>
<td>220,994</td>
</tr>
<tr>
<td>Fire infrequent</td>
<td>7,041</td>
<td>47,173</td>
<td>5,405</td>
<td>59,619</td>
</tr>
<tr>
<td>Shasta-Trinity National Forest total</td>
<td>227,382</td>
<td>1,336,168</td>
<td>891,120</td>
<td>2,454,669</td>
</tr>
</tbody>
</table>

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### Table 37. Siuslaw National Forest ecological departure by fire ecology group

<table>
<thead>
<tr>
<th>Fire Ecology Group</th>
<th>Low (acres)</th>
<th>Moderate (acres)</th>
<th>High (acres)</th>
<th>Grand Total (acres)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Frequent-fire dependent</td>
<td>116</td>
<td>215</td>
<td>198</td>
<td>529</td>
</tr>
<tr>
<td>Fire diverse (mixed severity)</td>
<td>103,995</td>
<td>5,763</td>
<td>22</td>
<td>109,779</td>
</tr>
<tr>
<td>Fire infrequent</td>
<td>317,371</td>
<td>177,198</td>
<td>95</td>
<td>494,664</td>
</tr>
<tr>
<td>Siuslaw National Forest total</td>
<td>421,482</td>
<td>183,175</td>
<td>315</td>
<td>604,972</td>
</tr>
</tbody>
</table>

### Table 38. Six Rivers National Forest ecological departure by fire ecology group

<table>
<thead>
<tr>
<th>Fire Ecology Group</th>
<th>Low (acres)</th>
<th>Moderate (acres)</th>
<th>High (acres)</th>
<th>Grand Total (acres)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Frequent-fire dependent</td>
<td>19,298</td>
<td>716,665</td>
<td>307,520</td>
<td>1,043,483</td>
</tr>
<tr>
<td>Fire diverse (mixed severity)</td>
<td>62,108</td>
<td>85,758</td>
<td>8,717</td>
<td>156,583</td>
</tr>
<tr>
<td>Fire infrequent</td>
<td>7,939</td>
<td>4,809</td>
<td>661</td>
<td>13,410</td>
</tr>
<tr>
<td>Six Rivers National Forest total</td>
<td>89,345</td>
<td>807,232</td>
<td>316,898</td>
<td>1,213,476</td>
</tr>
</tbody>
</table>

### Table 39. Umpqua National Forest ecological departure by fire ecology group

<table>
<thead>
<tr>
<th>Fire Ecology Group</th>
<th>Low (acres)</th>
<th>Moderate (acres)</th>
<th>High (acres)</th>
<th>Grand Total (acres)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Frequent-fire dependent</td>
<td>14,062</td>
<td>79,472</td>
<td>13,795</td>
<td>107,329</td>
</tr>
<tr>
<td>Fire diverse (mixed severity)</td>
<td>628,475</td>
<td>140,956</td>
<td>4,851</td>
<td>774,281</td>
</tr>
<tr>
<td>Fire infrequent</td>
<td>6,137</td>
<td>63,580</td>
<td>10,698</td>
<td>80,415</td>
</tr>
<tr>
<td>Umpqua National Forest total</td>
<td>648,674</td>
<td>284,007</td>
<td>29,344</td>
<td>962,025</td>
</tr>
</tbody>
</table>

### Table 40. Willamette National Forest ecological departure by fire ecology group

<table>
<thead>
<tr>
<th>Fire Ecology Group</th>
<th>Low (acres)</th>
<th>Moderate (acres)</th>
<th>High (acres)</th>
<th>Grand Total (acres)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Frequent-fire dependent</td>
<td>6,699</td>
<td>9,800</td>
<td>4,127</td>
<td>20,626</td>
</tr>
<tr>
<td>Fire diverse (mixed severity)</td>
<td>1,101,932</td>
<td>111,218</td>
<td>576</td>
<td>1,213,726</td>
</tr>
<tr>
<td>Fire infrequent</td>
<td>91,675</td>
<td>279,190</td>
<td>314</td>
<td>371,180</td>
</tr>
<tr>
<td>Willamette National Forest total</td>
<td>1,200,306</td>
<td>400,209</td>
<td>5,017</td>
<td>1,605,532</td>
</tr>
</tbody>
</table>