

# **Supplement 7 – Biological Assessment and Determination for Federally Listed Species for the Non-MA 3.63 Decision**



---

# **Biological Assessment<sup>1</sup>**

*FOR*

*BLACK-TAILED PRAIRIE DOG MANAGEMENT IN NON-MANAGEMENT AREAS 3.63*

---

on units of the

*Nebraska National Forest:*

*Buffalo Gap and Fort Pierre National Grasslands, South Dakota*

*Oglala National Grassland, Nebraska*

---

**Prepared and reviewed by:**

**John Sidle, Fish and Wildlife Biologist**

**Biological Determinations Made By:**

/s/ John Sidle

**John Sidle, Fish and Wildlife Biologist**

June 23, 2008

**Date:**

<sup>1</sup> Meets the standards for Biological Assessment (50 CFR 402.12(f))

# Table of Contents

1. Introduction .....	1
2. Description of the Proposed Action.....	1
Purpose of and Need for Action.....	1
3. Area Affected .....	4
Existing Condition .....	7
4. Proposed Action (Alt. 1) and Alternatives Analyzed in the FEIS.....	10
The Alternatives.....	10
5. Endangered Species Considered in the Analysis.....	12
6. The Black-footed Ferret’s ( <i>Mustela nigripes</i> ) Status, Ecology, and Prairie Dog Habitat.....	16
Distribution and Status.....	16
Prairie Dog Habitat of the Black-footed Ferret.....	17
Prairie Dog as a Keystone Species in Creating Habitat for the Black-footed Ferret .....	21
Black-footed Ferret Use of Prairie Dog Colonies.....	23
7. Black-footed Ferret ESA Status and Other Organizational Rankings.....	26
8. Recovery and Conservation Planning.....	27
9. Analysis of Effects: Black-footed Ferret.....	32
Direct Effects .....	32
Indirect Effects.....	39
Cumulative Effects.....	39
10. Determination of Effect and Rationale for the Black-footed Ferret .....	49
The Determination .....	49
Rationale .....	49
11. Whooping Crane ( <i>Grus americana</i> ) Distribution and Status .....	50
Migration .....	50
Migration Habitat.....	51
12. Whooping Crane ESA Status and Other Organizational Rankings.....	55
Existing Conditions.....	55
Direct, Indirect, and Cumulative Effects .....	55
13. Determination of Effect and Rationale for the Whooping Crane.....	55
Entire Project Area.....	55
Rationale .....	55
14. References for the Biological Assessment and Biological Effects Report.....	57

## 1. Introduction

The purpose of this Biological Assessment (BA) is to determine the likely effects of black-tailed prairie dog (prairie dog) (*Cynomys ludovicianus*) management in the interior management zone (IMZ) on Nebraska National Forest (NNF) units (Buffalo Gap, Fort Pierre, and Oglala national grasslands) for federally listed species and proposed species under the Endangered Species Act (ESA). For South Dakota national grasslands, the BMZ has been defined on the Buffalo Gap National Grassland as less than one-half mile from private land and less than a quarter mile on the Fort Pierre National Grassland. On the Oglala National Grassland in Nebraska, the BMZ is less than one-half mile from private land.

Section 7 of the ESA requires federal agencies to use their authorities to carry out programs to conserve endangered and threatened species, and to insure that actions authorized, funded, or carried out by them are not likely to jeopardize the continued existence of listed or proposed species, or result in the destruction or adverse modification of their designated critical habitats. A BA must be prepared for federal actions that are “major construction activities” (defined under the National Environmental Policy Act (NEPA) as a project significantly affecting the quality of the human environment) to evaluate the potential effects of the proposal on listed or proposed species. This BA is being conducted in conjunction with an environmental impact statement (EIS) analyzing the situation. The contents of the BA are at the discretion of the federal agency, and will depend upon the nature of the federal action (50 CFR 402.12(f)).

## 2. Description of the Proposed Action

### Purpose of and Need for Action

The 2005 *Record of Decision for Black-tailed Prairie Dog Conservation and Management on the Nebraska National Forest* ((USDA Forest Service 2005e) focused on the encroachment of prairie dog colonies from National Grasslands onto adjoining private or tribal agricultural lands, where ranchers and farmers are concerned about losses in agricultural production, costs of managing prairie dogs, effects on land values, and risks to health and safety. That effort dealt with the prairie dog colonies near the Forest boundaries in the BMZ and the impacts as colonies expanded onto non-Forest Service lands. The Forest Plan did not set acre objectives for prairie dog colonies outside the BMZs, and it limited rodenticide use to very specific situations.

The purpose of this proposed action is to determine the techniques and objectives for managing prairie dog colonies in the interior of the Oglala, Buffalo Gap, and Fort Pierre national grasslands (National Grasslands) in an adaptive fashion through the following:

- ◆ Setting objectives for desired acres of prairie dog colonies within the interior of the National Grasslands to move toward desired prairie dog acres, and to maintain or move toward desired vegetation cover, protect top soil, and prevent the establishment of noxious and invasive species.
- ◆ Managing black-tailed prairie dog habitat designated as a black-footed ferret management area (MA 3.63) in the 2002 Forest Plan to sustain populations of black-footed ferrets and associated species.

However, this proposed action addresses only non-MA 3.63 areas. This proposed action selects Alternative 1, for non-MA 3.63 areas only, in the FEIS. Current direction would continue in the two MA 3.63 areas, Conata Basin and Smithwick.

This BA tiers to and incorporates still valid information and analysis from the 2005 BMZ decision. The BA and other information are in the administrative record for that decision at the NNF Supervisor’s Office in Chadron, Nebraska. For South Dakota national grasslands, the BMZ has been defined on the Buffalo Gap National Grassland as less than one-half mile from private land and less than a quarter mile

on the Fort Pierre National Grassland. On the Oglala National Grassland in Nebraska, the BMZ is less than one-half mile from private land.

The following aspects of the proposed action are relevant to this BA:

### ***Prairie Dog Management in Non-MA 3.63 Areas***

Prairie dog acreage objectives would be established for non-MA 3.63 areas in all geographic areas (GA) (Table 1). For all but two GAs, a minimum and maximum range of acres would be set. In addition, desired vegetation conditions for prairie dog colonies would be described, along with a suite of tools that could be used to achieve those desired conditions.

**Minimum Range of Acres:** The minimum number of acres is 1,000, based on the 2001 Forest Plan definition of a prairie dog colony complex. The 2001 Forest Plan defines a prairie dog colony complex as a group of at least 10 prairie dog colonies with nearest-neighbor inter-colony distances not exceeding 6 miles and with a total colony complex acreage of at least 1,000 acres. Two of the GAs, Fall River West and Wall North GA, currently are below the minimum prairie dog acreage requirement while the others are at or above 1,000 acres.

**Maximum Range of Acres:** The maximum range of prairie dog colony would be 3 percent of the gross national grassland acres in each GA. When the maximum range of prairie dog acres has been exceeded, rodenticide would be used to reduce acres. Specific colony locations may change based on other management objectives; for example, high vegetative structure for prairie grouse. During the range allotment management planning (RAMP) process, the District Ranger will use adaptive management to decide where the prairie dog colonies will occur, using existing colony locations as a starting point.

For Fall River Southeast and Wall Southwest GAs which contain MA 3.63 areas, prairie dog colony acreage objectives will be set at a later date through a separate proposed action.

Table 1. Range of prairie dog colony acres by geographic area.

Geographic Area	Current (2006) occupied acres in IMZ	Acres of prairie dog colonies	
		Minimum	Maximum
Oglala	1,125	1,000 <sup>a</sup>	2,800
Fall River Northeast	1,130	1,000	2,800
Fall River West	210	1,000	3,600
Fall River Southeast (excludes MA 3.63)	42	No acreage objective	
Wall North	454	1,000	2,100
Wall Southeast	1,414	1,000	2,900
Wall Southwest (excludes MA 3.63)	214	No acreage objective	
Fort Pierre	1,735	1,000 <sup>a</sup>	3,500
<sup>a</sup> No change from current Forest Plan direction			

**Desired Vegetation Condition:** Plant communities identified in ROD Supplement 1, Table 6 would describe desired vegetation condition on prairie dog colonies. The desired condition would primarily

be a blue grama plant community. Similarity Index<sup>2</sup> (SI) would be used to monitor the current condition of plant communities. The threshold for rodenticide use would be an SI of 25% of the Historic Climax Plant Community (HCPC) which generally equates to a blue grama community as described in Table 6, ROD Supplement 1.

- ◆ The District Ranger would use non-lethal methods before employing lethal methods to maintain the range of prairie dog acres and achieve desired vegetation conditions (see Supplement 1 – Implementation Plan).
- ◆ Prairie dogs would have priority over livestock for forage in allotments with prairie dog colonies. This may include annual removal of livestock from allotments with prairie dog colonies as a form of non lethal treatment.
- ◆ Livestock grazing would be managed to maintain prairie dog habitat to meet desired vegetation conditions and minimize the potential for soil loss. This management may include annual modifications to livestock grazing and other tools as described in ROD Supplement 1, Table 1. Long-term modifications to livestock grazing would be addressed in the range allotment management planning (RAMP) process.
- ◆ This proposed action would allow the use of rodenticide in the interior-colony management zone outside the MA 3.63 areas.
- ◆ Adaptive management as described in ROD Supplement 1 – Implementation Plan would be used. This includes the suite of management tools listed in Table 1 in that supplement. The District Ranger would be the decision-maker for the site-specific, on-the-ground actions. Monitoring would be used to determine if the thresholds (discussed below) that trigger action have been met. Cumulative effects of the actions would be considered and the actions would be modified, if needed. An integrated plan using non-lethal and lethal treatment methods will be developed. The integrated plan would include items to be monitored during the action. The plan would be specific to a colony. When the plan is implemented, we would monitor the results and use the findings to adjust future management actions.

There are two thresholds that could initiate use of rodenticide: when the maximum range of prairie dog colony acres is exceeded and when necessary to achieve a desired condition for vegetation on prairie dog colonies.

**1. Maximum acreage limit exceeded (see ROD Supplement 1 – Implementation Plan):**

- ◆ When prairie dog colony acres in a GA exceed the maximum, rodenticide use may occur on up to 1/3 of the maximum range of acres (i.e., if the aggregate acres exceed 3%, reduce to about 2%).
- ◆ Poisoning would usually occur for 1 to 3 years, until the District Ranger determines the prairie dog colony acres are at or below the maximum acreage. Colony reduction, due to exceeding acreage limits, may occur every year based on available funding. However, the District Ranger should avoid rodenticide use for more than three to five consecutive years. The intent is not to apply rodenticide annually.

<sup>2</sup> Similarity Index is a method to evaluate an ecological site. This method compares the present plant community on an ecological site to the desired vegetation that can exist on the site. The SI is expressed as the percentage of a vegetation plant community presently on the site to the desired vegetation plant community. The desired vegetation plant community must be identified as the reference plant community. (NRCS 2006).

**1. Maximum acreage limit exceeded cont.**

- ◆ Acres that have been poisoned would not be used to calculate prairie dog acreage requirements until monitoring shows these areas have been recolonized.
- ◆ Before, during and following poisoning to reduce prairie dog colony acres, livestock would be removed for a period of 1 to 3 years or until the District Ranger determines that desired prairie dog acreage requirements have been achieved.

**2. To achieve desired condition for vegetation:**

- ◆ Before rodenticide use can occur, the minimum range of prairie dog colony acres for the GA must be achieved. Non-lethal methods can be used at any time (see Supplement 1).
- ◆ Rodenticide would be used to reduce prairie dog densities within a colony. By reducing densities, we would reduce the amount of grass the prairie dogs are utilizing. This allows the vegetation to grow, making it possible to achieve desired vegetation condition. By poisoning some, but not all, of the prairie dogs (reducing density), we would maintain enough prairie dogs to ensure the colony still exists and can meet the needs of species associated with this habitat. In a colony in which prairie dog densities have been reduced but not eliminated, the treated acres would be used to calculate the minimum and maximum prairie dog colony acreage requirements for the GA.
- ◆ If monitoring indicates that the existing condition of the plant community is below the desired condition and the SI is at or below 25% of HCPC, poisoning could occur as determined by the District Ranger (see Supplement 1).
- ◆ Before, during and following poisoning to reduce acres, livestock will be removed for a period of 1 to 3 years or until the District Ranger determines that desired vegetation conditions are being met.

**3. Area Affected**

The proposed action only addresses non-MA 3.63 GAs. GAs are administrative units on the National Grasslands. Their administrative boundary encompasses federal land (national grasslands) and non-federal lands (mostly private). See Figures 1 and 2 for examples of GAs and depictions of IMZ, BMZ, and extent of prairie dog colonies; see maps of other GAs in Appendix F.

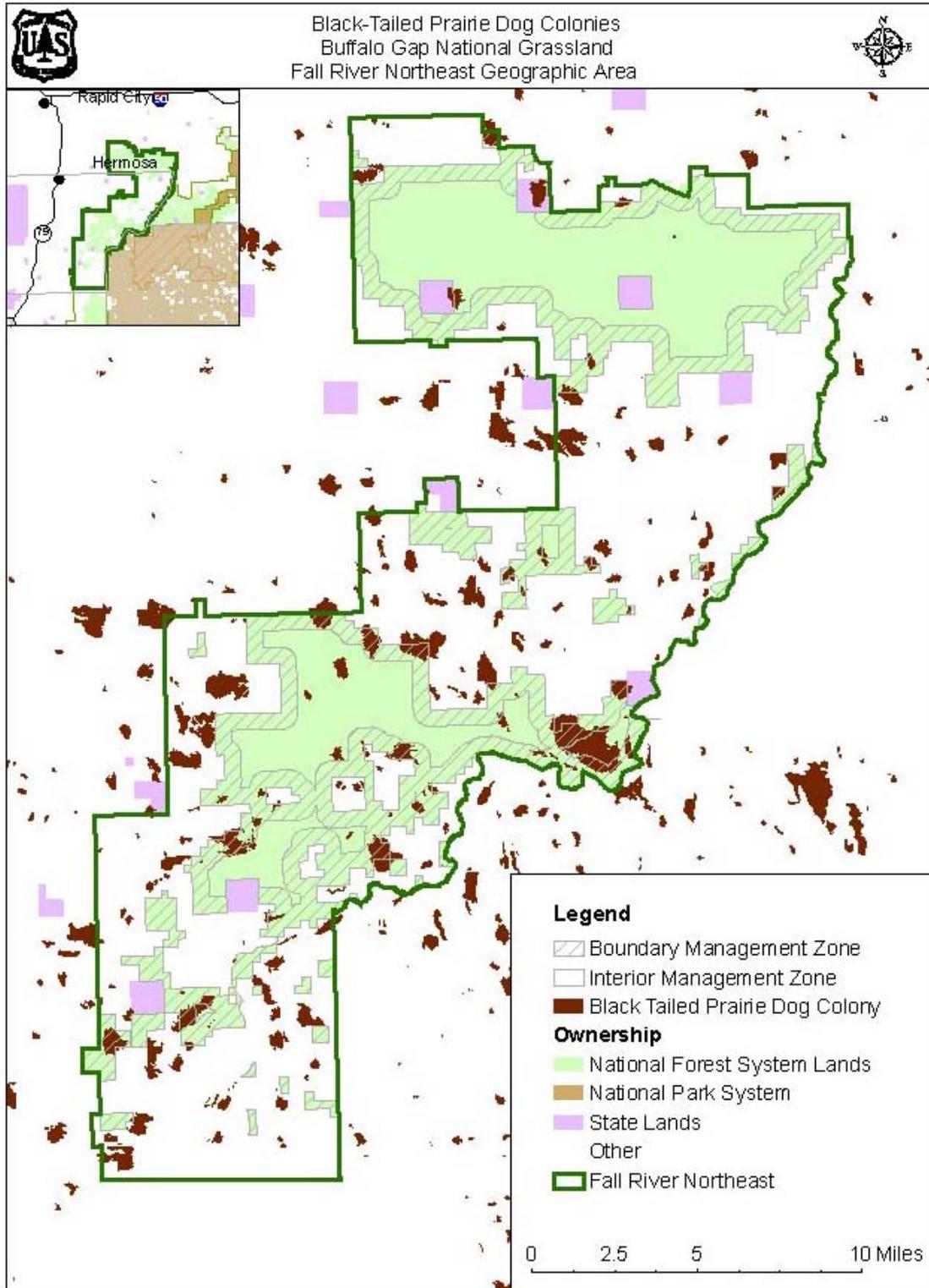


Figure 1. An example of a Geographic Area (GA) showing the extent of black-tailed prairie dog colonies on and in the vicinity of Buffalo Gap National Grassland (Fall River Northeast GA), South Dakota.

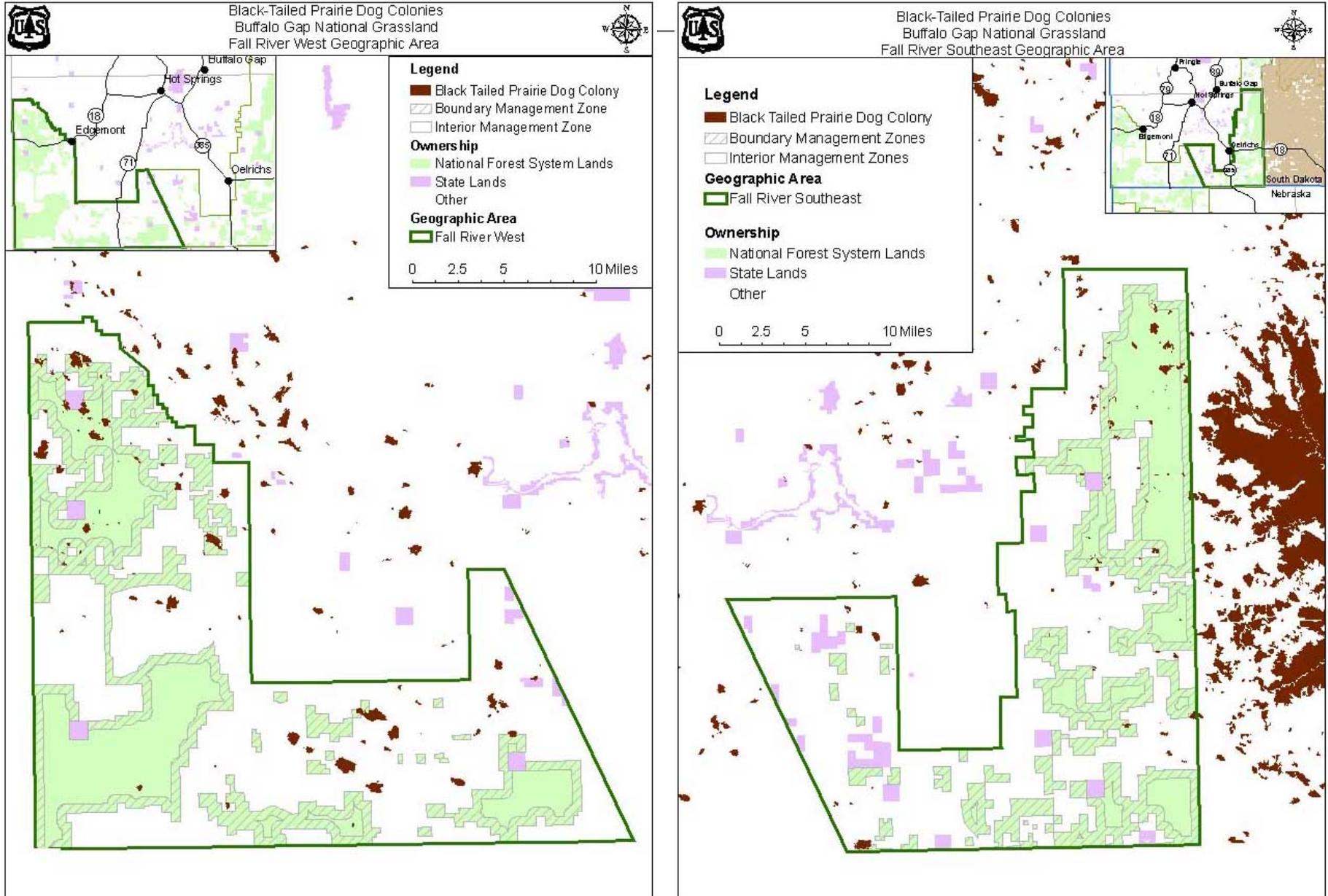


Figure 2. The extent of black-tailed prairie dog colonies on and in the vicinity of Fall River Southwest and Fall River Southeast GAs, Buffalo Gap National Grassland, South Dakota. The large prairie dog colonies on the Pine Ridge Reservation (far right) have been affected by plague but vast colony acreages remain suitable for black-footed ferret reintroduction.

**Buffalo Gap National Grassland and Geographic Areas.** The Buffalo Gap National Grassland (Buffalo Gap) is located in southwestern South Dakota and includes more than 589,000 acres of land that borders private, state, Indian reservation, and national park lands. The eastern half of this unit extends from near Kadoka to the Cheyenne River on the west, north to U.S. Highway 14, and south to the Pine Ridge Indian Reservation. The Wall Ranger District (WRD) at Wall, administers the eastern half. The WRD is divided into 3 GAs (Wall North, Wall Southeast, and Wall Southwest). Wall Southeast contains a 3,910-acre area designated as a Management Area 3.63 for black-footed ferret reintroduction. Wall Southwest contains Conata Basin, a 73,243-acre Management Area 3.63 for black-footed ferret reintroduction. Black-footed ferrets have been successfully reintroduced into this area. For clarity, the 3,910 and 73,243 acres are combined as the 77,155-acres Conata Basin Management Area 3.63. However, the focus of this BA is on the non-MA 3.63 GAs and the impacts that any further prairie dog control could have on the black-footed ferret.

The western half of Buffalo Gap extends from the Cheyenne River on the east to the Wyoming and Nebraska borders on the west and south, respectively, and is administered by the Fall River Ranger District (FRRD), Hot Springs. The FRRD is divided into 3 GAs (Fall River West, Fall River Southeast, and Fall River Northeast). Fall River Southeast contains Smithwick, a 25,307-acre MA 3.63 for black-footed ferret reintroduction. No black-footed ferrets have been reintroduced.

**Fort Pierre National Grassland and Geographic Area.** The Fort Pierre National Grassland (Fort Pierre) (116,053 acres) lies south of Pierre, South Dakota, north of Interstate 90, and west of the Lower Brule Indian Reservation. The Fort Pierre consists of mixed-grass vegetation on a rolling hill landscape just west of the Missouri River.

**Oglala National Grassland and Geographic Area.** The 94,484-acre Oglala National Grassland lies in Dawes and Sioux counties of northwestern Nebraska and contains mostly mixed-grass vegetation. Topography consists of rolling hills and badlands. The grassland is administered by the Pine Ridge Ranger District, Chadron, Nebraska.

## Existing Condition

Prairie dog colonies occur throughout the GAs but are most numerous in the Wall Southwest GA, especially Conata Basin (Tables 1 and 2; Figure 3), an area dedicated to large complexes of prairie dog habitat for black-footed ferret recovery. There are currently only modest acreages of prairie dog colonies in the other GAs. Indeed, the number, density, and distribution of colonies in non-black-footed ferret areas on national grasslands in southwest South Dakota is similar to that found on private land in the same area. Maps of prairie dog colonies, potential prairie dog habitat, and buffer zones are in Appendix F. Please review these maps to understand the extent of prairie dog colonies and suitable prairie dog habitat.

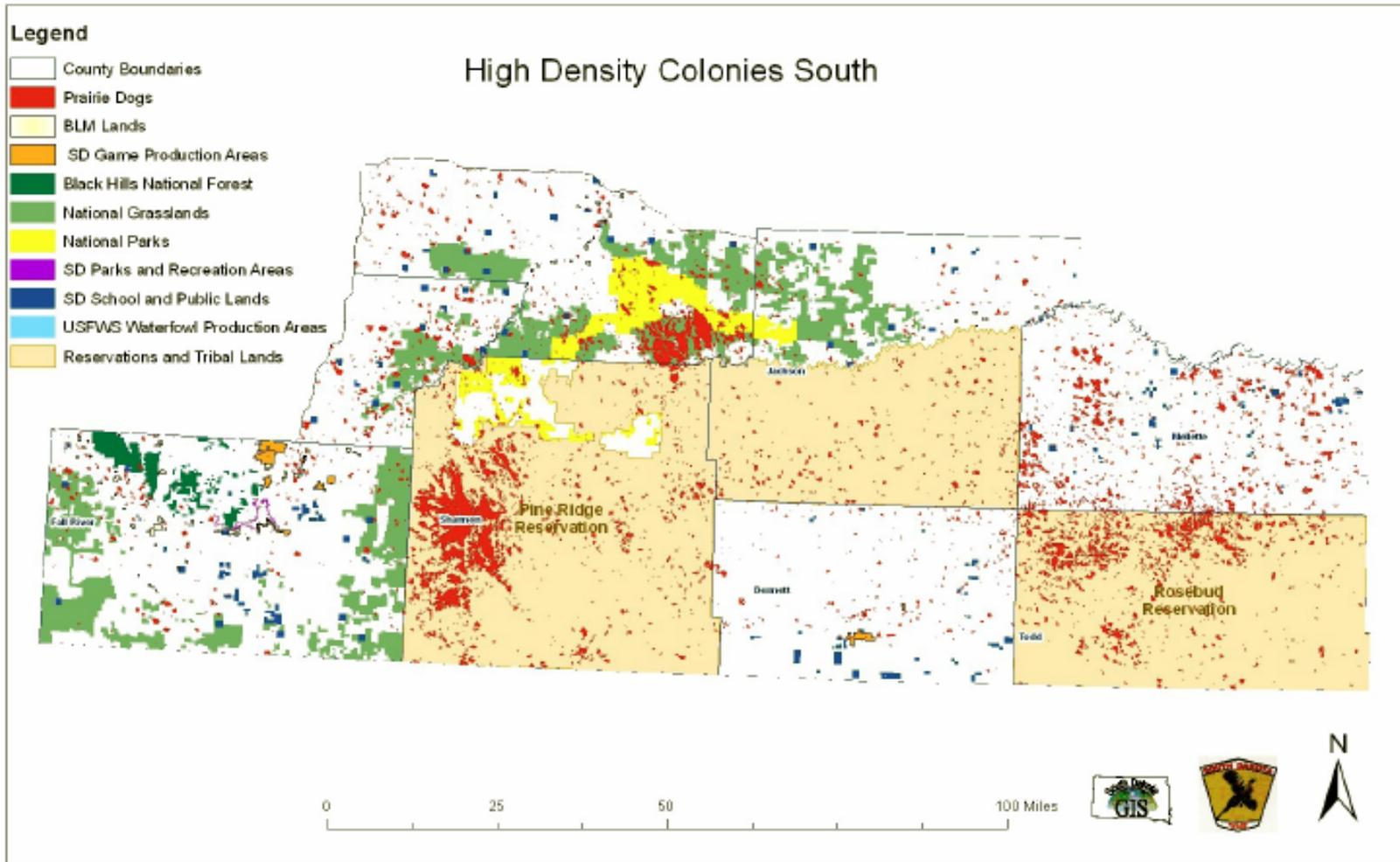


Figure 3. Distribution of black-tailed prairie dog colonies in southwestern South Dakota in 2006 (Kempema 2007). Most of the prairie dog colony acreage is located on reservations and tribal lands (extensive tribal lands in Mellette and Bennett counties are not depicted). The large red areas in the center of the map are colonies in Conata Basin, Buffalo Gap National Grassland.

Table 2. The extent of black-tailed prairie dog colonies in the non-MA 3.63 Geographic Areas during 2002 and 2006. Also displayed are the extents of the BMZ, IMZ, suitable habitat, and the proposed action's caps on prairie dog colonies (Source: Nebraska National Forest, Geographic Information System). There are no prairie dog colony acreage objectives for Wall Southwest and Fall River Southeast Geographic Areas.

Geographic and Management Areas	Total Non-NFS Acres in Geographic and Management Areas	2002 Prairie Dog Colony Acreage (% of non-NFS acres in GA)	2006 Prairie Dog Colony Acreage	Total NFS Acres in Geographic Area	Total Suitable NFS Acres for Prairie Dogs	Total Acres in BMZ	Total Acres Suitable for Prairie Dogs in BMZ	Acres of Prairie Dog Colonies Poisoned in BMZ, 2006	Total Acres in IMZ	Total Suitable Acres (IMZ)	2002 Prairie Dog Colony Acreage (% of NFS acres in GA)	2006 Prairie Dog Colony Acreage (% of NFS acres in GA)
Oglala	121,274	NA <sup>1</sup>	NA	94,484	62,347 (66%)	64,902 (69%)	43,701	1,011	29,582	18,646	2,192 (2.3%)	1,125 (1.2%)
Fall River Northeast	119,265	5,008 (4%)	4294 (3.6%)	91,298	78,806 (86%)	51,334 (56%)	45,328	1,998	39,964	33,478	3,851 (4.2%)	1,130 (1.2%)
Fall River West	192,868	2,438 (1.2%)	2112 (1%)	119,951	116,434 (97%)	70,088 (58%)	68,014	780	49,863	48,420	1,358 (1.1%)	210 (0.17%)
Fall River Southeast	189,165	821 (0.4%)	722 (0.4%)	86,666	84,142 (97%)	61,515 (71%)	59,448	411	25,151	24,694	1,509 (1.7%)	42 (0.04%)
Smithwick MA 3.63	1	0	0	25,307	24,187 (95%)	7,658 (30%)	7,177	68	17,649	17,010	990 (4%)	503 (2%)
Wall North	76,450	426 (0.5%)	454 (0.6%)	69,437	63,894 (92%)	53,418 (77%)	49,093	940	16,019	14,801	397 (0.5%)	454 (0.6%)
Wall Southeast	86,107	975 (1%)	982 (1%)	90,840	79,017 (87%)	58,000 (64%)	51,132	1,370	32,840	27,885	1,312 (1.4%)	1,414 (1.5%)
Wall Southwest	57,810	1,231	1,701	28,580	15,125 (53%)	14,456 (50%)	8,432	279	14,124	6,693	186 (0.6%)	214 (0.7%)
Conata Basin MA 3.63	353	81	77	77,155	68,183 (88%)	25,149 (37%)	21,784	5,251	52,006	46,399	22,530 (29%)	26,484 (34%)
Fort Pierre	74,921	130 (0.17%)	163 (0.2%)	116,053	108,409 (93%)	50,639 (44%)	47,195	621	65,414	61,214	1,346 (1.1%)	1,735 (1.5%)
Total (Fall River Northeast through Fort Pierre)	918,214	11,110 <sup>2</sup> (1.4%)	10,505 <sup>2</sup> (1.3%)	799,771	700,544 (87%)	457,159	401,304	12,729	342,612	299,240	35,671 (4.4%)	33,311 (4.1%)

<sup>1</sup>No information is available on the extent of prairie dog colonies on the Oglala GA.

<sup>2</sup>Based upon total non-NFS acres (428,337) in Fall River Northeast, Fall River west, Wall North, Wall Southeast, and Fort Pierre.

## 4. Proposed Action (Alt. 1) and Alternatives Analyzed in the FEIS

The proposed action is to amend current management direction in the 2001 Forest Plan to: 1) specify the desired range of acres of prairie dog colonies on in non-MA 3.63 GAs on Buffalo Gap, Fort Pierre, and Oglala national grasslands; and 2) allow toxicants if the acreage exceeds the desired range and for multiple use objectives. The effects considered in this analysis are:

- ◆ Effects of setting minimum and maximum areas of prairie dog colonies for non-MA 3.63 GA.
- ◆ Effects of rodenticide application in the IMZ on black-footed ferrets and whooping cranes, the two species listed under the ESA that are known to occur or may occur in all or some of the non-MA 3.63 GAs.

This analysis assumes prairie dog colonies within the BMZ will be treated with rodenticide within the life of this project. For the purposes of this analysis, however, the BMZ will be considered to be free of prairie dogs.

### The Alternatives

Below is a summary of each alternative. Refer to Chapter 2 of the FEIS for more details on each alternative. Table 3 displays the extent of prairie dog colonies under each alternative. The alternatives propose a range of prairie dog colony acres by GA. The “no action” alternative, Alternative 2, does not specify any specific range of acres, however, estimated acreages are provided based upon observed colony expansion rates. Alternative 4 only proposes a range of acres for the Conata Basin Black-footed Ferret Management Area 3.63. Alternatives 1, 3, and 4 also provide for use of rodenticide to reduce prairie dog density if a threshold based on Ecological Similarity Index or Desired Vegetation Condition is exceeded.

#### *Alternative 1*

This alternative employs adaptive management in emphasizing a mix of multiple uses while sustaining black-footed ferrets and associated species within Management Area 3.63 (MA 3.63). This alternative is based on a moderate objective for prairie dogs while incorporating adaptively applied active and passive management tools. Alternative 1 strives for a more balanced allocation between social/economic and biological issues while also sustaining black-footed ferrets. In the Conata Basin MA 3.63 where black-footed ferrets currently exist, this alternative prioritizes black-footed ferrets and the associated need for prairie dog colonies over other multiple uses. This alternative utilizes an Adaptive Response Protocol (See FEIS Appendix H) to help make implementation decisions at the site-specific level.

Outside of MA 3.63, maximum prairie dog colony acres are 3% of the total aggregate acres within each GA. The minimum acres center around an objective to establish one prairie dog colony complex on each GA that does not include 3.63 Black-footed Ferret Reintroduction Habitat. By definition, a prairie dog colony complex is a minimum of 1,000 acres of prairie dogs in 10 or more colonies with inter-colony distances not exceeding 6 miles. Higher acreages of active prairie dogs will be maintained in the 3.63 Black-footed Ferret Reintroduction Habitat.

#### *Alternative 2*

This alternative is the current prairie dog management as defined in the 2001 Forest Plan and 2005 amendment. The 2001 Forest Plan did not set specific acreage objectives (caps) for prairie dog colonies. The current management objective for prairie dogs located in the IMZ is to regulate and

manage prairie dog populations through non-lethal methods and limited rodenticide use where human health and safety or infrastructure is threatened.

The 2001 Forest Plan's prairie dog expansion model estimated 24,400-39,800 acres of prairie dog colonies on the NNF by 2012. The 2005 Forest Plan amendment estimated 29,600-41,400 acres 2012. There are currently 33,310 acres of occupied prairie dog colonies. Observed annual growth rates on the NNF indicate an average expansion rate of 25 percent, the rate used in the current EIS to predict acres on the NNF in the next 10 years. This rate includes recent years when extended drought conditions have facilitated prairie dog colony expansion, although not necessarily prairie dog populations. The 2001 Forest Plan has allowed black-footed ferret habitat (prairie dog colonies) to increase consistent with the priority of 3.63 management areas.

### ***Alternative 3***

This alternative ensures that there is not a disproportionate share of prairie dog colony acres in any county containing federal lands. In South Dakota, this alternative only allows a maximum acreage of prairie dog colony acreage of 3% the total aggregate federal ownership in each county. Any minimum numbers are essentially set by the vegetative condition of each prairie dog colony based on a NRCS Ecological Site Similarity Index threshold (Natural Resource Conservation Service – USDA 2006). If the similarity index falls below 25%, the prairie dogs will be reduced to 10% of the active colony acreage. On the Oglala National Grassland, Nebraska, the range of acres of active prairie dog colonies is 100-900 acres.

### ***Alternative 4***

This alternative is derived from the South Dakota State Prairie Dog Plan (Cooper and Gabriel 2005). It calls for 8,000-12000 acres of prairie dog colonies in Conata Basin. The state plan did not specify colony acreage outside of Conata Basin. The plan set objectives at the state-wide level for tribal and non-tribal acreage and did not set specific acreage objectives for all federal lands included in this analysis.

### ***Alternative 5***

This alternative emphasizes prairie dogs and black-footed ferrets. In non-MA 3.63 areas, prairie dog colony acreage represents 10% of the total acres in the GA as a minimum and 20% as a maximum. In the 3.63 Black-footed Ferret Reintroduction Habitat in Conata Basin, the maximum acre objective is based on what is required to maintain 200 breeding adult ferrets (using average black-footed ferret home range data collected in Conata Basin and adjusted for unoccupied areas (Livieri 2007a, c), and includes a range of acres based on prairie dog densities. During drought, prairie dog colonies expand but there are fewer prairie dogs per acre. This low density of prairie dogs leads to larger home ranges of black-footed ferrets. During wet periods, prairie dog colonies exhibit higher densities of prairie dogs and black-footed ferret home ranges contract. The minimum acre objective for Conata Basin is based on what is required to maintain 120 breeding adult black-footed ferrets. The prairie dog colony acre objective for the Smithwick black-footed ferret area is based on what is required to maintain 50 breeding adult ferrets during wet and dry periods.

**Table 3. Black-tailed prairie dog colony acreage under the alternatives for non-MA 3.63 areas.**

Geographic Area	Current Acreage (2006)	Desired range of acres of prairie dog colonies				
		Alt 1 <sup>1</sup>	Alt 2 <sup>2</sup>	Alt 3 <sup>1</sup>	Alt 4 <sup>3</sup>	Alt 5 <sup>1</sup>
Oglala	1,125	1,000-2,800	1,125-13,097	100-900	N/A	9,500-18,900
Fall River Northeast	1,130	1,000-2,700	1,130-13,155	2,700	N/A	9,100-18,300
Fall River West	210	1,000-3,600	210-2,445	3,600	N/A	12,000-24,000
Fall River Southeast	42	0	42-489	2,500	N/A	8,700-17,300
Wall North	454	1,000-2,100	454-5,285	2,100	N/A	6,900-13,900
Wall Southeast	1,414	1,000-2,700	1,414-16,461	2,700	N/A	9,100-18,200
Wall Southwest	214	0	214-2,491	830	N/A	2,600-5,100
Fort Pierre	1,735	1,000-3,500	1,735-20,198	3,470	N/A	11,600-23,200

<sup>1</sup>Desired minimum and maximum range of prairie dog colony acres.

<sup>2</sup>Range of acres is current acreage (2007) and projected acreage in ten years.

<sup>3</sup>This alternative does not contain a desired range of prairie dog colony acreage for non-MA 3.63 areas.

## 5. Endangered Species Considered in the Analysis

On April 5, 2007 a list of species to be evaluated was received from USFWS. Four species, the least tern, piping plover, Ute's Ladies' Tresses, and pallid sturgeon do not occur on any units of the Nebraska National Forest nor is there any suitable habitat that could be affected, and therefore these species are not analyzed further. Two federally-listed species are analyzed in detail in this report: black-footed ferret and whooping crane, which are known to occur or may be present on the Nebraska National Forest (Table 4).

**Table 4. Species listed as endangered under the Endangered Species Act and located in the Project Area.**

	Buffalo Gap National Grassland						Fort Pierre National Grassland	Oglala National Grassland
	Fall River West GA	Fall River SE GA	Fall River NE GA	Wall North GA	Wall SW GA.	Wall SE GA		
Black-footed ferret <sup>1</sup>	---	---	---	---	K	K	---	---
Whooping crane	P	P	P	P	P	P	P	P

K = Known occurrence; P = Possible but unconfirmed occurrence

<sup>1</sup>Non-essential experimental population (USFWS 1994)

The black-footed ferret MA 3.63 areas are not the subject of this BA, however, for context and clarity it is useful to briefly discuss these areas. The Conata Basin/Badlands reintroduction area was designated as

non-essential experimental population under Section 10(j) of the ESA (59 Federal Register 42682-42694; August 18, 1994) (Figure 4). A 10(j) black-footed ferret population is not afforded the same level of protection as a fully endangered species but is treated as a “proposed” species for the purposes of managing the reintroduction sites on National Forest System lands. On National Park Service lands, the species’ status is considered threatened. The relaxed standards associated with a 10(j) designation are sometimes necessary to provide management flexibility while attempting to establish the population and also to sustain support for reintroduction efforts at the state and local level. A non-essential experimental population designation does not convey that the population is not important for the recovery of the species. To the contrary, Conata Basin/Badlands is foundational to black-footed ferret recovery and Section 10(j) populations are counted towards meeting recovery goals.

The reintroduction area (three units) in 1994 was about 42,000 acres. The experimental population area is about 1.2 million acres. In 1994, there were about 8,000 acres (much of the acres in Badlands National Park) of black-tailed prairie dog colonies in the three units. Other colonies occurred outside the units. The 1994 record of decision designated 206,300 acres of Buffalo Gap National Grassland outside the reintroduction area as dispersal habitat for black-footed ferrets that move outside the reintroduction area. USFWS et al. (1994) envisioned that black-footed ferrets would not simply remain within the three units illustrated above but rather would move to other colonies and that greater acreages of prairie dog colonies would be required. USFWS (1992; cited in USFWS et al. (1994) recommended that a site with 11,700 to 23,500 acres (or more) of active prairie dog colonies with an average prairie dog density of 5 prairie dogs per acre represents habitat for sustaining a population of 95 breeding black-footed ferrets.

The **Conata Basin and Smithwick MA 3.63** was established on Buffalo Gap National Grassland by the 2001 Forest Plan (Figure 5). The 3.63 Area encapsulates the 1994 reintroduction area and much of the dispersal habitat on Buffalo Gap into the 2001 Forest Plan. The theme or purpose of the area is uniquely black-footed ferret reintroduction habitat, that is providing prairie dog colonies to support a viable black-footed ferret population. Livestock grazing and other uses may occur in the 3.63 Area but fundamentally the emphasis is on habitat of the black-footed ferret. Specific acreages of prairie dog colonies were not enumerated in the 2001 Forest Plan because of uncertainties surrounding black-footed ferret habitat requirements and population size. The 2001 Forest Plan afforded flexibility in managing an endangered species.

Under the 2001 Forest Plan, “black-tailed prairie dog colony complexes are to be actively and intensively managed as reintroduction habitat for black-footed ferrets.” Desired conditions are “large prairie dog colony complexes established and maintained as suitable habitat for black-footed ferret reintroductions. Land uses and resource management activities are conducted in a manner that is compatible with maintaining suitable black-footed ferret habitat. The Forest Service works with other agencies and organizations to pursue conservation agreements or easements with adjoining land jurisdictions to achieve black-footed ferret recovery objectives. Where landownership patterns are not conducive to effective and successful prairie dog and black-footed ferret management, landownership adjustments with willing landowners may also be used to help resolve management issues.” The Forest Service “authorizes only those uses and activities that do not reduce the suitability of the area as black-footed ferret reintroduction habitat.” All prairie dog colonies within this area “are managed as though they are occupied by black-footed ferrets.” In addition, “Relocation of prairie dogs to establish new colonies and accelerate growth of prairie dog populations in selected areas may occur only after consultation with appropriate state and federal wildlife agencies. To help expand and maintain suitable black-footed ferret habitat, prairie dog shooting is prohibited. In 2005, the 2001 Forest Plan was amended and a boundary management zone (BMZ) was established within 0.5 miles of private and tribal lands where control of prairie dogs may occur. The BMZ effectively removed 25,149 acres (21,784 acres suitable for black-tailed prairie dogs) from the Conata Basin MA 3.63, assuming that all adjoining landowners would file valid complaints when prairie dogs expanded onto or approached their land. The BMZ effectively removed 7,658 acres

(7,177 acres suitable for black-tailed prairie dogs) from the Smithwick MA 3.63, again assuming that all adjoining landowners would not tolerate colonization by prairie dogs from the national grasslands.

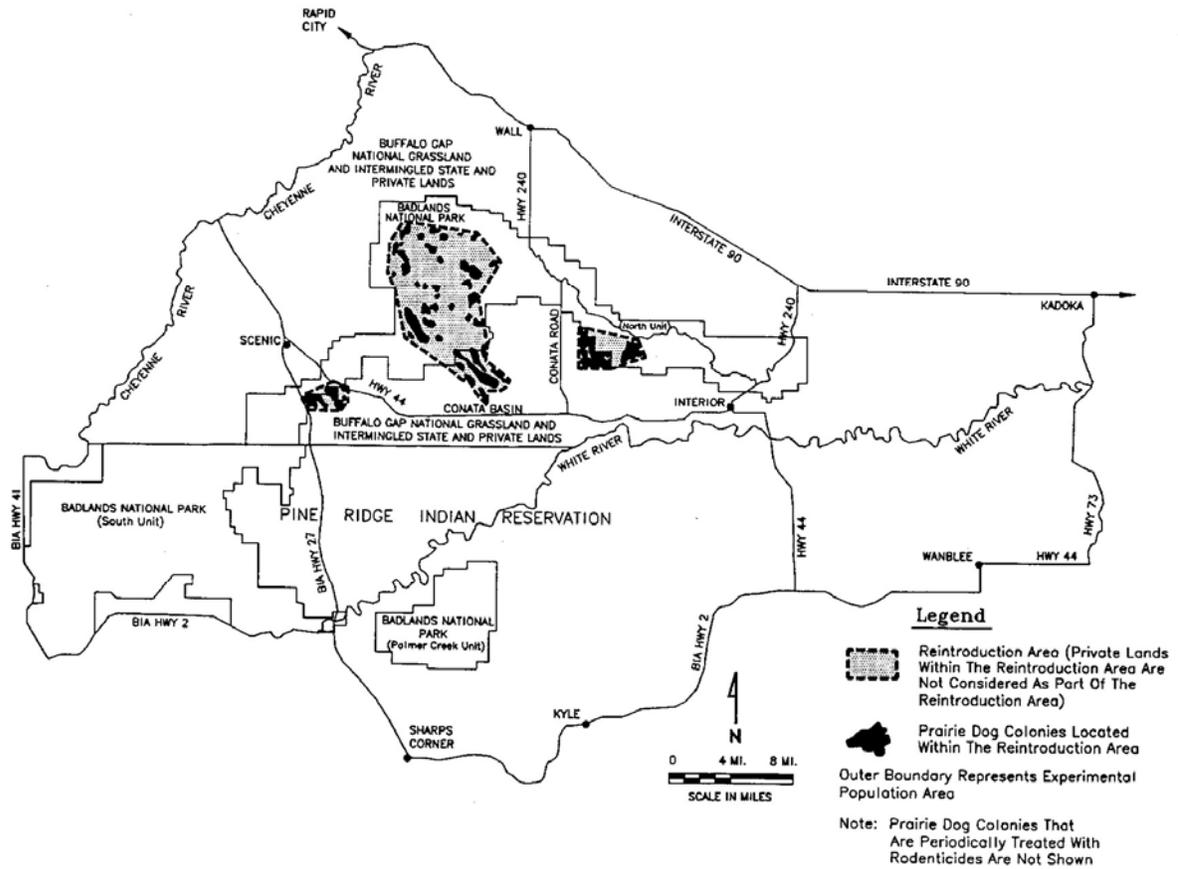
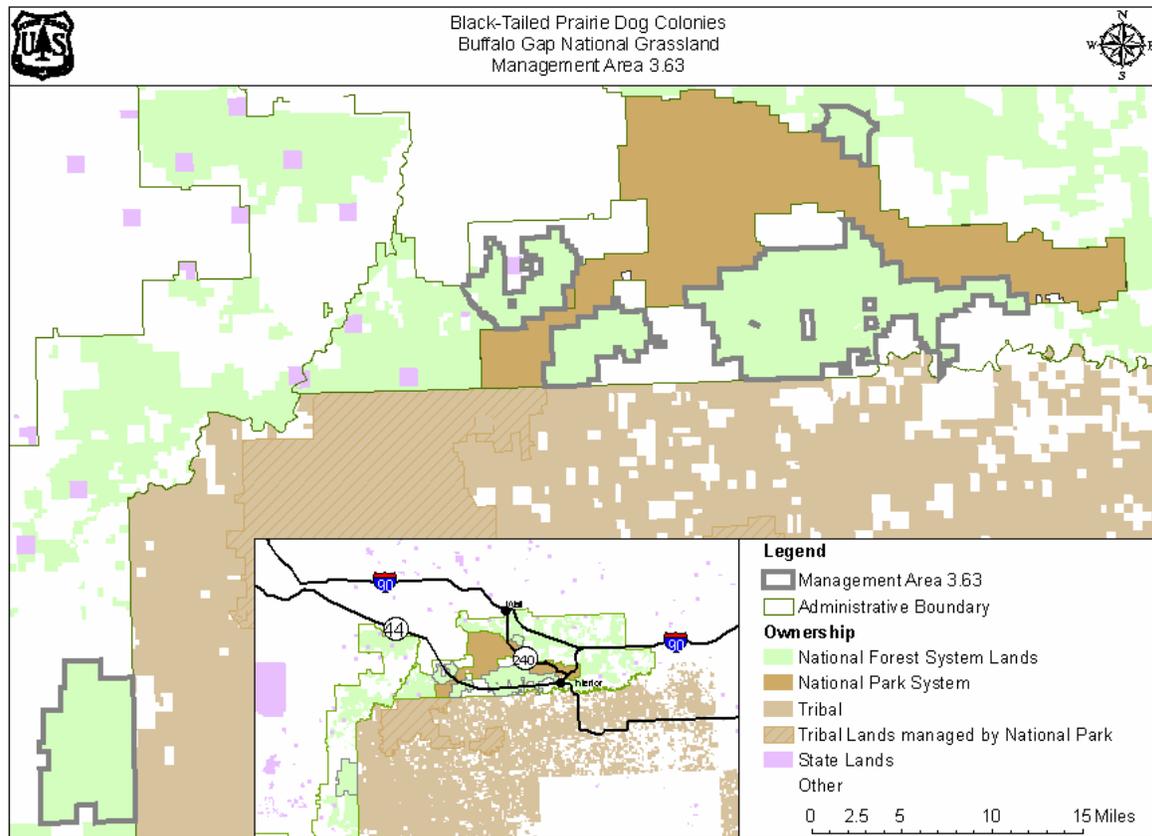


Figure 4. The Conata Basin/Badlands Black-footed Ferret Reintroduction Area and Experimental Population Area (outer boundary) established in 1994; the reintroduction area was selected by the U.S. Forest Service (31 May 1994 record of decision) and based upon USFWS et al. (1994).



Management Area 3.63	Total Acreage	Suitable	Unsuitable	BMZ Acreage	Suitable	Unsuitable	IMZ Acreage	Suitable	Unsuitable
Smithwick	25,307	24,187	1,120	7,658	6,696	481	17,649	17,010	639
Conata Basin	77,500	68,182	8,972	21,860	18,495	3,365	55,294	49,687	5,607

Figure 5. The **Conata Basin and Smithwick Management Areas 3.63** (outlined in grey above). Conata Basin is in the center and Smithwick is in the lower left corner. Suitable and unsuitable habitat (acres) for black-tailed prairie dogs in the Management Area 3.63 (black-footed ferret management areas) on Buffalo Gap National Grassland, including the size and suitability of the boundary management zone (BMZ) and interior management zone (IMZ). About 88% (68,182 acres) of Conata Basin 3.63 is suitable habitat for prairie dog colonies (black-footed ferret habitat). However, the 2005 BMZ decision changed suitable habitat for prairie dog colonies from 68,182 acres to 49,687 acres, a 27% reduction. The 2005 BMZ decision changed suitable habitat for prairie dogs in the Smithwick from 24,187 acres to 17,649 acres, a 27% reduction. There are currently about 26,000 acres of prairie dog colonies in Conata Basin 3.63, leaving another 23,687 acres for potential expansion or colonization as prairie dog colony extent shifts from current locations.

## 6. The Black-footed Ferret's (*Mustela nigripes*) Status, Ecology, and Prairie Dog Habitat

The black-footed ferret is in the Order Carnivora, Family Mustelidae, Genus *Mustela*, and Subgenus *Putorius* (Hillman and Clark 1980). The species is one of five *Mustela* in North America that include the ermine (*M. erminea*), long-tailed weasel (*M. frenata*), least weasel (*M. nivalis*), and American mink (*M. vison*) (Wilson and Ruff 1999). The black-footed ferret is the only ferret species native to the Americas.

Other ferret species in the genus include the Siberian polecat (*M. eversmanni*) and the European ferret (*M. putorius*). The black-footed ferret is most closely related to the Siberian polecat, from which ancestral black-footed ferrets arose (Anderson et al. 1986, Hillman and Clark 1980). The earliest fossil record of the black-footed ferret is from approximately 100,000 years ago (Anderson et al. 1986). The black-footed ferret was formally described in 1851 by J.J. Audubon and J. Bachman (Anderson et al. 1986, Clark 1986).

The black-footed ferret is a medium-sized mustelid weighing 1.4-2.5 pounds and measuring 19-24 inches in total length. Upper body parts are yellowish buff, occasionally whitish; feet and tail tip are black; and a black “mask” occurs across the eyes (Anderson et al. 1986, Hillman and Clark 1980).

### Distribution and Status

Our understanding of historic black-footed ferret distribution derives from the literature and museum specimens. Black-footed ferrets from Arizona, Colorado, Kansas, Montana, Nebraska, New Mexico, North Dakota, Oklahoma, South Dakota, Texas, Utah, Wyoming, Alberta, and Saskatchewan have been collected since the late 1800s (Anderson et al. 1986). The black-footed ferret formerly occurred at colonies of three species of prairie dogs (Anderson et al. 1986) (Figure 6). Over 84 percent of black-footed ferrets occurred at black-tailed prairie dog colonies, the species with the largest range of any prairie dog species (Ernst et al. 2006).

The black-footed ferret was probably common, although its secretive habits (nocturnal and often underground) made it difficult to observe (Anderson et al. 1986, Clark 1989, Forrest et al. 1985). Given the historic extent of prairie dog colonies, Anderson et al. (1986) estimated that 500,000 to 1,000,000 black-footed ferrets once populated prairie dog colonies. For most of the twentieth century, however, the species was believed almost extinct due to the near eradication of prairie dogs, especially complexes of thousands of acres of prairie dog colonies. The first opportunity to study black-footed ferrets was during 1964-1974 at black-tailed prairie dog colonies in Mellette County, South Dakota. This remnant

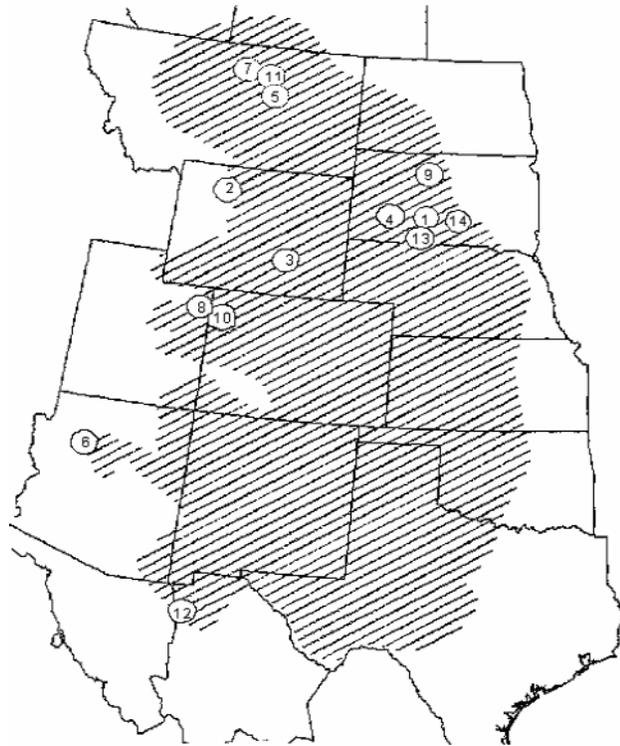


Figure 6. Probable historic range of black-footed ferrets as defined by composite ranges of black-tailed (*Cynomys ludovicianus*), white-tailed (*C. leucurus*) and Gunnison's prairie dogs (*C. gunnisoni*). Numbers 1 and 2 are the Mellette County, South Dakota and Meeteetse, Wyoming populations that no longer exist. Other numbers are re-introduction sites that are referenced in the table on the following page (USFWS 2006b).

population was disappearing and removed black-footed ferrets died in captivity during captive breeding attempts. Black-footed ferrets were discovered in 1981 in white-tailed prairie dog colonies near Meeteetse, Wyoming (Clark et al. 1986, Miller et al. 1996). In 1985, sylvatic plague, a lethal disease to prairie dogs and black-footed ferrets, struck the prairie dogs at Meeteetse killing most prairie dogs.

Lethal canine distemper also was discovered in the Meeteetse black-footed ferrets. In 1986, the remaining eighteen Meeteetse black-footed ferrets were removed to protect them from distemper and to start a captive breeding program (Miller et al. 1996). Seven of the black-footed ferrets were successfully bred in captivity (Garrelle et al. 2006, Hutchins et al. 1996). The extant population today, both captive and wild, descends from these seven “founder” animals. Several thousand black-footed ferrets have been produced in captivity (Conservation Breeding Specialist Group 2004). A goal of the breeding program is to retain as much genetic diversity as possible, although intensive searches throughout the black-footed ferret’s range have not discovered any wild black-footed ferrets since the Meeteetse find. It is very unlikely that any undiscovered wild populations remain (Hanebury and Biggins 2006, Lockhart et al. 2006).

There are about 250 breeding adult black-footed ferrets in the wild. About 100 occur in Conata Basin. The number of breeding adults is the standard for measuring the status of black-footed ferret populations. The current wild population is well short of the objective to ensure the immediate survival of the species by establishing 1,500 breeding adults in 10 or more populations by 2010 (USFWS 1988). Since 1991, black-footed ferrets have been reintroduced into twelve sites including one site in Mexico. Two sites have failed to sustain a black-footed ferret population while other sites display varying results. The largest populations currently exist at Shirley Basin, Wyoming, Conata Basin, South Dakota and Cheyenne River Sioux Reservation, South Dakota. Wild animals from Conata Basin and Cheyenne River have been translocated to other reintroduction sites. Agencies and others tasked with black-footed ferret recovery desire large self-sustaining populations with the goal of eliminating or reducing the need for supplementation with captive-reared kits and thus the eventual elimination or reduction of the costly captive breeding program at the Black-footed ferret Conservation Center near Wellington, Colorado and participating zoos.

There are few successful black-footed ferret sites, that is, sites free of plague and containing large prairie dog colony acreages. In 2006, Wind Cave National Park, South Dakota and northwestern Kansas were selected in preliminary black-footed ferret allocations. These sites and Lower Brule Sioux Reservation are much smaller than previous reintroduction sites, but provide an opportunity for new sites east of the current plague line, cooperative efforts with new partners, and potential refugia in the event of increased plague activity.

More populations are needed to enhance overall species’ viability and reduce the risk of extinction. In the past, the black-footed ferret had an enormous range in the center of North America and perhaps up to one million existed (Anderson et al. 1986). Fundamentally, most small animal populations face a particularly high risk of extinction in the modern world (Gaggiotti and Hanski 2004). This notion is as old as the writings of Charles Darwin and as recent as today (Holsinger 2000). The high extinction risk of small populations is not surprising because this is the expectation based on several mechanisms of extinction: demographic and environmental stochasticity, Allee effect, inbreeding depression, and so forth. As the different mechanisms tend to make populations ever smaller, they amplify the effect on each other and lead to extinctions (Gilpin and Soulé 1986).

## **Prairie Dog Habitat of the Black-footed Ferret**

The exact acreage of prairie dog colonies that once existed is unknown and debate on the issue continues (Vermeire et al. 2004). Historical information on species is almost always lacking. Anderson et al. (1986) estimated a 90 percent decrease in occupied habitat for all species of prairie dogs. Recent estimates of prairie dog occupied habitat include 1,600,000 acres of black-tailed prairie dog occupied

habitat (69 Federal Register 51217-51226, August 18, 2004; Manes 2006; Van Pelt 2007). The existence of an obligate carnivore, the black-footed ferret, suggests strongly that prairie dog colonies were once numerous and extensive. Moreover, using the USFWS (2004) current estimates for prairie dogs, the declining population trend is clear whether there were once 40,000,000 ha, 30,000,000 ha, 20,000,000 ha, or 10,000,000 ha of colonies. Respectively, these figures indicate declines of roughly 98%, 97%, 96%, or 92% (Miller et al. 2007).

The black-footed ferret is known to inhabit almost exclusively large complexes of prairie dog colonies of thousands of acres. Colonies provide the black-footed ferret with shelter and its only food source, prairie dogs. As prairie dog colonies and suitable habitat declined due to conversions from native prairie rangeland to cropland, poisoning and plague during the late 19<sup>th</sup> century and the first half of the twentieth century, black-footed ferret populations likewise declined (Biggins 2005, Cully 1993, Cully et al. 2006, Fagerstone and Biggins 1986, Lockhart et al. 2006). A typical landscape with prairie dogs in the Great Plains today contains widely scattered colonies of less than 50 acres each. Large complexes of thousands of acres of prairie dog colonies are uncommon and are found almost entirely on tribal and federal lands. Over 100 years ago, Merriam (1902) described a 25,000 square mile prairie dog colony in the north Texas short grass prairie. He also noted that colonies 20-30 miles long were common. Colonies of this magnitude no longer exist. Poisoning, plague, and land fragmentation confine black-footed ferret habitat to small areas. Grassland continues to be converted to cropland. Conversion of grassland that had no prior cropping history to cropland in South Dakota, for example, was 54,404 acres in 2005 and 47,167 acres in 2006 (U.S. Government Accountability Office 2007).

Black-footed ferret habitat, that is, the prairie dog colony, is a unique feature in the Great Plains. The colonies are easily visible on the ground, to aircraft, and even to orbiting, imaging satellites because they are fundamentally different from surrounding grassland. The U.S. Geological Survey's vegetation mapping program classifies prairie dog colonies as the "black-tailed prairie dog town grassland complex." This complex occurs widely throughout the Great Plains.

Visible features (conditions) at prairie dog colonies include numerous dome burrow mounds (up to 2.5 m in diameter) and crater mounds barren of vegetation, short-statured vegetation, and bare areas. Prairie dogs are diurnal and short vegetation, barren mounds, and bare areas greatly facilitate the detection of predators. In addition, prairie dogs feed on grasses and forbs. Prairie dogs are attracted to livestock and other ungulate grazing. Such grazing along with prairie dog herbivory helps establish and even expand prairie dog colonies by increasing the amount of short-statured vegetation and bare areas. Drought restricts plant growth leading to expansion of prairie dog colonies, more short-statured vegetation, and more bare areas than during times of normal precipitation. Prairie dog densities decrease because food resources decrease during drought leading to an increase in the home range sizes of black-footed ferrets and other carnivores.

Fundamentally, the clipping and foraging habits of prairie dogs create a unique habitat (unique conditions) of bare ground and short, sparse vegetation (Archer et al. 1987, Natural Resources Conservation Service 2003). Some species such as the mountain plover require prairie dog colonies containing from 40-60% bare ground (Knowles et al. 1982). Prairie dog colonies can contain from near 0% bare ground to 80% bare ground (Magle 2003) depending upon precipitation, grazing, and colony age. Vegetation species diversity and per cent bare ground changes under different levels of impacts such as grazing and drought (Whicker and Detling 1988). Plant biomass production and forage quality is positively correlated to precipitation. Black-footed ferrets occur even in large areas of bare ground in colonies (Travis Livieri, personal communication), however, the relationship between productivity of black-footed ferrets and plant biomass production and forage quality is not well understood.

Black-footed ferret habitat (prairie dog colonies) is located on a wide variety of soils, including clay, clay loam, silty loam and some sandy loam soils deposited following erosion from adjacent uplands, including badlands formations. Soils are deep, structured and not easily eroded. This type is found on level sites

along drainages, in broad valleys, on gentle to moderately sloping hills, and flats on tables and buttes (Von Loh et al. 1999). Prairie dogs create extensive burrows in their towns. Large volumes of soil are moved, improving filtration, hastening the incorporation of organic matter, facilitating nutrient cycling, and increasing the spatial heterogeneity of vegetation, soils, and other ecosystem components (Whicker and Detling 1988).

Although several plant species are consistently found in black-footed ferret habitat, overall vegetation characteristics are highly variable depending upon size and age of the town and its position on the landscape. The vegetation in all colonies usually tends toward a prostrate growth form because of intensive grazing by prairie dogs and livestock. The overall pattern of vegetation appears in relatively concentric zones relating to the outward expansion of town boundaries over time. Vegetation cover varies from <25% to almost 100%.

Prairie dog colonies (black-footed ferret habitat) are located in open mixed grass or short grass prairie, and their activity has both direct and indirect effects on the vegetation. Prairie dogs keep the surrounding vegetation clipped close to the ground, presumably to improve their ability to detect stalking predators. This clipping gives the impression of a mowed lawn, or overgrazed rangeland. Prairie dogs repeatedly clip and graze plants, rarely allowing shoots to reach full size. Thus, canopy height within the colony is about 5-10 cm, compared to 20-50 cm in nearby, uncolonized grassland (Whicker and Detling 1988). Changes in plant species composition may begin as early as 2 or more years after colonization. Shortgrass species, such as blue grama (*Bouteloua gracilis*) and buffalo grass (*Buchloe dactyloides*), and annual forbs, become abundant and replace mid-height or tall grasses. Continued heavy grazing may eventually result in complete dominance by a few species of forbs or dwarf shrubs (Whicker and Detling 1988). Grazing may even cause genetic shifts within species. The shorter, more prostrate, growth forms of western wheatgrass (*Pascopyrum smithii*), for example, on prairie dog colonies have been shown to be more abundant than those away from colonies, suggesting that some genotypes within the species may tolerate grazing better than others (Jaramillo and Detling 1988, Whicker and Detling 1988).

Bison and livestock may be attracted to prairie dog colonies, and may preferentially graze them (Coppock et al. 1983, Coppock and Detling 1986, Day and Detling 1990). The forage on the colonies is more nutritious than off, with higher nitrogen content and younger shoots, apparently because animal waste products are deposited there. In turn, the presence of bison and livestock waste products further increases soil fertility and forage quality (Knight 1994).

Plant species diversity is increased by the small-scale disturbances caused by the digging of prairie dogs, and animal species diversity may also increase because of the habitat provided for the badger, rattlesnake, burrowing owl, black-footed ferret, and cottontail, in addition to the bison and pronghorn (Knight 1994). Prairie dog colonies can also move over time, expanding and contracting, and, as larger colonies can cover thousands of acres at a time (Augustine et al. In Press), the effect on the landscape is substantial. The plant community types on a prairie dog colony are roughly indicative of the extent of herbivore disturbance and reflect the cumulative impact of grazing intensity, grazing duration, activities of other animals, soil characteristics, and weather (Whicker and Detling 1988). Early stages of the colony may have a typical mixed grass or shortgrass prairie type. With continued grazing and age of the colony, the composition may shift to a mix of annual species and dwarf-shrubs. Species richness appears to be highest under moderate levels of disturbance, because grass species have not yet begun to disappear, but forb species begin to increase. In general, in the mixed grass prairie, prairie dog colonies reduce plant productivity and plant species richness by reducing cool-season perennial grasses and litter, and increasing bare ground (Johnson-Nistler et al. 2004).

There is a strong relationship between prairie dog colonies and ungulate grazing. Cattle-prairie dog interactions clearly show a synergistic relationship between cattle grazing and prairie dog habitat occupancy (Belak 2001). Osborn and Allen (1949) credited regenerating tall vegetation in a fallow pasture for the elimination of a small prairie dog colony located in tall-grass prairie. Snell and

Hlavachick (1980) found that elimination of cattle grazing during the growing season within and around a large colony in Kansas and the resulting vegetation growth was associated with a 90% decrease in the size of the colony. Knowles (1982), studying prairie dog distribution in eastern Montana, concluded that prairie dog distribution was mainly influenced by heavy livestock grazing pressure and other land disturbances created by humans. Uresk et al. (1982) found that the burrow densities within Conata Basin increased twice as fast in areas grazed by cattle vs. ungrazed areas. Cincotta (1985) reported that prairie dog colony expansion on the adjacent Badlands National Park was greater in heavily grazed areas and areas previously disturbed by homesteading activity. This was corroborated by Langer (1998) using a GIS analysis based on comparison of old homestead records with known prairie dog distributions. Cincotta et al. (1988) built a linear regression model to predict prairie dog establishment adjacent to existing colonies. They found that population density, visibility through vegetation, and the interaction of these two terms were significantly associated with colony expansion. These studies indicate that prairie dogs prefer areas with reduced vegetation density created by grazing pressure.

The role of livestock grazing in creating black-footed ferret habitat has been evident at the boundary of Buffalo gap National Grassland and Badlands National Park. Badlands National Park contains a comparatively small acreage of prairie dog colonies because much of the park is comprised of Badlands topography and because little ungulate grazing occurs in the park (Figure 7). For over 20 years prairie dog colonies on the National Grasslands in Conata Basin stopped at the fence line with the Park, although with severe drought some incursions of colonies into the Park have occurred. About 4,000 acres of colonies existed in Badlands during the 1994 reintroduction in Badlands. Colony acreage has increased from 5,629 to 7,867 during 2003-2006.



Figure 7. The boundary and fence line between Badlands National Park and Buffalo Gap National Grassland in part of Conata Basin during the early 1990s is illustrated by a different color tone. The stair-stepped lighter areas in the lower right are black-tailed prairie dog colonies on Buffalo Gap. The white areas, located mostly in the left section of the photograph, are Badlands topography in Badlands National Park. Extensive livestock grazing on Buffalo Gap fosters prairie dog colony development. No livestock grazing and a small bison herd in Badlands do not create the conditions for extensive prairie dog colonies required by black-footed ferrets. Buffalo Gap prairie dog colonies stop at the fence line. In recent drought years, there have been some prairie dog colony incursions into Badlands.

Black-footed ferret habitat can exhibit considerable antiquity. Carbon dating of mounds at Wind Cave National Park, South Dakota indicate that prairie dog colonies have existed there for about 400 - 800 years and other areas for several thousand years (White 1986, Carlson and White 1987), suggesting that despite periodic drought and varying levels of domestic and native ungulate grazing, black-footed ferret habitat persists in the same locations.

### **Prairie Dog as a Keystone Species in Creating Habitat for the Black-footed Ferret**

Paine (1969) developed the keystone species concept when species composition and physical appearance of the ecosystem is greatly modified by the activities of a single native species high in the food web. Populations of such species are the “keystone of the community’s structure, and the integrity of the

community and its unaltered persistence through time are determined by their activities and abundances.” The term keystone was borrowed from an old architectural word referring to the wedge-shaped stone at the highest point of an arch that locks the other stones in place and keeps the arch from collapsing (Mills 2007). The keystone concept can promote conservation not only for the prairie dog, but also for its grassland ecosystem (Kotliar 2000; Miller et al. 2000; Soulé et al. 2003). Soulé et al. (2005, 2007) labeled the prairie dog as a “strongly interacting species” in an ecosystem. They viewed as essential prairie dog population densities or levels that maintain interaction effectiveness rather than mere persistence at minimal numbers.

Miller et al. (2007) summarize the unique role played by prairie dogs. Like cattle, prairie dogs graze, but prairie dogs represent much more than just grazers (Kotliar et al. 2006). They also move soil (Detling and Whicker 1988, Detling 1998), influence nutrient cycling (Coppock et al. 1983a; Detling and Whicker 1988; Whicker and Detling 1988, 1993; Detling 1998), increase nitrogen content of soil and plants (Holland and Detling 1990, Detling 1998), change vegetation structure and community dynamics (Coppock et al. 1983a, Whicker and Detling 1988, Weltzin et al. 1997, Detling 1998, Fahnestock and Detling 2002), aerate the ground (Whicker and Detling 1988, Outwater 1996, Detling 1998), alter soil chemistry (Munn 1993), and deepen water penetration (Outwater 1996, Detling 1998). They provide a ready source of prey to many predators and burrows for shelter to other animals and insects (Goodrich and Buskirk 1998, Kotliar et al. 1999, Shipley and Reading 2006). This combination of effects gives the prairie dog its role as a highly interactive (keystone) species in the ecosystem, creating a matrix of different habitats that increases diversity across the grassland (Kotliar et al. 1999, 2006; Miller et al. 2000).

Kotliar et al. (2006) describe the prairie dog as a keystone species because the species has a unique, significant, disproportionately large impact on its ecosystem. The species’ influence on grassland ecosystems varies with its abundance and certain keystone functions also vary with abundance. Even though small, isolated colonies of prairie dogs are better than no colonies at all in the Great Plains, they cannot support the full complement of species that naturally associate with prairie dogs (Kotliar 2000, Lomolino et al. 2003, Johnson and Collinge 2004). About 100 years ago, both large and small colonies occurred throughout the prairie dog's geographic range (Proctor et al. 2006). A similar combination of small and large colonies (e.g., Conata Basin) is probably a good mechanism for maintaining today's prairie dogs and their keystone functions (Lomolino et al. 2003; Trudeau et al. 2004). Lomolino and Smith (2003) found that while species richness per se was not necessarily higher in prairie dog colonies, they harbored significantly more rare and imperiled species than paired sites in the adjacent landscape.

Agricultural interests have dictated artificially small limits on prairie dog populations for more than 100 years (Schenbeck 1981, 1985; Miller et al. 1990; Lamb et al. 2006, Miller and Reading 2006). Indeed, few prairie dog populations today are sufficiently large to support viable populations of black-footed ferrets. A healthy grassland ecosystem includes areas with and without prairie dogs, allows for dispersal among colonies, and promotes colonization of new areas and recolonization of sites where prairie dogs lived in the recent or distant past. Because they currently inhabit only a tiny fraction of their original geographic range, more prairie dogs are pivotal for a healthy grassland ecosystem.

Regarding herbivory, redundancy (i.e., consumption of plants by several species rather than by a single species) is important for the persistence of ecosystems (Walker 1991; Ehrlich and Walker 1998). As redundancy decreases (e.g., following the drastic declines of American bison and prairie dogs from the grasslands of western North America), the vulnerability of an ecosystem increases.

In summary, the prairie dog is a keystone species because its effects on its ecosystem are unique, significant, and disproportionately large relative to its abundance (Kotliar et al. 2006):

- ◆ Prairie dogs influence the grassland ecosystem via three primary pathways: grazing, burrowing, and as prey. Numerous species use colony-sites for food and shelter. Black-footed ferrets, mountain plovers, burrowing owls, and at least 6 other species clearly depend on prairie dogs for survival and reproduction; at least 20 species opportunistically benefit at colony-sites; and at least 117 additional species have a natural history that suggests a benefit from associating with prairie dogs.
- ◆ Besides vertebrates, other types of organisms that associate with prairie dogs include numerous species of protozoans, arachnids, and insects, and at least four plant species.
- ◆ Burrowing by prairie dogs mixes subsoil and topsoil, redistributes nutrients and minerals, and promotes penetration and retention of moisture. The combination of burrowing, foraging, and clipping alters floral species composition and affects the concentration of nitrogen in the soil and the rate of nitrogen uptake by plants. Prairie dogs reduce vegetation biomass, but, especially at young colonies, they often enhance digestibility, protein content, and productivity of grasses and forbs.
- ◆ The prairie dog's influence on plant and animal communities is substantial and unique, disproportionately large relative to its abundance, and critical to the integrity of grassland ecosystems. The prairie dog is therefore a legitimate keystone species. Because some of its ecological functions are directly proportional to its abundance (e.g., nutrient cycling), the prairie dog is also a foundation species.
- ◆ Because it is a keystone and foundation species, conserving the prairie dog will help to save its grassland ecosystem. The concepts of keystone and foundation species thus are helpful for conservation.

## **Black-footed Ferret Use of Prairie Dog Colonies**

Preliminary knowledge on how black-footed ferrets utilize prairie dog colonies has been gained from research in Conata Basin. Jachowski (2007) and Jachowski et al. (2008) examined resource selection by black-footed ferrets in relation to the spatial distribution of prairie dogs in Conata Basin and Montana (Figures 8 and 9). Although biologists have long mapped the boundaries of prairie dog colonies, the density of prairie dogs can vary considerably within a colony. Within prairie dog colonies, areas with high densities of active and inactive prairie dog burrows form patches and the distribution of these patches change in size, shape and connectivity over time, often a short amount of time (Jachowski 2007). The ability of an area to support a black-footed ferret population may be influenced by the size, availability, and density of high-density patches of prairie dogs (Biggins et al. 2006). Prairie dog distribution shifts likely in response to changes in vegetation species composition and productivity (Coppock et al. 1983, Detling and Whicker 1987). Prairie dog distribution moves to the perimeter of colonies but also back to the interior over time. Jachowski (2007) hypothesized:

...the relationship between vegetative cover and prairie dog occupancy is likely a dynamic process, where activity areas of prairie dogs shift spatially within colonies over time to enable long-term occupancy of a defined area or burrow system. Thus the inactive portion of prairie dog colonies should not be viewed as low in habitat value over the long term because these sites might again become populated.

Jachowski (2007) stated:

From a conservation perspective, our study supports the hypothesis that spatial distribution of patches of high prairie dog density is a major determinant of the ability of reintroduction sites to maintain ferret populations. Ferrets have relatively short life spans of 3-5 years (Forrest et al. 1988), experience a high rate of loss due to predation (Forrest et al. 1988, Biggins 2000), and have a low reproductive output relative to other mustelids (Forrest et al. 1985). These factors contribute to the difficulty of ensuring their recovery and underscore the importance of maintaining multiple breeding females at each recovery site. Our findings indicate that the ability of a site to support a greater number of female ferrets, and to achieve a self-sustaining ferret population, can be enhanced by management to create and maintain sufficiently large areas of high prairie dog density that are in close proximity to each other.

Black-footed ferrets compete for some space to some degree. Regardless of overlap, there is some competition and discussions continue of just how much spatial overlap exists and how to separate trespass from sharing, temporal overlap, and kinship effects. Mindful of the above uncertainty, black-footed ferret use of a colony, that is, its use of space, varies according to active prairie dog burrow densities (Figures 8 and 9). Lower prairie dog densities that are commonly associated with drought as well as disease, poisoning and possibly other unknown factors, lead to an expansion in the black-footed ferret's use of space. Conata Basin has experienced wet and drought years during the past 13 years. Home range estimates for female ferrets ranged from 51.2 – 162 acres during drought (Jachowski 2007). Male home range sizes were considerably larger, averaging 304 acres. Home range size increased as the mean prairie dog utilization distribution value within the home range decreased (Jachowski 2007).

Livieri (2007a, c) estimated the mean home range of female black-footed ferrets on one colony in Conata Basin at 79.3 acres during wet years (1998-2000) and 144.3 acres during a drought (2005), an 82% increase in home range size. Across Conata Basin the mean home range of 11 females was 193 acres and 351.2 acres for the wet and dry years, respectively. Black-footed ferrets are territorial but there is approximately 15% overlap of female home ranges and typically 21.1% of a colony may not be occupied by black-footed ferrets, a difference of 6% over home range overlap. Livieri (2007c) excluded home range overlap area and included unused areas of a colony. That is, it is too simplistic to restrict black-footed ferret spatial requirements to the specific acres of a home range. A 500-acre prairie dog colony containing three black-footed ferrets is not likely to use every acre of that colony but as indicated by Jachowski (2007) over the long and short term the black-footed ferrets will use most of the area. Males may overlap two female black-footed ferret home ranges. Therefore, and mindful of the pitfalls of using home range values, 125 breeding black-footed ferrets (95% population persistence) would require about 16,878 to 31,189 acres of prairie dog colonies during wet and dry years, respectively.

The findings on black-footed ferret home range between wet and dry years at Conata Basin are preliminary, but in general are consistent with patterns observed in many other species, where home ranges are not constant from year to year but typically become smaller when resources are plentiful. However, larger-than-average home ranges do not alone confirm that insufficient habitat quality or amount alone may be the driving factor (Bolen and Robinson 2003, Ewer 1973, MacDonald 1983). Other factors such as density of conspecifics, body size, competitors, predators and landforms can influence home range size. In essence, home range is an expression of the optimization process an animal uses to select habitat, a process that involves many factors (Morrison et al. 2006). Providing an area to allow a species to expand their home ranges to meet their resource needs under varying environmental conditions is fundamental to the long-term presence of the species and fundamental to reserve design (Temple and Cary 2002). The black-footed ferret can only achieve a viable population on substantial acreages of prairie dog habitat. If small areas of habitat were sufficient to sustain black-footed ferrets then one would expect to encounter black-footed ferrets throughout the Great Plains on remaining small patches of habitat, however, such encounters are non-existent.

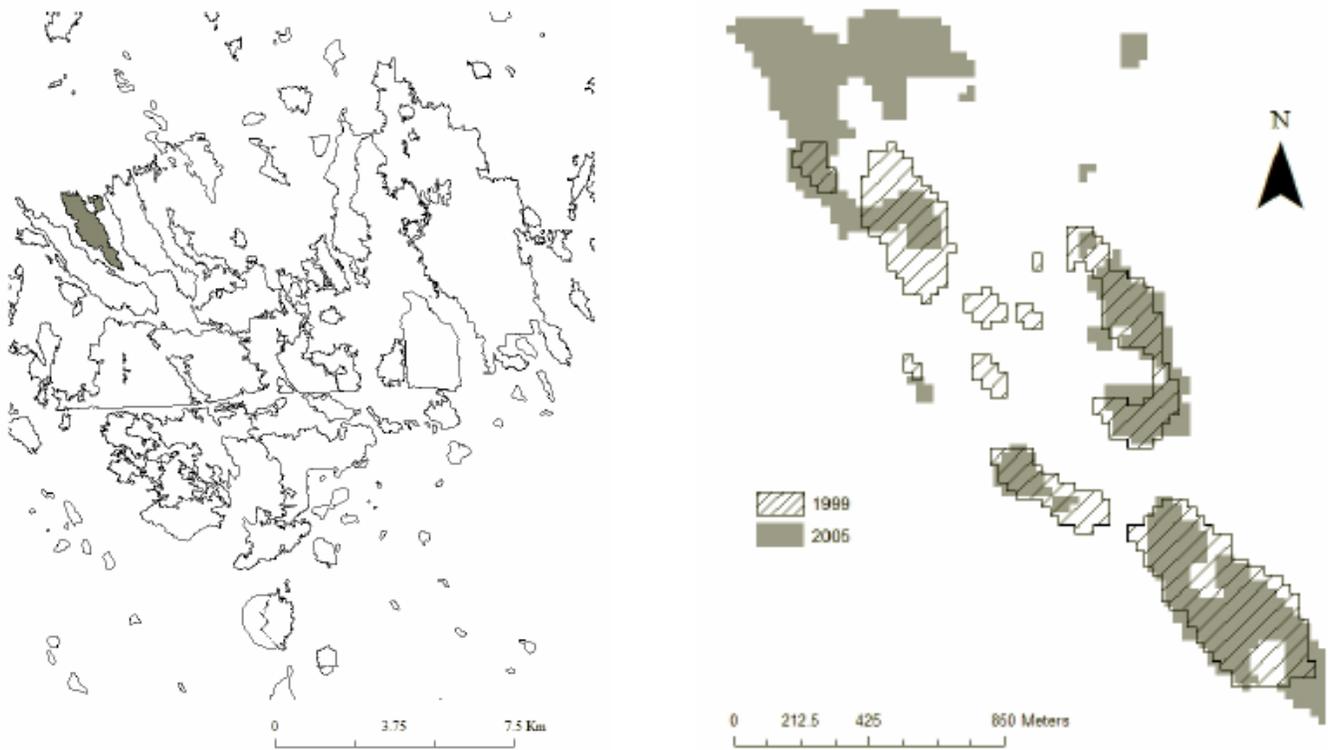


Figure 8. Perimeter map of black-tailed prairie dog colonies in Conata Basin (left). At right are the locations of areas of high densities of active burrows in 1999 and 2005 as illustrated by Jachowski (2007) who classified active burrow distributions into a five-level, ordered factor based upon the quantiles of the prairie dog burrow utilization distribution (Type 1 contained the lowest density patches of prairie dog burrows).

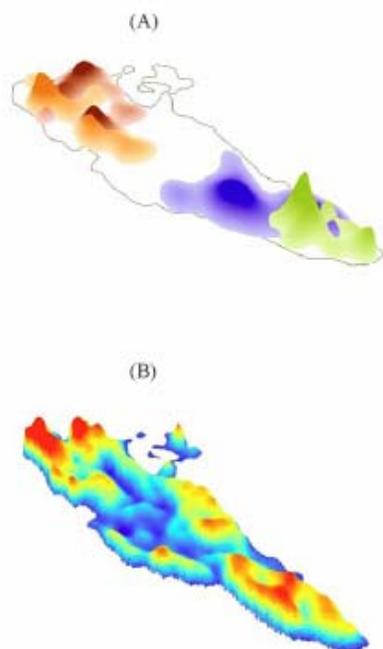


Figure 9. Three-dimensional representation (A) of four female black-footed ferret utilization distributions in 2006 at one large prairie dog colony in Conata Basin, 2006; and the prairie dog utilization (B) at the same colony. The yellow to red colors indicate high densities of active prairie dog burrows (Jachowski 2007). Blue shades indicate low densities or no prairie dogs.

## 7. Black-footed Ferret ESA Status and Other Organizational Rankings

### ESA Status

Endangered

### Conservation Status (<http://www.natureserve.org>)

G1, N1; Nebraska – SH; South Dakota - S1

The black-footed ferret is the 11<sup>th</sup> rarest mammal in the world and was listed in 1967 as endangered under the Endangered Species Preservation Act of 1966, and in 1970 under the Endangered Species Conservation Act of 1969, forerunners to the ESA of 1973. Critical habitat has not been designated for the species. Nebraska and South Dakota list the species as endangered. The black-footed ferret is listed as extirpated by Canada and a recovery team and plan have been established (<http://www.cosewic.gc.ca>; <http://www.speciesatrisk.gc.ca>). Mexico has cooperated with the re-introduction of black-footed ferrets into northern Mexico. The U.S. Government is a state member of the International Union for the Conservation of Nature and Natural Resources (IUCN) which lists the black-footed ferret as extinct in the wild (IUCN 1996; <http://www.iucn.org/themes/ssc/redlist2006/redlist2006.htm>).

The Canada/Mexico/United States Trilateral Committee for Wildlife and Ecosystem Conservation and Management has identified the black-tailed prairie dog as a species of common conservation because of its role in the conservation of the black-footed ferret. The North American Agreement on Environmental Cooperation calls for action to encourage conservation of wildlife and wildlife habitat, and specifically, the protection of species in danger of extinction (<http://www.cec.org>). The black-footed ferret is listed on Appendix I of the Convention on International Trade in Endangered Species of Wild Fauna and Flora (CITES). Appendix I lists species that are the most endangered among CITES-listed animals and plants. They are threatened with extinction and CITES prohibits international trade in specimens of these species except when the purpose of the import is not commercial, for instance for scientific research.

## 8. Recovery and Conservation Planning

Although the black-footed ferret has been listed for nearly 40 years, recovery is still problematic given severe habitat limitations and disease (Conservation Breeding Specialist Group 2004). Challenges facing black-footed ferret conservation include unknown influence of low genetic diversity; canine distemper hazard; direct effects of plague on prairie dog populations and on black-footed ferrets; and minimal suitable habitat for reintroduction (Seal et al. 1989, Roelle et al. 2006). Consequently, reintroductions at most black-footed ferret recovery sites have not succeeded in establishing self-sustaining populations (Figure 10).

Recovery plans developed under Section 4 of the ESA delineate actions which the best available science indicates are required to protect and recover listed species. They provide guidance to state and federal agencies and others. The goal of the black-footed ferret recovery plan is to recover the endangered black-footed ferret to the point where reclassification to threatened status, and eventually delisting, are possible. The recovery priority number for the black-footed ferret is 2 on a scale of 1-18, with 1 equaling the highest priority. This indicates that the black-footed ferret faces a high degree of threat. The ranking also reflects the black-footed ferret's taxonomic status as a full species.

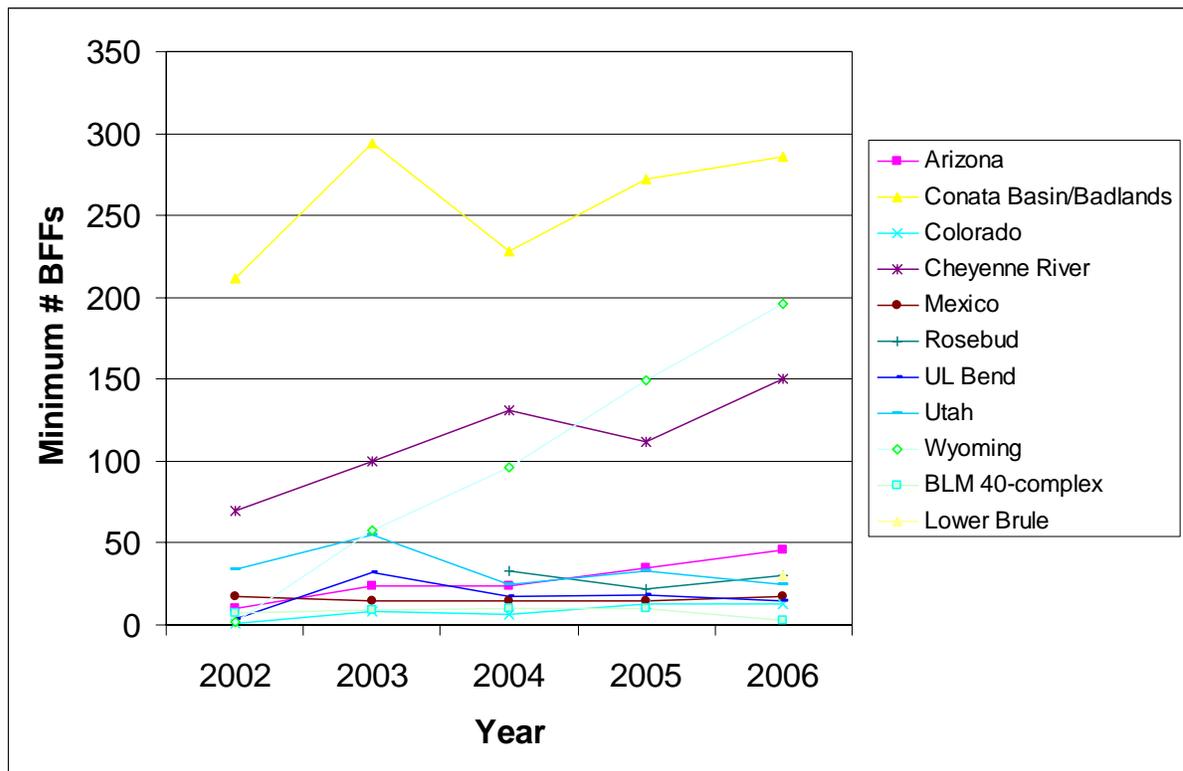


Figure 10. Success of black-footed ferret reintroduction sites through 2006.

To reclassify the black-footed ferret from endangered to threatened status, recovery criteria have been established in the 1988 recovery plan (USFWS 1988).

- ◆ Maintain a core breeding population of a minimum of 240 adults (90 males, 150 females)
- ◆ Establish a pre-breeding census population of 1,500 free-ranging black-footed ferret breeding adults in 10 or more populations with no fewer than 30 breeding adults in any population by 2010; and

- ◆ Encourage the widest possible distribution of reintroduced black-footed ferret populations.

USFWS recommends the following actions:

1. Maintain a captive black-footed ferret population of optimal size and structure to support genetic management and reintroduction efforts.
2. Reduce disease-related threats in wild populations of black-footed ferrets and associated species.
3. Ensure sufficient habitat to support a wide distribution of self-sustaining black-footed ferret populations.
4. Establish free-ranging populations of black-footed ferrets to meet reclassification and delisting goals.
5. Promote partner involvement and adaptive management through regular programmatic review and outreach.

The recovery plan is being revised and recovery of the black-footed ferret may require a pre-breeding census population of 3,000 free-ranging black-footed ferret breeding adults in 30 or more populations with no fewer than 30 breeding adults in any population and at least 10 populations with 100 or more breeding adults (Scott Larson, USFWS, personal communication).

Section 7(a) (1) of ESA requires all federal agencies to further the purposes of the ESA by carrying out programs for the conservation of endangered species.

The Forest Service, National Park Service, and their partners host the most successful black-footed ferret reintroduction efforts at Conata Basin/Badlands (Figure 10). Conata Basin has been a donor site for translocation of wild kits to other reintroduction projects on tribal lands in South Dakota, and in three other states.

The Forest Service has also identified two other potential black-footed ferret reintroduction sites on the Little Missouri National Grassland, North Dakota (Horse Creek) and on the Thunder Basin National Grassland, Wyoming (Cheyenne River). Implementation of reintroduction projects at these sites would improve recovery, but planning and implementation efforts are incomplete. Moreover, the Forest Service manages the widest extent of federal land ownership across Great Plains states east of the plague line which were historically occupied by black-footed ferrets and could establish recovery projects elsewhere in North Dakota, South Dakota, Colorado, Kansas, Oklahoma, Texas and New Mexico.

The Forest Service has the capability of more rapidly and effectively improving the number and distribution of wild black-footed ferret populations than any other federal agency, and more recovery attention is needed on all national grasslands with potential to support black-footed ferret populations (Mike Lockhart, USFWS, personal communication, 2007). Even with development of more recovery sites across national grasslands, the amount of habitat that would be needed to manage prairie dog and black-footed ferret populations would represent a small fraction of the 3.8 million acres of national grasslands.

In addition to agency efforts to identify black-footed ferret reintroduction sites, others have used a geographic information system (GIS) to identify areas with restoration potential within the former range of the black-tailed prairie dog, a species for which there are existing models of habitat suitability (Proctor, 1998, Proctor et al. 2006). Focal areas for black-footed ferrets have been identified, including Fort Pierre, Buffalo Gap (in addition to Conata Basin), and Oglala national grasslands (Figure 11).

Recently, the USFWS and the Forest Service met to discuss black-footed ferret recovery opportunities on national grasslands across the central plains. In a letter dated March 16, 2007, the USFWS stated:

Previous Forest Service commitments to ferret recovery and its success at Conata Basin significantly advanced the overall ferret recovery. No other reintroduction effort has yet achieved the level of success of Conata Basin, and given current habitat extent and

presence of plague over much of the western prairie habitats, there are few sites that can realistically hope to match that success today. Moreover, Conata Basin is an essential donor source for wild born ferret kits and has helped initiate and/or bolster reintroduction efforts at three other sites in South Dakota, as well as recovery sites in three other States. The Forest Service's contribution to ferret recovery is enormous and deserving of recognition and appreciation of the Service and all of the other partners involved in ferret recovery

However, despite considerable recovery progress since the mid-1980s, the ferret remains a critically endangered species. Potential recovery sites across North America are limited. Perhaps no other agency or entity can contribute to more assured and rapid recovery of the ferret than the Forest Service. Ferret recovery cannot be achieved on National Grasslands alone, but likewise, the establishment of adequate numbers of ferret populations across the historical range of the species (a recovery plan objective) may not be possible without concerted support by the Forest Service and expansion of field recovery efforts across more of the Forest Service's vast western holdings. Even with more focused Forest Service management and development of additional sites for prairie wildlife, the amount of managed land actually required to meet these needs would represent a small percentage of the almost 4 million acres of National Grasslands.

Forest Service manages more lands distributed over plague-free, high-value black-tailed prairie dog habitat than any other State or Federal agency (i.e., in North and South Dakota). There may be excellent opportunities to expand reintroduction efforts into portions of other National Grasslands in Colorado, Kansas, Oklahoma, New Mexico, Texas, and Wyoming. I have enclosed a draft "recovery site" evaluation table (*see at end of ferret section*) currently being finalized by the Conservation Subcommittee of the Black-footed Ferret Recovery Implementation Team. That document is designed to identify existing and potential recovery areas; it shows that National Grasslands throughout the Plains States represent the greatest long-range potential to expand ferret recovery. We understand that many Forest Service sites are highly fragmented in Forest Service/private ownership and that plague is present at many as well. It may or may not be feasible in some cases to consolidate land ownership in a similar manner as was accomplished at Conata Basin. Still, the Service and other recovery partners are pursuing establishment of smaller "nursery" populations and development of plague management capabilities with a view that successful distribution of small sites in the near term may be crucial to spreading population extinction risk, while simultaneously increasing numbers and distribution in the wild. Moreover, we are working on some innovative administrative procedures (via Endangered Species Act section 10(A)(1)(a) whereby we can implement experimental recovery actions in a more timely and cost effective manner (as opposed to requiring a costly and protracted 10(j) rulemaking process, while still ensuring the protections afforded by 10(j) on adjacent landowners). Our hope is that we can get more small populations started and maintained, while simultaneous efforts can be made to manage for expanding habitat base and consolidating land ownership.

It is important then that the Forest Service examine all possible opportunities to expand recovery efforts across grassland units. While habitats the size and quality of Conata Basin are certainly desirable, we understand that such habitat development is not possible for most units, and that actual development of suitable prairie dog habitats for ferrets may take considerable time to accomplish. Nevertheless, the planning for recovery sites/grassland unit should be initiated as soon as feasible. Wherever possible (especially on plague free units), small nursery populations of ferrets should be established on

suitable complexes where 1,500 to 3,000 acres of prairie dogs occur and can be maintained and managed for ferret recovery.

Finally, it is important to understand where the recovery program stands with respect to implementation and monitoring of ferret projects. Recovery partners are now trying to establish as many populations as possible as broadly as possible across the ferret's historical range. Funding for monitoring and research that was typically involved in earlier recovery efforts is largely unavailable today. Consequently, we are looking at more "hands-off" approaches to recovery that entail only basic monitoring requirements (e.g., one fall spotlight effort/year for the first 3 years, perhaps only every 3- 5 years thereafter). We know that the cost of implementing recovery actions is becoming a greater problem for partners and it has become more important from a recovery perspective to establish more and better distributed populations.

The Forest Service responded to its staff:

To date, the most successful black-footed ferret reintroduction site is in Conata Basin on the Buffalo Gap National Grassland in South Dakota. A population of about 100 breeding adults has been established, and wild born kits from that site have even been available to supplement other recovery sites. As part of the Northern Great Plains Land Management Plan revision process, the Forest Service designated future black-footed ferret reintroduction sites at another area on Buffalo Gap, and on portions of the Little Missouri and Thunder Basin National Grasslands. These areas are envisioned to contain large acreages of black-tailed prairie dog colonies to support ferret populations.

Additional reintroduction sites are needed to make better progress toward recovery objectives. Finding or even establishing large complexes of black-tailed prairie dog colonies is often problematic. Most black-tailed prairie dog colonies are small and scattered across the landscape. FWS is indicating that smaller complexes of black-tailed prairie dog colonies (1,500 to 3,000 acres) can play an increasingly important role in national recovery by supporting small nursery populations of black-footed ferrets. The enclosed letter from FWS provides the rationale for such nursery populations.

FWS is also working on developing some innovative administrative procedures that may help to expedite reintroduction actions. Recognizing that funding for monitoring may be limited; FWS is also reducing expectations for post-release monitoring.

Despite our important contributions to the national recovery program to this point, recovery of the black-footed ferret still remains tenuous at best. Opportunities likely remain for the Forest Service to continue to be a leader in the national recovery effort. Please join in the conversations and meetings as you can to better understand and discuss these opportunities in Region 2.

Limitations on recovery program efforts caused by declining budgets and redirected priorities over many areas are a critical concern. The lack of response of federal and state agencies to ESA Section 7(a) (1) responsibilities is a critical concern. The March, 2007 letter from the USFWS to the Forest Service not only highlights again the importance of Conata Basin but beckons the Forest Service to establish more black-footed ferret reintroduction sites on national grasslands. More detailed information on the status and recovery needs of black-footed ferret is presented by the U.S. Geological Survey (Roelle et al. 2006).



## 9. Analysis of Effects: Black-footed Ferret

This section analyzes the direct effects, indirect effects, and cumulative effects of the proposed action on the black-footed ferret. It describes and analyzes the effects of the action that would directly affect the species. For example, actions that would immediately remove or destroy habitat or displace animals are considered direct effects. Direct effects are effects occurring during implementation of an action.

Indirect effects are those that are caused by the action and are later in time (after the action is completed) but still reasonably certain to occur. Examples include changes to ecological systems, such as predator/prey relationships, long-term habitat changes, or anticipated changes in human activities, including changes in land use. The geographic extent of the above effects is the action area, defined as all areas to be affected directly or indirectly by the federal action and not merely the immediate area involved in the action.

The effects of the action must be evaluated in the context of cumulative effects. These are defined in the ESA as the effects of future state, tribal, local, or private activities that are reasonably certain to occur in the foreseeable future within the action area. NEPA requires disclosure of *cumulative impacts*, which may result from the incremental impact of the action when added to other past, present, and reasonably foreseeable future actions regardless of what agency (federal or non-federal) or person undertakes such other actions. Cumulative impacts can result from individually minor but collectively significant actions taking place over a period of time. The larger the extent of the project action area, the more extensive this aspect becomes. Once identified, the cumulative effects are evaluated with the direct and indirect effects of the action, to provide the context for the Services' jeopardy/adverse determination.

### Direct Effects

#### *Application of Rodenticide*

Rodenticide is proposed in the IMZ under all of the action alternatives. Under current direction (the no action alternative, Alternative 2 in the 2008 FEIS), rodenticide use occurs in the BMZ per the 2005 forest plan amendment 2 and in limited circumstances in the IMZ regarding threats to public health and infrastructure. Prairie dog rodenticide (2 percent zinc phosphide bait) is highly effective in reducing prairie dog populations, the sole prey of the black-footed ferret. The direct effect of poisoning is the elimination of the prey and habitat currently used by the black-footed ferret and the elimination of habitat where black-footed ferrets can disperse to.

Poisoning of non-target species can occur but is minimized if the rodenticide is applied according to label specifications, time of year, and favorable weather. In Conata Basin, measurable reductions in non-target populations were observed for deer mice (*Peromyscus maniculatus*), ants (Hymenoptera), and darkling beetles (Coleoptera), but there was no measurable reduction in avian and other invertebrate populations (Apa et al. 1991, Deisch et al. 1990, Uresk et al. 1985a, Uresk et al. 1985b, and Uresk et al. 1986).

Zinc phosphide is a heavy, finely ground gray black powder that is practically insoluble in water. When exposed to moisture it decomposes slowly and releases phosphine gas. Phosphine generates rapidly when in contact with dilute acids in the stomach causing death. Animals that ingest lethal amounts of bait usually are asphyxiated within 3-5 hours (Timm 1983). Zinc phosphide is also a strong emetic (causes vomiting) which can determine how much zinc phosphide is required to kill the animal and whether or not an animal dies after ingesting the chemical (Schitoskey 1975).

Grain bait (oats) is treated with zinc phosphide for consumption by prairie dogs. Untreated grain (pre-bait) is typically applied to prairie dog colonies a few days prior to zinc phosphide treated bait to

promote grain consumption. Prairie dogs usually will not eat the grain bait until early in the fall when their natural forage matures and dries (South Dakota Department of Agriculture et al. 1994). When proper procedures are followed, efficacy of zinc phosphide bait is typically 90 percent or higher (South Dakota Department of Agriculture et al. 1994).

Zinc phosphide is highly toxic to wild birds. It is also toxic to non-target mammals. Nearly sixty studies have been conducted on the toxicity of this rodenticide to wild animals. The most sensitive bird species which have been evaluated are geese (LD50 of 7.5 mg/kg for the white-fronted goose). Pheasants, morning doves, quail, mallard ducks and the horned lark are also very susceptible to zinc phosphide. Blackbirds are less sensitive (Extension Toxicology Network 1993). Seed-eating animals on national grasslands will be at risk from zinc phosphide treated oats. The whooping crane is granivorous.

Zinc phosphide breaks down rapidly in the digestive tract of affected animals, so predators and scavengers are generally not exposed to the compound. Species that were fed zinc phosphide-poisoned prey during laboratory studies and showed no negative physiological symptoms included Siberian polecats, mongooses, coyotes, kit foxes, mink, black vultures, bald eagles, golden eagle, and great-horned owls (USDA Animal Plant and Health Inspection Service 1994).

Zinc phosphide is not stored in the muscle or other tissue of poisoned animals and therefore there is no true secondary poisoning. However, zinc phosphide remains toxic for several days in the gut of dead rodents. Other animals can be poisoned if they consume enough of the gut content of rodents recently poisoned by zinc phosphide (Timm 1983), however, such poisoning is limited because most prairie dogs poisoned with zinc phosphide die in their burrows (Tietjen 1976).

There is only a small amount of deterioration of zinc phosphide baits due to the evolution of phosphine gas. Dry baits are considered toxic indefinitely. Lecithin-mineral oil, added to zinc phosphide to adhere to grain bait, offers protection against moisture, and therefore increases its stability. Under field conditions, zinc phosphide baits may remain toxic several months until eroded by weather, decomposition of the carrier, or removal by insects (Timm 1983). Translocation of phosphine gas has been demonstrated but it is rapidly converted to harmless phosphates (Timm 1983).

The 2001 Forest Plan prohibits the use of rodenticides (above-ground baits) outside October 1 to January 31 to reduce risks to migratory birds. To reduce risk to other wildlife, the LRMP does not allow burrow fumigants in prairie dog colonies. The act of applying rodenticide may also directly affect some species. Trucks are used to haul pre-bait and bait over two-track trails to the application site. Once at the site, all-terrain vehicles are operated on prairie dog colonies to allow people applying the oats to reach all prairie dog holes. The use of vehicles may directly impact some species.

Currently, black-footed ferrets occur only in the Conata Basin MA 3.63, with a 2008 observation of ferret snow tracks immediately adjacent to but outside of this area. Over time, some additional dispersal could occur within the vicinity of the reintroduction area. As explained above, no secondary poisoning of black-footed ferrets would be anticipated as a result of rodenticide application in non-MA 3.63 areas.

### ***Minimum and Maximum Prairie Dog Colony Extent***

A major aspect of the proposed action would be a minimum and maximum range of prairie dog colony acres to be allowed in each GA (Table 2). A consideration in the analysis of potential effects on the black-footed ferret is how setting a minimum of 1,000 acres of colonies and a maximum acreage of 3% of gross GA acreage (2,100 acres to 3,600 acres depending of the GA; Table 2), and setting a maximum inter-colony distance of 6 miles, will affect options for future black-footed ferret reintroductions on the six GAs. Will such a range of prairie dog colony acreage and inter-colony spatial standard going provide habitat for a viable population of black-footed ferrets?

One of the reasons for the near extinction of the black-footed ferret is the scant acreage of prairie dog colonies throughout the prairie dog's range in the Great Plains. If black-footed ferrets could survive in an area of low prairie dog colony acreage, the species would still occur in the Great Plains. In the Great Plains, prairie dog colonies typically are small and constitute a small extent of private rangeland. Examples of the low percentage of private land base in prairie dog colonies are displayed on the non-federal land within the GAs (Table 2; Figures 1 and 2). The percentage in prairie dog colony ranges from 0.17% on Fort Pierre GA non-federal land (much of that non-federal land is in cropland) to 4% on Fall River Northeast GA. The percentage of non-federal land in prairie dog colonies remained stable during 2002-2006 despite the drought (Table 2). Total acres fell from 8,977 to 8,005 acres (2.0 - 1.9%).

On federal land in the GAs, the percentage of land base in prairie dog colony ranged from 0.5 - 4.2% in 2002, depending on the GA, and from 0.17 - 1.5% in 2006 after BMZ poisoning (Table 2). Total prairie dog acreage on federal land fell from 8,264 to 6,068 acres (1.7 - 1.2%). Almost all of the current acreage is in the IMZ. The proposed action, Alternative 1, would establish a minimum of 1,000 acres (0.8 - 1.4% of federal land base, depending on the GA size) for black-footed ferret reintroduction. At this minimum acreage, the proposed action establishes a prairie dog colony extent slightly less than what exists on private land within the GAs. The minimum total of 6,000 acres for the six non-MA 3.63 GAs represents 1% of total federal land in the GAs, 1.2% of the total suitable prairie dog habitat, 2.5% of the total IMZ acreage, and 3% of the total suitable prairie dog habitat in the IMZ (Table 2).

Will the proposed action's maximum prairie dog colony acreage per GA (2,100 acres to 3,600 acres depending on the GA (Table 2)) of prairie dog colonies provide viability for a black-footed ferret population? The proposed action cites the recent reintroduction of black-footed ferrets at Wind Cave National Park for the proposed action's contention that the minimum and maximum range of acres of prairie dog colonies per GA will keep options open for black-footed reintroduction. The USFWS and others have long viewed large acreages of prairie dog colonies as essential for recovery of the species (Roelle et al. 2006). Prairie dog colony extent for black-footed ferrets has been a minimum of 10,000 acres, 5,000 acres, and now the USFWS will consider a reintroduction site if the site contains at least 1,500 - 3,000 acres of prairie dog colonies. This consideration by USFWS does not convey that 1,500 - 3,000 acres is all the acres a site needs for black-footed ferret reintroduction. Indeed, these small reintroduction sites are experiments to test whether a population of black-footed ferrets can be established at such acreages. The experiment may fail or the small sites such as Wind Cave and Smokey Valley ranch, Kansas may require periodic supplementation with captive-reared or wild born black-footed ferrets. The USFWS and others believe that it is crucial to attempt to establish and maintain as many black-footed ferret populations as possible in native habitats in order to achieve recovery of this species. The experiments at low prairie dog colony acreage will provide useful information but they are in another sense a sign of desperation because it is very difficult to find large extents of prairie dog colonies for black-footed ferret reintroduction.

The introduction of black-footed ferrets at Wind Cave is a test of the viability of using a reintroduction site with less than 5,000 acres of prairie dog colonies (National Park Service 2006). The black-tailed prairie dog management plan for Wind Cave sets the limits of prairie dog colonies within the park between 1,000 and 3,000 acres (3.6 - 11% of the park's land base or 12 - 35% of its 8,566 acres of suitable prairie dog habitat). The limits were established before a black-footed reintroduction was proposed.

The primary goal of the project is to test and evaluate the viability of ferret populations in a small prairie dog complex (i.e., approximately 2,500 acres). The information obtained from this project will have significant implications to the ferret recovery program. The five-year goal of the proposed reintroduction is to establish a self-sustaining population of black-footed ferrets in the park. Specifically, the population would consist of at least

30 breeding females after five years. In the long-term (*i.e.*, continuing five years and longer after release), the goal would be to have a ferret population with size and distribution in a proportional relationship to the extent of prairie dog complexes in the park.

Again, it is unknown at this time if a complex of 1,000 – 3,000 acres will be able to support a self sustaining population of ferrets, therefore the park will be testing the viability of a site <5,000 acres (National Park Service 2006). Black-footed ferrets were reintroduced into Wind Cave in 2007 when the prairie dog colony acreage was 2,800 acres (greater than the acreage depicted in the Wind Cave map below). A similar reintroduction has recently occurred at the 16,800-acre Smokey Valley ranch in Kansas where there are about 2,600 acres of prairie dog colonies.

In light of the above, especially the misunderstanding about whether low acreages of prairie dog colonies constitute viability for the black-footed ferret, an understanding of prairie dog colony acreages, including the origin of minimum and maximum acre values, and black-footed ferret habitat requirements is therefore important. Such an examination is necessary to substantiate the BA's conclusions about black-footed ferret options under the proposed action.

The biological evaluation (BE) for the proposed action lists the sources for the origin of a minimum of 1,000 acres of prairie dog colonies per GA as the 2001 Forest Plan and the multi-state conservation plan for the black-tailed prairie dog (Luce 1999, 2003). The 2001 Forest Plan defines a prairie dog colony complex as:

a group of at least 10 prairie dog colonies with nearest-neighbor inter-colony distances not exceeding 6 miles with a total colony complex acreage of at least 1,000 acres.

The 2001 Forest Plan is silent on minimum and maximum acres except for two GAs: Fort Pierre and Oglala, where it calls for at least one complex as defined above. At that time prairie dog colony acreage was low in those two GAs and the 2001 Forest Plan desired to encourage prairie dog colony expansion. The 1,000 acres was simply viewed as a target to achieve over the first few years of the 2001 Forest Plan. The ten colonies was taken from a general discussion on meta-populations by Hanski (1997) who suggested that for species in general an adequate successful network of small habitat fragments should have a minimum of 10-15 well-connected fragments. Hanski indicated that even this number may be insufficient if regional stochasticity is strong and local dynamics are strongly correlated. The six-mile maximum for inter-colony distance is taken from Knowles (1985) who published the farthest known dispersal distance for the black-tailed prairie dog.

The multi-state conservation plan for the black-tailed prairie dog calls for:

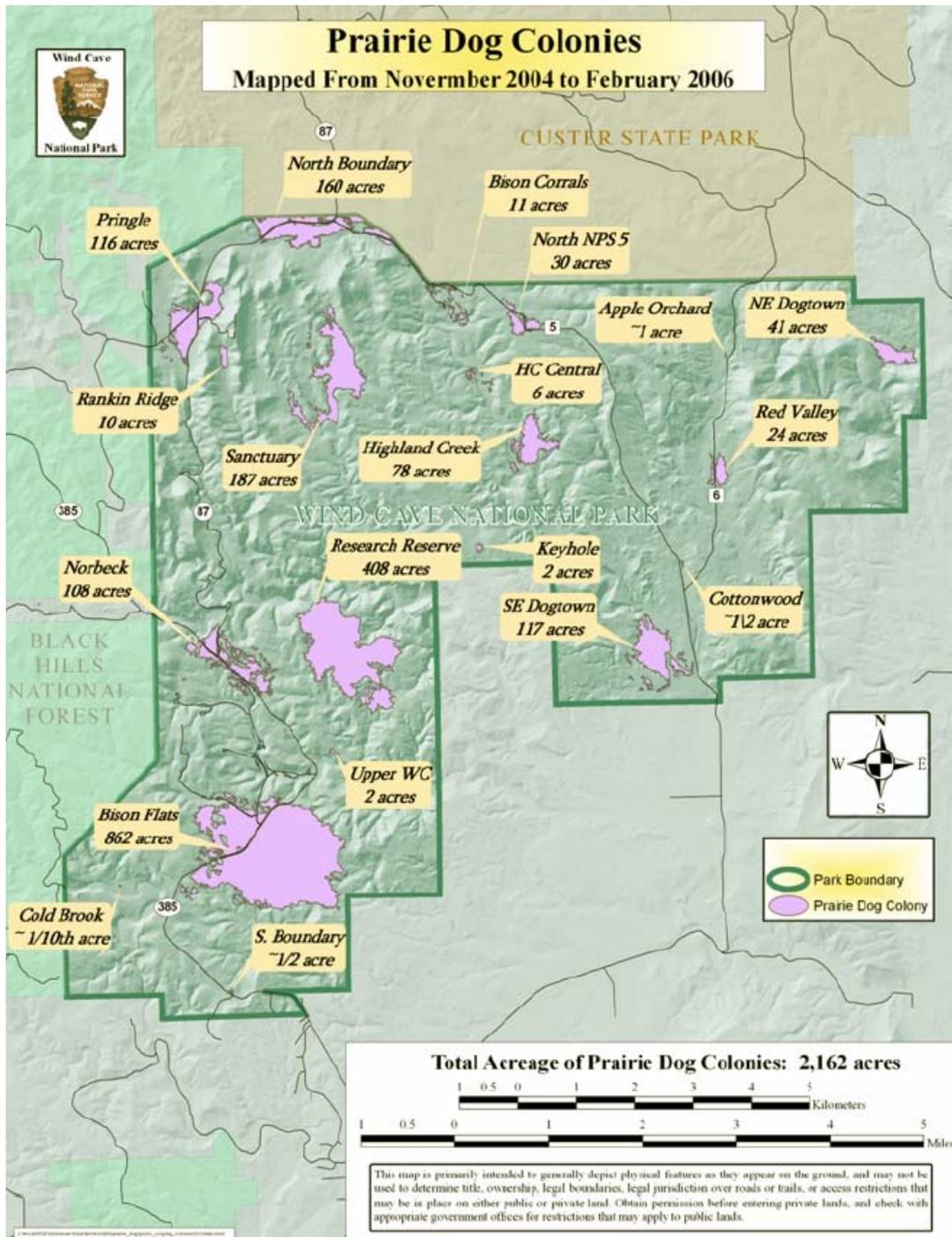
an area greater than 1,000 acres of suitable prairie dog habitat, encompassing either an existing complex of occupied prairie dog colonies or an area where a complex of colonies can be created to sustain a viable population of prairie dogs for long-term management.

However, the multi-state conservation plan for the black-tailed prairie dog provides no substantiation for the 1,000 acres.

The 1,000 acres commonly appears in documents but there is no empirical basis for the figure even in the multi-state conservation plan. Moreover, there has never been any viability assessment to validate the accuracy of 1,000 acres colonies for viability of prairie dogs and associated species. Dr. Dean Biggins (USGS, personal communication, 2008) has been studying several prairie dog species for over 25 years and states clearly that there is no scientific basis for the 1,000 acre figure for prairie dog or black-footed ferret viability. The BE for this proposed action also recognizes the interface of legitimate social-political issues when grappling with prairie dog colony acres. Such issues arguably are significant when setting prairie dog colony acres for prairie dogs, black-footed ferrets, or grassland ecosystems.

The direct effect of establishing minimum and maximum acreages of prairie dog colonies does not preserve options for future black-footed ferret reintroductions. The proposed action is clear that current GAs with prairie dog colony acreage below 1,000 acres will be increased to 1,000 but is silent on any aggressive action to go beyond 1,000 acres. However, the desire to enhance options for black-footed ferret reintroductions into non-MA 3.63 is stated, suggesting that the maximum acreages will be pursued.

In addition to colony acreage triggers for prairie dog control in the IMZ, desired vegetation conditions in colonies must be met. If such conditions are not met, colonies can be poisoned to reduce prairie dog densities, or, if acreage exceeds the 3%, the acreage can be poisoned back to 2%. Such actions would negatively impact black-footed ferrets, however, the proposed adaptive response protocol is designed to give black-footed ferrets priority. For example, if a 50-acre prairie dog colony does not meet similarity index value and if black-footed ferrets were present, the prairie dogs would not be controlled out of deference to the black-footed ferrets. If the overall maximum acreage were exceeded, other arrangements would have to be made to translocate black-footed ferrets.



**Distribution of prairie dog colonies at Wind Cave National Park in 2006. By 2007, the time of black-footed ferret reintroduction, the colony acreage had increased to 2,800 acres.**

Table 5. The extent of prairie dog colonies under each alternative and potential black-footed ferret outcomes from establishing minimum and maximum acres of prairie dog colonies (black-footed ferret habitat) in non-3.63 Geographic Areas. Black-footed ferrets are not known to currently exist in the non-3.63 Geographic Areas, although three sets of snow tracks were observed in January, 2008 Wall Southeast.

Geographic Area	Total Acres in Geographic Area	Total Acres in IMZ	Total Suitable Acres (IMZ)	Current Prairie Dog Colony Acreage	Range of Prairie Dog Colonies by Alternative					Expected Black-footed Ferret Population <sup>1</sup>				
					ALT 1	ALT 2	ALT 3	ALT 4	ALT 5	ALT 1 <sup>2</sup>	ALT 2 <sup>3</sup>	ALT 3 <sup>4</sup>	ALT 4 <sup>5</sup>	ALT 5
Oglala	94,484	29,582	18,646	1,125	1,000 - 2,800	1,125 - 13,097	100 - 900	N/A	9,500 - 18,700	11	-	0	0	73
Fall River Northeast	91,298	39,964	33,478	1,130	1,000 - 2,800	1,130 - 13,155	0 - 2,700	N/A	9,100 - 18,300	11	-	0	0	68
Fall River West	119,951	49,863	48,420	210	1,000 - 3,600	210 - 2,445	0 - 3,600	N/A	12,000 - 24,000	11	-	0	0	95
Wall North	69,437	16,019	14,801	454	1,000 - 2,100	454 - 5,285	0 - 2,100	N/A	6,900 - 13,900	11	-	0	0	50
Wall Southeast	90,840	32,840	27,885	1,414	1,000 - 2,900	1,414 - 16,461	0 - 2,800	N/A	9,100 - 18,200	11	-	0	0	68
Fort Pierre	116,053	65,414	61,214	1,735	1,000 - 3,500	1,735 - 20,198	0 - 3,470	N/A	11,600 - 23,200	11	-	0	0	82

<sup>1</sup>Based upon the area requirements of black-footed ferrets (Table 7)

<sup>2</sup>Alternative 1 does not provide enough acres of prairie dog colonies to sustain a population of black-footed ferrets. The GAs areas may provide enough acres at the maximum range for a few black-footed ferrets. This population may have to be supplemented periodically with wild born or captive reared black-footed ferrets. However, the low end of one thousand acres is not adequate for even a nursery population of black-footed ferrets.

<sup>3</sup>Alternative 2 provides substantial maximum acreages that could sustain a black-footed ferret population in several geographic areas. However, the minimum acreage (the current 2006 acreage) is too low to support a population of black-footed ferrets. The problem in evaluating Alternative 2 ranges of prairie dog colony acres stems from their presentation in the FEIS. Alternative 2 is the current Forest Plan which allows prairie dog colony expansion and contraction to occur in the IMZ and does not specify colony acreages. The current plan also allows tools to expand colonies. The above range of acreage simply presents the current acreage and the expected acreage in ten years. In reality, Alternative 2 will support black-footed ferrets if the minimum acreages are adjusted upwards.

<sup>4</sup>Alternative 3 allows zero acres of prairie dog colonies and hence cannot support a population of black-footed ferrets.

<sup>5</sup>Alternative 4 provides no prairie dog colony acres in the non-MA 3.63

## Indirect Effects

An indirect effect is the loss of habitat and ecosystem processes as a result of rodenticide use and reductions in prairie dog populations in black-footed ferret habitat. Prairie dogs cut tall vegetation creating the low structure grassland typical of black-footed ferret habitat. Permanently removing prairie dog populations could result in a shift of the vegetative community from a buffalo grass/blue grama sod to a western wheatgrass/green needlegrass community (depending on the soil type at the prairie dog colony). This in turn could alter habitat suitability for wildlife species. Many wildlife species prefer low structure grassland and bare ground created by prairie dogs. Soulé et al. (2005, 2007) labeled the prairie dog as a “strongly interacting species” in an ecosystem. They viewed as essential prairie dog population densities or levels that maintain interaction effectiveness rather than mere persistence at minimal numbers.

The BE and this BA also address the role of the prairie dog in the prairie ecosystem of the black-footed ferret. Both documents cite numerous sources for the role of the prairie dog as a keystone or highly interactive species, a species whose ecologically effective densities are far greater than densities required for mere population persistence. Ecologically effective densities influence large areas of grassland, thereby creating black-footed ferret habitat. Miller et al. (2007) summarize the unique role played by prairie dogs in black-footed ferret habitat. Like cattle, prairie dogs graze, but prairie dogs represent much more than just grazers (Kotliar et al. 2006). They also move soil (Detling and Whicker 1988, Detling 1998), influence nutrient cycling (Coppock et al. 1983a; Detling and Whicker 1988; Whicker and Detling 1988, 1993; Detling 1998), increase nitrogen content of soil and plants (Holland and Detling 1990, Detling 1998), change vegetation structure and community dynamics (Coppock et al. 1983a, Whicker and Detling 1988, Weltzin et al. 1997, Detling 1998, Fahnestock and Detling 2002), aerate the ground (Whicker and Detling 1988, Outwater 1996, Detling 1998), alter soil chemistry (Munn 1993), and deepen water penetration (Outwater 1996, Detling 1998). They provide a ready source of prey to many predators such as the black-footed ferret and burrows for shelter to other animals and insects, including the black-footed ferret (Goodrich and Buskirk 1998, Kotliar et al. 1999, Shipley and Reading 2006). This combination of effects gives the prairie dog its role as a highly interactive (keystone) species in the ecosystem, creating a matrix of different habitats that increases diversity across the grassland and sustains black-footed ferrets (Kotliar et al. 1999, 2006; Miller et al. 2000). (see *Prairie Dog as a Keystone Species in Creating Habitat for the Black-footed Ferret and other Species*).

The minimum and maximum prairie dog colony acreage caps would impose a steep limitation to the prairie dog’s ecological role in a grassland ecosystem. Only 1.2% of the total suitable prairie dog habitat would be exposed to the ecological role of the prairie dog. In contrast, almost all of the non-MA 3.63 is annually exposed to the important ecological processes of grazing and browsing by domestic and native ungulates. The BE refers to strong regulatory mandates for viable populations and also references concerns that ecologically effective densities of highly interactive species like the prairie dog are not mandated (Soulé et al. 2005).

## Cumulative Effects

Drought, rodenticide use, livestock grazing, plague, prairie dog shooting, and land use conversions, are reasonably foreseeable to occur in the project area.

## ***Drought and Black-footed Ferret Habitat***

Drought is significant in the natural history of black-footed ferret habitat. Periods of low precipitation can occur annually intermixed with normal and above normal precipitation periods. However, an extended period of low precipitation (drought) has occurred during the last 6 years, resulting in reduced plant productivity and expansion of prairie dog colonies, albeit, with low densities of prairie dogs. Drought also reduces prairie dog productivity. Climate history indicates that drought may persist into the future, however, periods of drought are normal. Alternatives that favor increasing prairie dog colony acreages during drought benefit the black-footed ferret because black-footed ferret home ranges must increase significantly during drought to compensate for reduced densities of prairie dogs. Alternatives that decrease prairie dog acreages during drought reduce food availability to black-footed ferrets.

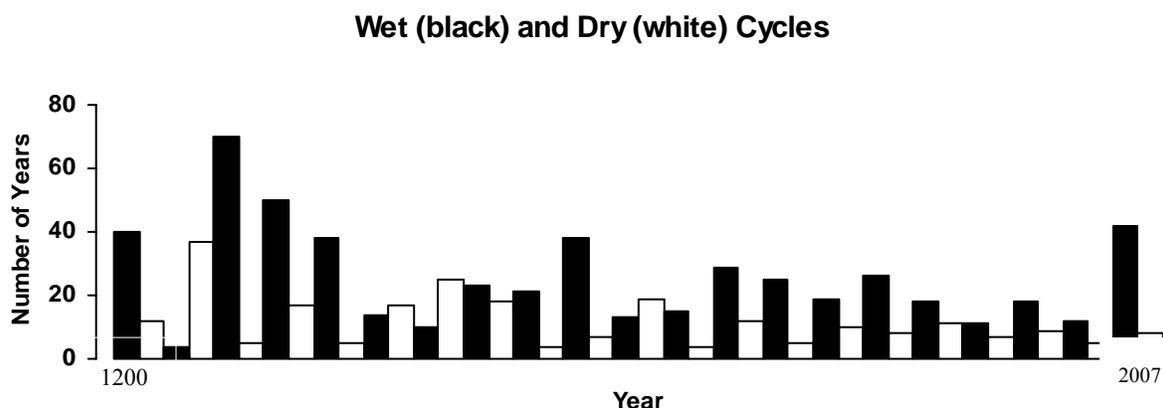


Figure 12. Tree ring data show repeated droughts of five years or longer in western Nebraska during 1200 - 2007 ((Source: University of Nebraska).

Drought is one of the stimulants for this FEIS. However, drought must be put into ecological and historical context. Drought promotes short-statured vegetation and prairie dog colonies can expand rapidly establishing large colonies with low densities of prairie dogs as plant food sources decline under drought. Drought has played a major role in the expansion of black-tailed prairie dog colonies on the NNF, including in Conata Basin, and throughout the Great Plains. However, droughts occur regularly and prairie dog colonies and black-footed ferret populations persist even continuously for hundreds or thousands of years in the same location despite drought or wet periods (White 1986, Carlson and White 1987). The 42-year wet cycle (1957-1999) was unusually long and it has been the period during which most living Americans have lived. Extended drought is a new experience for most people. But again, the phenomenon is normal during the past 800 years (Figure 12). What we see today in Conata Basin black-footed ferret habitat and elsewhere in regards to bare ground has been seen before on prairie dog colonies during drought as noted by Smith (1958):

As the drought continued in 1956, some areas of the dog town were bare and other areas were fifteen per cent covered by vegetation whereas outside the dog town the vegetation covered twenty-five per cent of the ground. The pond dried up in June and was completely dry for eight months. In the autumn the cattle were taken off the area. Severe dust storms occurred; wind eroded areas around the pond and windmill. In December of 1956, and in January, February, and March of 1957, the prairie dogs dug for roots until parts of the dog town looked as though they had been cultivated. April and May brought ten inches of rain, which came in slow drizzles, soaking in with little or no runoff.

Annuals, especially peppergrass, began to green up the area and form a dense ground cover.

If the non-MA 3.63 areas are allowed to approach maximum prairie dog colony acres and a black-footed ferret nursery population is established, and if drought expands the colonies to beyond the maximum acreage, or if similarity index is violated, the proposed action’s implementation plan would have to address the expanded home range needs of black-footed ferrets through translocation or some other tool. The adaptive response protocol appears to recognize this and addresses black-footed ferrets in MA and non-MA 3.63.

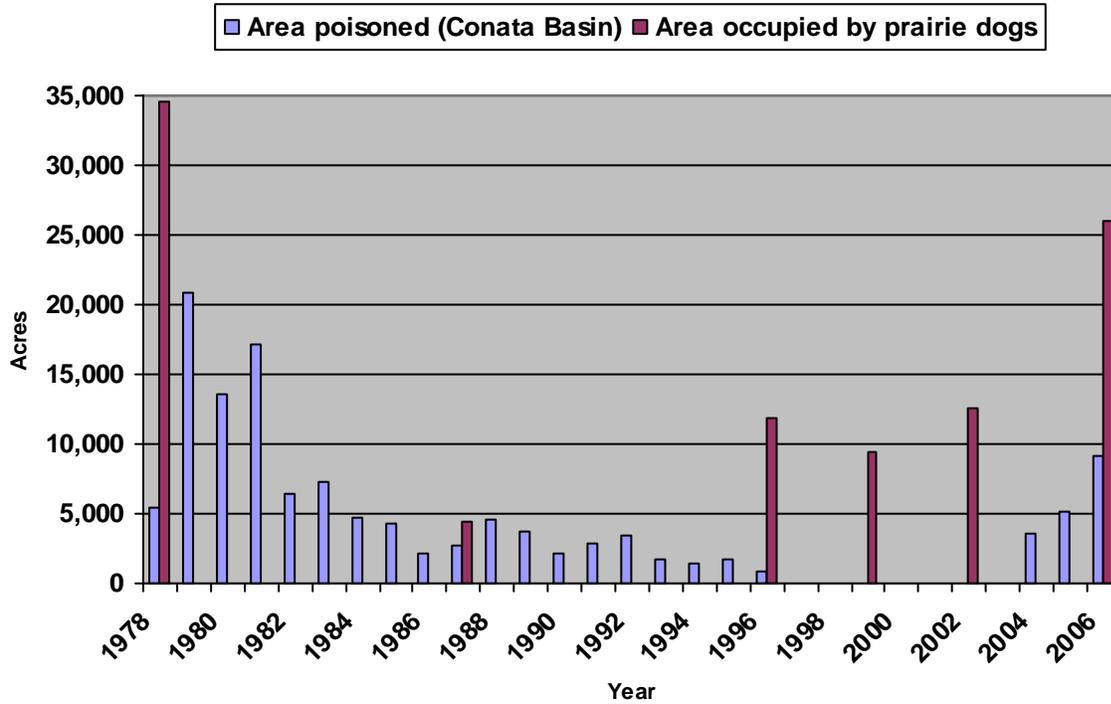


Figure 13. Extent of black-tailed prairie dog colonies and extent of poisoning in Conata Basin, Buffalo Gap National Grassland, South Dakota.

### ***Poisoning of Black-footed Ferret Habitat***

The cumulative effect of former and current poisoning prairie dog colonies in the Great Plains has been to eliminate almost all black-footed ferret habitat. Substantial acreages of prairie dog colonies occur on some federal lands and tribal lands where poverty and regulation curtail prairie dog poisoning (Figure 3). The century-old poisoning of black-footed ferret habitat persists because prairie dogs consume the grasses and forbs consumed by domestic livestock. Therefore, private landowners and agencies (Figure 13) have directed extensive programs to eliminate prairie dogs. The general dislike of black-footed ferret habitat persists today (Chace 1973, O’Meilia et al. 1982, Fox-Parrish 2002, Reading et al. 1999, 2002, 2005, McCain et al. 2002, Lamb et al. 2006, Detling et al. 2006, Miller and Reading 2006, Hoogland 2006). Although scientists have identified the prairie dog, the species upon which the black-footed ferret completely depends, as a keystone species, much of the

rural public in the Great Plains views the prairie dog as a pest whether on private or federal land (Carr 1973, Lee and Henderson 1989, Reading and Kellert 1993, Kayser 1998, Zinn and Andelt 1999, Lamb et al. 2001, Sexton et al. 2001, Wyoming Agricultural Statistics Service 2001, Fox-Parrish 2002). A minority of ranchers favors the retention of small to medium-sized prairie dog colonies on federal lands (Carr 1973, Reading 1993).

Forrest and Luchsinger (2006) summarize the extent of destruction of black-footed ferret habitat since European settlement (Figure 14). By 1911, prairie dogs were gone from 2 million acres in Kansas. From 1903 through 1912, strychnine eliminated 91% of Colorado's prairie dogs. They state:

South Dakota, for example, poisoned approximately 1.2 million hectares (3 million acres) of prairie dogs from 1915 to 1965. During the same period, Montana and Texas poisoned 3.4 and 3.8 million hectares (8.4 and 9.4 million acres), respectively. Arizona poisoned about 0.5 million hectares (1.2 million acres) from 1918 to 1922, and extirpated the prairie dog from the state by 1932 (Alexander 1932; Oakes 2000). Records for some states combined poisoning statistics for all species of prairie dogs, and occasionally with statistics for ground squirrels as well, so that data specifically for black-tailed prairie dogs are elusive. Wyoming, for example, reported only combined statistics for white-tailed and black-tailed prairie dogs through 1965.

The total area across all states poisoned from 1915 to 1965 attributable exclusively to black-tailed prairie dogs was more than 12.1 million hectares (30 million acres) (Figure 5 below), and probably was more than 15 million hectares (37 million acres). The peak year was 1923, when poisoning affected 1.5 million hectares (3.7 million acres). Precise estimates for the cost of the war against prairie dogs are unavailable, but the cumulative cost over the years has been billions of dollars.

By 1960, prairie dogs probably inhabited less than 5% of their geographic range of the early 1800s (Berryman and Johnson 1973; Miller et al. 1994). Survivors persevered on lands with low productivity, in areas with poor access, and on scattered public lands (Anderson et al. 1986; Oakes 2000). Most colonies inhabited less than 10 hectares (25 acres), and only a few complexes (i.e., groups of nearby colonies) inhabited more than 2,500 hectares (6,200 acres) (Linder et al. 1972; Chapter 12). Victims of their own success, poisoning programs started to wane. For the first time, federal agencies began to talk about "stabilizing" efforts to control prairie dogs (Berryman and Johnson 1973).

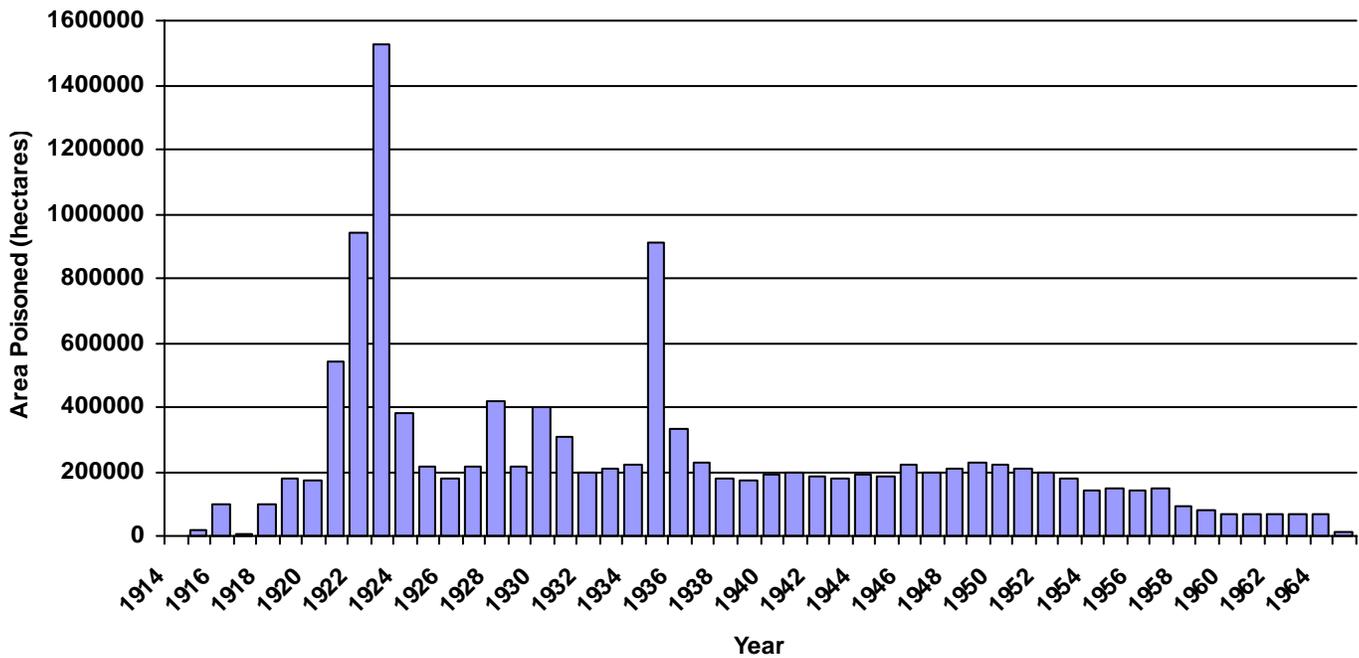


Figure 14. Area poisoned annually for elimination of black-tailed prairie dogs from 1915 to 1965 for all states except Wyoming, and including only the years 1918-1935 for New Mexico and 1921-1923 for Colorado (Forrest and Luchsinger 2006).

The magnitude of today’s poisoning is low relative to poisoning over the past century simply because today’s cumulative colony acreage is less than 5% of the former extent of colonies. There are fewer prairie dogs left to poison. Despite these trends regarding the magnitude of today’s level of poisoning relative to levels of poisoning over the last 100 years, poisoning on federal, state, and private lands has increased since 2004 following the removal of the prairie dog from the list of candidate species for addition to the federal list of endangered and threatened species (Forrest and Luchsinger 2006). Continued drought conditions in much of the Great Plains have led to prairie dog colony expansion and coincided with additional poisoning activities compared to pre-drought.

An obvious cumulative effect arises from the reduction of prairie dog populations due to the current application of rodenticide in the BMZ on MA and non-MA 3.63 (Table 6). Although a suspension of prairie dog control on national grasslands occurred during 1999 to the fall of 2004, poisoning resumed in the BMZ in November, 2004. In addition, the state of South Dakota poisoned 32,055 acres in 2004, 14,725 in 2005, and 29,502 in 2006 (Smith 2007, Kempema 2007).

Since July 1, 2006 the South Dakota bait plant sold 274,500 lbs of oats treated with zinc phosphide (South Dakota State Department of Agriculture 2007). About 6 ounces of treated oats are typically applied per acre of prairie dog colony (Andelt 2006). Enough bait was sold to control 732,000 acres of prairie dogs between July 1, 2006 and the issue date of the winter newsletter of the South Dakota State Department of Agriculture. The state of South Dakota and U.S. Forest Service used only bait from the Pierre, South Dakota bait plant. All of the bait produced at the bait plant is not sold in the study area or even in the state of South Dakota. The South Dakota bait plant is also not the only

source of zinc phosphide treated oats in the area. Nevertheless, a large number of prairie dogs have been controlled in 2006 and there is no indication that this trend will cease.

Given the poisoning of black-footed ferret habitat on private land, it is reasonable that federal lands play an important role in the establishment and management of suitable acreage of prairie dog colonies required for self-sustaining populations of black-footed ferrets (Sidle et al. 2006).

If black-footed ferrets disperse into non-MA 3.63 from Conata Basin or Badlands National Park, the Implementation Plan's Adaptive Response Protocol addresses black-footed ferrets first before carrying out poisoning or other management in response to too many prairie dog colony acres or an inadequate Similarity Index.

Table 6. A summary of prairie dog control conducted on the Nebraska National Forest by geographic area in the fall and winter of 2004, 2005, and 2006.

Geographic Area	Acres Poisoned		
	2004	2005	2006
Oglala	0	1,011	926
Fall River Northeast	2,106	1,998	2,844
Fall River West	768	780	372
Fall River Southeast	366	480	312
Wall North	60	497	940
Wall Southeast	237	950	1,370
Wall Southwest	3,196	2,184	5,530
Fort Pierre	0	211	621
<b>Total</b>	<b>6,733</b>	<b>8,110</b>	<b>12,905</b>

### ***Prairie Dog Shooting***

The possibility of accidental black-footed ferret mortality exists with prairie dog shooting (Joslin and Youmans 1999). Prairie dog shooting could accidentally kill a black-footed ferret, although the potential is probably small because the black-footed ferret is largely nocturnal. However, black-footed ferrets have been known to be aboveground during daylight hours, particularly early morning and late evening periods when prairie dog shooting may be occurring. Prairie dog colonies in non-MA 3.63 are not closed to shooting.

Prairie dog shooting is extensive in South Dakota and affects prairie dog populations. In 2000, 12,219 residents and 3,081 non-residents of South Dakota shot prairie dogs during over 100,000 recreational shooting days (Gigliotti 2000). Shooters killed 1.23 million prairie dogs on non-tribal lands with about 84% shot on private land. Prairie dog shooting occurs throughout the year with peak shooting in the summer.

Although small mammals such as lagomorphs and squirrels recover quickly from hunting via density-dependent vital rates, such recovery is not observed in the black-tailed prairie dog (Pauli and Buskirk 2007a). Because of their coloniality, prairie dogs possess certain life-history traits that predisposed them to be particularly susceptible to hunting associated disturbances, which had cascading effects on population-level processes (Pauli and Buskirk 2007a). Shooting of prairie dogs may significantly reduce prairie dog densities (Vosburg and Irby 1998) and indefinitely maintain reduced densities in colonies (Knowles 1987).

Prairie dogs show no evidence of density dependence in overwinter survival or next-year natality. Rather, shooting induces not only additive effects on survival but also leads to reproductive near-collapse the summer following shooting (Pauli and Buskirk 2007a). Surviving prairie dogs increase alert behaviors eightfold and reduce both aboveground activity and time spent foraging by 66%. Changes in behaviour lower the body condition of surviving adults by 35%. Survivors of shooting,

especially juveniles, exhibited elevated stress levels; fecal corticosterone concentrations increased by 80% among juveniles. Pauli and Buskirk (2007a) discovered that overwinter survival rates did not increase in response to reduced prairie dog density. Pregnancy rates declined by 50% and reproductive output fell by 82%. Risk-disturbance overwhelmed any possible density-dependent effects of shooting in prairie dogs, which exhibited additive mortality in response to hunting, and reproductive failure 1 year after shooting. Risk-disturbance was the predominant mechanism whereby individuals and colonies were affected by hunting. The cumulative effect of shooting prairie dogs outside of Conata Basin/Badlands, where shooting is prohibited, is to lower the quality of the prairie dog colony habitat of black-footed ferrets and limit the effectiveness of black-footed ferret dispersal.

Federal lands are used extensively for prairie dog shooting with expanding lead alloy bullets that leave lead fragments in prairie dog carcasses posing a risk to scavengers (Pauli and Buskirk 2007). Their “results suggest that recreational shooting of prairie dogs contributes to the problem of lead intoxication in wildlife food chains that include prairie dogs. Indeed, some features of recreational shooting, including the killing of large numbers of animals, not removing carcasses from the field, and using expanding bullets, is in contrast to traditional forms of hunting and may present potentially dangerous amounts and particle sizes of metallic lead to scavengers and predators of prairie dogs. Recreational shooting of black-tailed prairie dogs occurs with minimal regulation, yet appears to provide a readily available source of lead to scavenging vertebrates. Few agencies regulate recreational shooting intensity and duration, and none currently regulate the type of ammunition that can be used. Managers should consider measures, such as using non-expanding or lead-free ammunition, to reduce the likelihood of lead poisoning in scavenging raptors and carnivores.”

Another effect is secondary lead poisoning of non-target species caused by lead fragments left in the prairie dog carcasses after they have been shot by prairie dog shooters. In a study conducted in eastern Wyoming two types of bullets were tested to determine how much lead was present in the prairie dog carcasses after they had been shot: a soft point and a full metal jacket (both from .223 caliber rifles). Eighty-seven per cent of prairie dogs shot with soft point bullets contained bullet fragments compared to 7 percent of those shot with full metal jackets. Furthermore, the amount of lead found in prairie dog carcasses differed between the two bullet types; full metal jacket only averaged 19.8 mg of lead, while soft point averaged 225.2 mg of lead (Pauli and Buskirk 2007). Therefore, a scavenger, such as a black-footed ferret, that eats a prairie dog carcass could succumb to lead poisoning. Since black-footed ferrets are known to exist only on the federally administered Conata Basin/Badland complex area, impacts from secondary lead poisoning of ferrets are unlikely. The deposit of lead into the prairie dog colony environment lowers the quality of prairie dog colony habitat as dispersal habitat for black-footed ferrets. The cumulative effect is to heighten the importance of prairie dog colonies in Conata Basin/Badlands, where shooting is prohibited, as habitat for black-footed ferrets.

Unless the small acreages of prairie dog colonies in non-MA 3.63 are not closed to prairie dog shooting, their value to dispersing black-footed ferrets would likely be reduced. Their utility to harbor a nursery population of black-footed ferrets would also be reduced considerably.

### ***Disease and Black-footed Ferret***

Two diseases, canine distemper and sylvatic plague, have notably impacted both wild and captive-reared populations of black-footed ferrets. Sylvatic plague is the principal disease threat to the black-footed ferret. It is an exotic disease foreign to the evolutionary history of North American species and did not exist on this continent prior to 1900 (Gage and Kosoy 2006). It was first observed in prairie dogs in 1932 in Arizona (Cully 1993). Plague is caused by the bacterium, *Yersinia pestis*, which fleas acquire from biting infected animals and can then transmit to other animals via a flea bite. The disease can also be transmitted pneumonically among infected animals or via the consumption of contaminated food items (e.g., black-footed ferrets eating plague-killed prairie dogs) (Godbey et al. 2006).

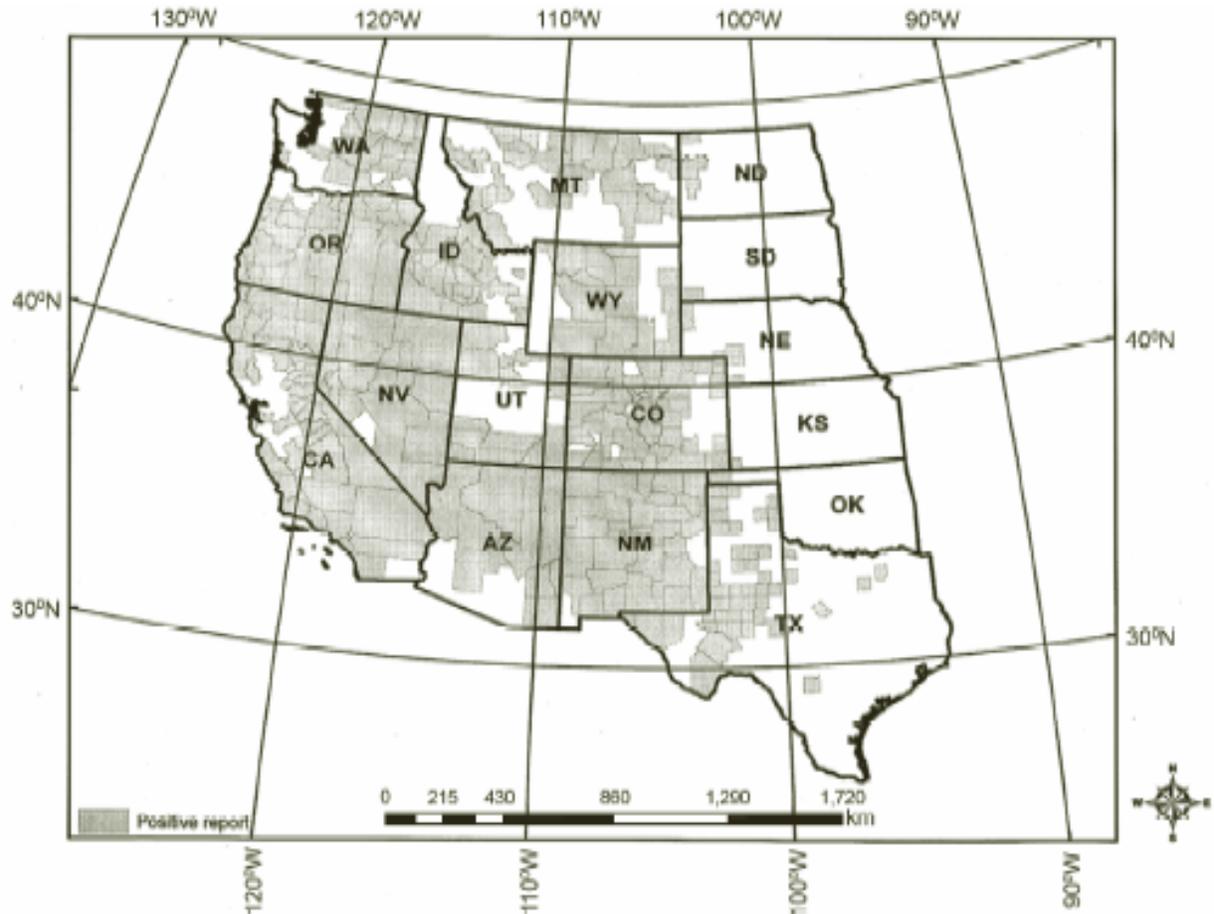


Figure 14. Counties with plague-positive mammals or fleas (1970-present) (Gage and Kosoy 2006). An imaginary plague line runs through the western Dakotas, Nebraska, Kansas, Oklahoma and Texas. The area east of that line is often referred to as plague-free and is the preferred area for black-footed ferret reintroductions.

Since the arrival of plague into North America 100 years ago, prairie dogs have not demonstrated evolved resistance to plague. The introduction of plague has subjected the black-footed ferret to a situation to which it is not evolutionarily adapted, thereby hastening extinction. The Siberian polecat, the black-footed ferret's closest relative, inhabits Asian steppe where plague is native. The polecat's

reproductive strategy has to be more responsive to its varying environment and rodent prey. In contrast, the black-footed ferret appears relatively K-selected, that is, its litters are smaller than the polecat, an expected attribute for a specialist that evolved in the plague-free Great Plains inhabited by a stable prey species, the black-tailed prairie dog (Biggins 2000).

Prairie dogs are highly susceptible to sylvatic plague, a serious threat to the persistence of local prairie dog populations. Plague has been annihilating prairie dogs in the western two thirds of their geographic range since the 1940s and greatly reduced the success of black-footed ferret reintroductions. Plague is rare, almost totally absent among prairie dogs east of a line that approximates the 103<sup>rd</sup> meridian (Cully et al. 2006), near the Wyoming state line (Figure 14). Plague was confirmed in a prairie dog colony in western Custer County, South Dakota in September, 2004 near the Wyoming border, and less than 10 miles from the northern most boundary of the Fall River West Geographic Area. Plague positive prairie dogs were found on tribal lands in Shannon County in 2005, less than 10 miles from the eastern boundary of the Fall River Southeast Geographic Area. Prairie dog colonies may have been affected on the Fall River ranger District because few prairie dogs are observed at many colonies. To date all of the prairie dog colonies thought to be infected by plague have some live prairie dogs within them. The potential for plague to persist in prairie dog populations on the national grasslands and forests in the project area is unknown, but it is acknowledged that plague can have dramatic impacts on prairie dog populations and black-footed ferrets.

Plague can impact the black-footed ferret directly via infection and subsequent mortality and by destroying the black-footed ferret's prey base of prairie dogs. A plague epizootic quickly kills close to 100 percent of prairie dogs in a colony. Plague is also highly prevalent in prairie dog colonies in an enzootic state without any obvious epizootic (Hanson et al. 2007).

Recovery efforts for the black-footed ferret are hampered because both black-footed ferrets and prairie dogs are extremely susceptible to plague (Barnes 1993, Gage and Kosoy 2006). The higher densities and higher rates of social contact of black-tailed and Gunnison's prairie dogs particularly enhance the spread of plague (Cully 1993). The disease is present throughout the range of white-tailed and Gunnison's prairie dogs and is present in approximately the western two-thirds of the range of the black-tailed prairie dog (Figure 6; Barnes 1993, Lockhart et al. 2006). Plague appeared in South Dakota during 2005 and in Conata Basin in 2008.

Until 2005 – 2008, plague had been documented at or within 25 miles of all black-footed ferret reintroduction sites, except the 4 active sites in South Dakota (Conata Basin, Cheyenne River Sioux Reservation, Lower Brule Reservation and Rosebud Sioux Reservation) and the reintroduction site in northern Chihuahua, Mexico (USFWS 2006a). Conata Basin supports the largest black-footed ferret population and has provided a surplus of kits for translocation to other reintroduction areas (Lockhart et al. 2006). However, in 2005, plague struck prairie dogs approximately 25 miles south of Conata Basin. Recovery of prairie dog colonies is occurring. During the late summer and fall of 2005, approximately 3,500 pounds of the insecticide, deltamethrin, a synthetic pyrethroid, were applied to 7,000 acres of occupied prairie dog burrows in known black-footed ferret habitat in an effort to eliminate fleas, the most likely plague vector. Dusting continued in 2008.

On the Buffalo Gap National Grassland, plague is probably responsible for prairie dog die-offs in all three of the Geographic Areas (GA) located on the Fall River Ranger District (Fall River Southeast GA, Fall River Northeast GA, and Fall River West GA), however, no prairie dog carcasses have been collected for analysis. Few prairie dogs can be seen on some colonies in the above areas. On the Fall River Northeast GA, plague has probably occurred on two colonies (approximately 375 acres). On the Fall River Southeast GA, plague has probably occurred on 20 colonies (approximately 1,940 acres). On the Fall River West GA, plague has probably occurred on 5 colonies (approximately 500 acres). Plague may have occurred on additional colonies because only colonies that were not controlled in the

past 3 years were examined. In 2008, plague has affected over 9,000 acres of colonies in Conata Basin MA 3.63. The epizootic continues.

In one instance, the black-footed ferret appears to have prospered despite the periodic presence of plague. In 1991, Shirley Basin, Wyoming was the first reintroduction site. Black-footed ferret releases were suspended there in 1994 due to plague and the small black-footed ferret population present was expected to be lost by the late 1990s. However, since 2002, the Shirley Basin black-footed ferret population has been growing rapidly (Lockhart et al. 2006, Grenier et al. 2007). This site is occupied by white-tailed prairie dogs. White-tailed prairie dog complexes are less densely populated than typical complexes of black-tailed or Gunnison's prairie dogs. Apparently, scattered populations of prairie dogs avoided contracting plague and were able to sustain a black-footed ferret population. Furthermore, outbreaks of plague appear to be associated with moisture and the Shirley Basin is one of the few areas in the white-tailed prairie dog range that has experienced prolonged dry conditions.

In summary, recent studies of plague underscore the negative impact of this disease on black-footed ferret recovery. Although the readily recognized epizootics of plague are indeed a hazard to black-footed ferret recovery (so far, one reintroduced population has failed due to epizootic plague), the chronic presence of cryptic levels of enzootic plague are equally ominous. Over a 12-year period, a population of black-footed ferrets was nearly extirpated when augmentation and plague management ceased at some study sites, but it was reestablished by additional releases and plague management; its future seems insecure without plague management. The South Dakota black-footed ferret reintroduction site at Conata Basin is now experiencing a plague epizootic. The extensive acres in Conata Basin may allow black-footed ferrets to survive the epizootic. Thus, while other black-footed ferret recovery sites, often with much smaller acreages of prairie dog colonies, have been lost due to plague, there is a highly disproportionate cumulative impact to the ferret recovery program of reducing the quality of the Conata Basin site by regulating the extent of prairie dog colonies (Dean Biggins, USGS Research Scientist, Science Consistency Review Report on Nebraska National Forest DEIS, 2007).

The cumulative effect of plague outside of South Dakota and the absence of plague throughout most of South Dakota heighten the importance of federal lands such as national grasslands for black-footed ferret sites. Enhancing options for black-footed ferret sites in the non-MA 3.63 through nursery or larger populations could potentially provide a relief valve for black-footed ferrets should plague spread in either MA-3.63 or non-MA 3.63.

## 10. Determination of Effect and Rationale for the Black-footed Ferret

### The Determination

The biological determination for the black-footed ferret is “*no effect*” for non-MA 3.63 areas of Oglala, Fort Pierre, Fall River West, and Fall River Southeast. The biological determination for the black-footed ferret for non-MA 3.63 area of Wall North, Wall Southeast, and Wall Southwest is “*not likely to jeopardize the continued existence.*” The biological determination for the black-footed ferret for non-MA 3.63 area of Fall River Northeast is “*may affect, not likely to adversely affect.*”

### Rationale

A finding of “no effect” is warranted for non-MA 3.63 areas of Oglala, Fort Pierre, Fall River West, and Fall River Southeast because black-footed ferrets do not occur there. Moreover, because of the GAs’ distance from a known black-footed ferret population, black-footed ferrets are not likely to disperse to the aforementioned GAs.

Two sets of black-footed ferret snow tracks were recently observed by Forest Service staff in Wall Southeast suggesting dispersal from Conata Basin/Badlands. Because of the proximity of the non-MA 3.63 Wall North, Wall Southwest and Fall River Northeast GAs, dispersal to these GAs is possible within the span of the Forest Plan. A finding of “not likely to jeopardize the continued existence” is warranted for non-MA 3.63 areas of Wall North, Wall Southeast, and Wall Southwest. The determination for Fall River Northeast is “may affect, not likely to adversely affect” since it is outside the Section 10(j) designated area.

The proposed action would not change prairie dog management within MA 3.63 areas and is not expected to affect the existing population of black-footed ferrets at Conata Basin/Badlands. Therefore the previous analysis and ESA section 7 conferencing for Forest Plan amendment 2 remain valid.

Setting the minimum and maximum acreage of prairie dog colonies, along with the threshold for vegetation conditions, will reduce and constrain the available habitat that otherwise could be occupied in the future by black-footed ferrets outside of MA 3.63. The proposed action would provide fewer acres of habitat on the upper end of the range than were originally projected under the 2001 Forest Plan.

It is unlikely that a prairie dog colony of 1,000 acres could sustain black-footed ferrets over time without periodic augmentation or other intervention. Nevertheless, small “nursery” populations may be able to play some role in ferret recovery. Wind Cave black-footed ferret reintroduction on 2,800 acres of prairie dog colonies is an experiment to test whether small acreages of prairie dog colonies can sustain a small black-footed ferret population. The outcome may not be known for some years.

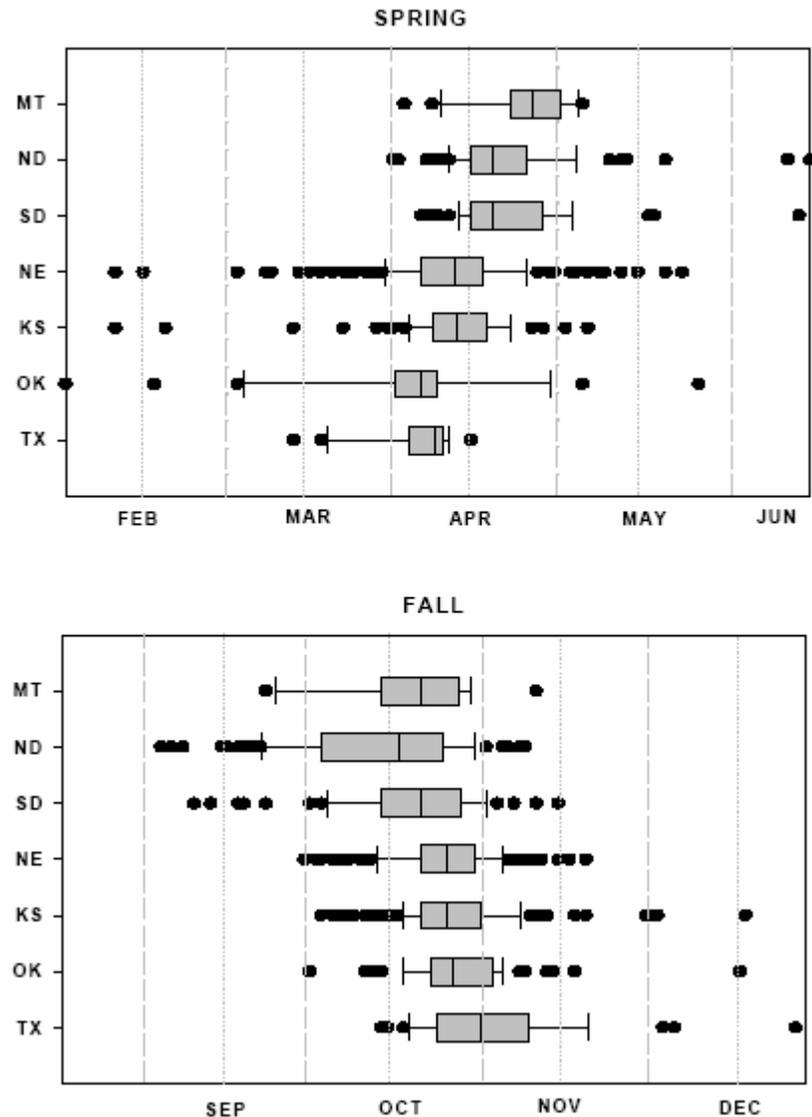
The effect of the proposed low acres of prairie dog colonies in non-MA 3.63 areas also reduces the prospects of successful black-footed ferret dispersal. The 1994 black-footed ferret reintroduction decision (USFWS et al. 1994) designated over 200,000 acres as dispersal habitat for black-footed ferrets, reasoning that black-footed ferrets would ultimately disperse from their point of re-introduction at Conata Basin/Badlands. The 2001 Forest Plan allowed prairie dog colonies to grow in non-MA 3.63 GAs, thereby allowing for future dispersal of black-footed ferrets from Conata Basin/Badlands. The proposed action’s small prairie dog colony acreages greatly curtails or terminates successful black-footed ferret dispersal and confines black-footed ferrets to MA 3.63 areas.

## 11. Whooping Crane (*Grus americana*) Distribution and Status

Whooping cranes were nearly extinct (15 or 16 wintering birds) in 1941, but by 1995 there were 257 birds in captivity and in the wild (Lewis 1995). These large white cranes are rare migrants across Buffalo Gap and Fort Pierre National Grasslands (Peterson et al. 1991, Graupman et al. 1991, Austin and Richert 2001) (Figures 15 and 16). Whooping cranes that migrate across the Great Plains nest in Canada and winter on the Texas Gulf Coast. The U.S. Geological Survey (Austin and Richert 2001) undertook a comprehensive review of observational and site evaluation data of migrant whooping cranes in the United States for the period. 1943-1999. The following discussion is based upon the USGS evaluation.

### Migration

Based on incidental observations, spring migration commences in mid-to late March and continues through mid-late May. Extreme dates include few observations in Oklahoma, Nebraska, and Kansas in February and early March, and in South and North Dakota in June. Some of these outliers were cranes that wintered in Oklahoma or summered in North Dakota. The peak of migration, as indicated by median dates of occurrence, was 8 April in Texas, 6 April in Oklahoma, 12 April in Kansas and Nebraska, 19 April in South and North Dakota, and 26 April in Montana. Few whooping cranes were observed in the United States after early May. The main periods of occurrence in each state over all years seem to be relatively short: the core 50% of the observations ranged from 6 days in Texas to 13 days in South Dakota. Fall migrants were first observed in North and South Dakota in early September; in Nebraska, Kansas, and Oklahoma in early October; and in Texas by mid-October. The peak of migration, as indicated by median dates, were 22 October in Montana, 18 October in North Dakota, 22 October in



**Figure 15. Dates of occurrence of whooping cranes in spring and fall, by state, 1943-1999. Box plots show median (vertical line box), 25<sup>th</sup> and 75<sup>th</sup> percentiles (box), 10<sup>th</sup> and 90<sup>th</sup> percentiles (bars) and outliers (dots) (Austin and Richert 2001).**

South Dakota, 27 October in Nebraska, 27 October in Kansas, 28 October in Oklahoma, and 1 November in Texas.

## Migration Habitat

Migrant whooping cranes use shallow water, including stock dams, as overnight roost sites (Ashton and Dowd 1991). Most wetlands used for roosting during migration were less than 10 acres in size and within 1 km of suitable feeding sites, croplands or wetlands (Lewis 1995). The birds are omnivorous and feed on plants, grain, and animals, including amphibians and invertebrates (Ashton and Dowd 1991).

Palustrine and riverine habitats in the central Great Plains provide roosting and foraging habitat to whooping cranes during spring and fall migration (Austin and Richert 2001). Characteristics of roost habitat have been examined in detail for the Platte River in Nebraska (Johnson 1982, Lingle et al. 1984, Faanes 1992, Faanes and Bowman 1992, Faanes et al. 1992), an area long recognized as a critical habitat for whooping cranes during migration. Although the Platte River is the best known spring stopover area for migrating whooping cranes, whooping cranes also use many other areas during spring and fall migration. Whooping cranes have been observed on various roosting and feeding areas throughout the migration path, which extends through North and South Dakota, Nebraska, Kansas, Oklahoma, and Texas. These areas play a key role in crane migration.

Austin and Richert (2001) evaluated the record of whooping crane sightings during migration. Palustrine wetlands accounted for >75% of records in all states except Nebraska; in that state, the proportions of observations occurring on palustrine and riverine systems were both high (56.0 and 39.6% of state records, respectively). Roost sites were most common on riverine systems only in Nebraska, primarily the Platte, Niobrara, and North and Middle Loup rivers. Most of the whooping cranes found on riverine roosts were single cranes or non-family groups, particularly on the Platte. Whooping cranes were most commonly observed on wetlands having seasonal and semipermanent water regimes. Cranes were observed on a wide range of wetland sizes in both spring and fall, with no apparent pattern relative to social groups. Cranes used portions of rivers that ranged in width from 27 to 457 m and averaged 267 + 87 (SD) m. Maximum depths of wetlands on which cranes were observed averaged 50.8 + 41.4 cm (20.0 + 16.3 inches), while specific sites within wetlands where cranes were observed feeding or roosting averaged 18.0 + 10.7 cm (7.1 + 4.2 inches). Most wetland shorelines were classified as having a slight slope (1 to <5% slope). In riverine systems, roosting cranes were more often observed on unvegetated sites than on vegetated sites, but palustrine roost sites had a broad range of emergent vegetation types.

Most feeding sites were upland crops, whereas dual-use sites were more often wetlands. On upland crop sites, 83% of grain stubble was wheat stubble, 75% of row-crop stubble was corn, and 80% of green crops was winter wheat. Habitats adjacent (<1.6 km [1 mi]) to roost sites were most frequently described as cropland (73.8%) and upland perennial cover (69.5%). Woodland habitat occurred adjacent to >70% of riverine roost sites but <8% of palustrine roost sites. More than two-thirds of sites where cranes were observed were <0.5 mi of human developments. Nearly half of the roost sites and two-thirds of feeding sites had unobstructed visibility of <0.25 mi). Private ownership accounted for >60% of all sites used by whooping cranes and >80% of feeding sites, which reflected the high use of crop fields.

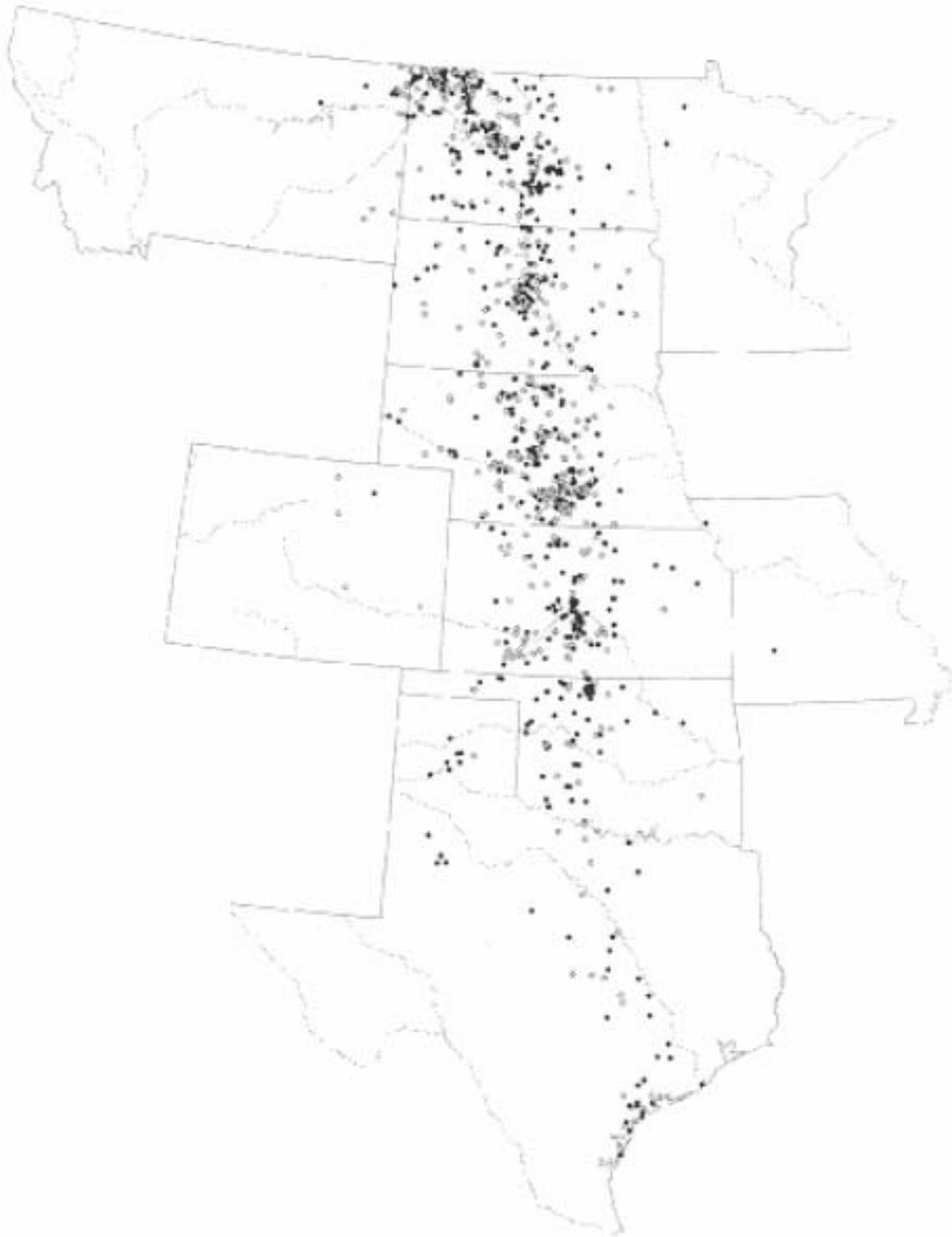


Figure 16. Whooping crane fall (●) and spring (°) sightings, 1943-1999 (Austin and Richert 2001)

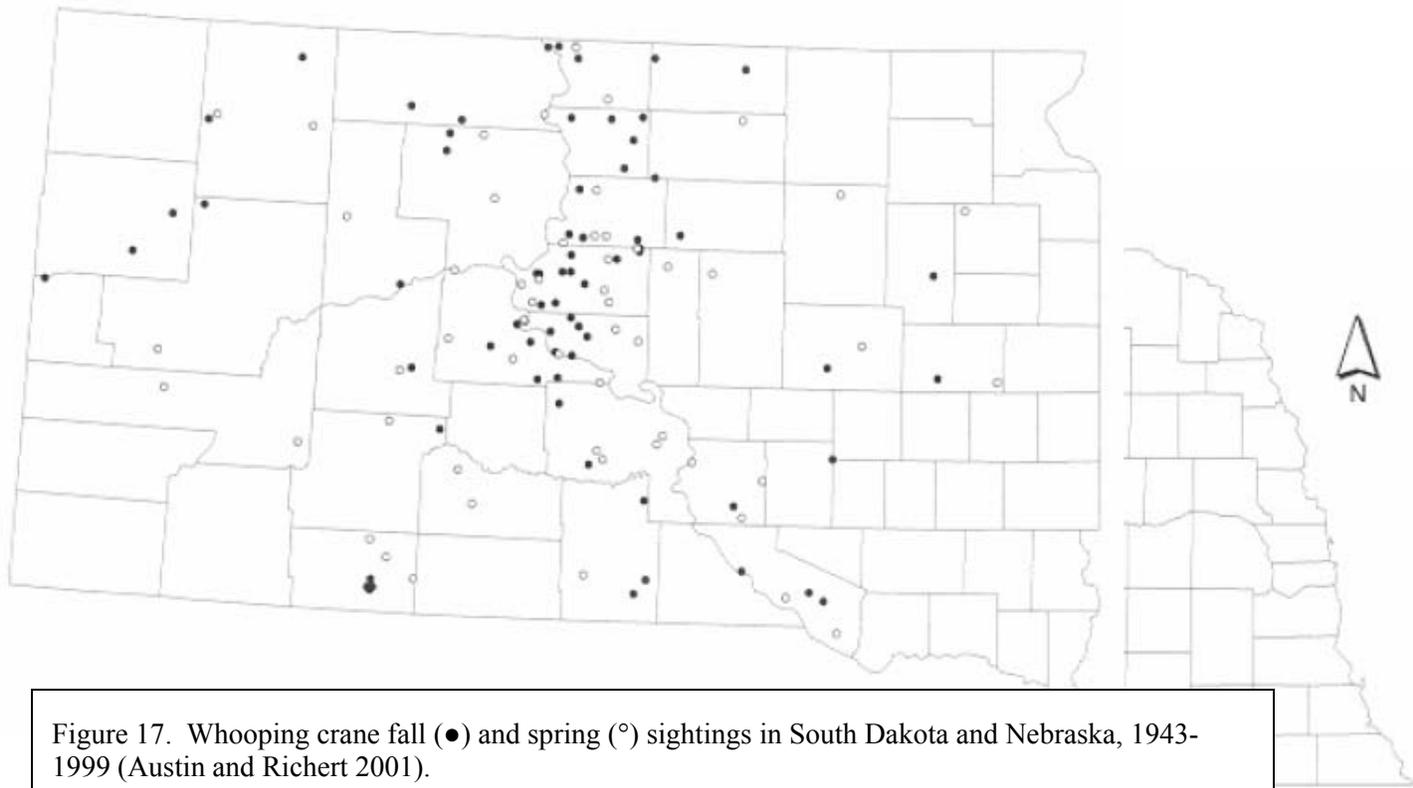
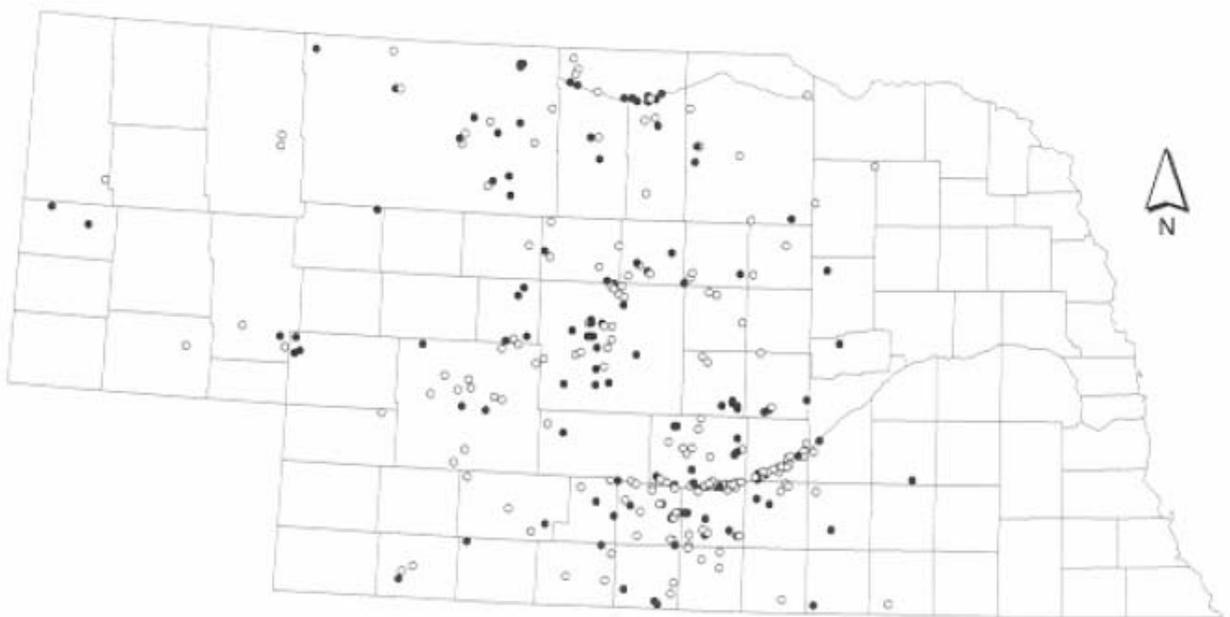


Figure 17. Whooping crane fall (●) and spring (°) sightings in South Dakota and Nebraska, 1943-1999 (Austin and Richert 2001).



Most sites where cranes were observed feeding were in upland crops whereas cranes observed at dual-use sites were more often in wetlands. Seasonally-flooded habitat was largely comprised of flooded pasture (47% of records) and seasonal wetlands (42% of records). Permanent wetlands were largely marshes (30–40%) and reservoirs (30–40%). Sixty percent of upland cover was described as pasture. For upland crops, wheat comprised 83% of small grain stubble, corn comprised about 75% of row-crop stubble, and winter wheat comprised 80% of green crops.

Most sites where cranes were observed feeding were upland crops, with lower occurrence of cranes seen in seasonally flooded habitats, permanent water, or upland perennial cover. No cranes were recorded feeding in woodland. Proportions of habitat types varied little between seasons. Although upland crops occurred in similar high proportions in descriptions of both feeding sites and adjacent habitat, cranes were less frequently observed in upland cover or on wetlands than occurred in adjacent habitat. In spring, cranes most frequently were observed feeding on row-crop stubble, with lesser use of small grain stubble and green crops; <10% of records were for standing small grain, standing row-crops and other. In fall, cranes were most frequently observed on green crops, small-grain stubble, and row-crop stubble. Cranes were infrequently observed in standing small grain, small-grain or row-crop stubble, or in other habitats such as CRP.

Habitats adjacent to roost sites (<1.6 km) most frequently were described as cropland (73.8%) and upland perennial cover (69.5%); permanent wetlands (36.2%) and upland cover (30.5%) were also common. Woodland habitat occurred adjacent to >70% of riverine roost sites but adjacent to <8% of palustrine roost sites. All riverine roosts also had adjacent upland cover, whereas only about half of palustrine roost sites had such adjacent cover; however, upland cropland was common. For both wetland systems, seasonal wetlands occurred more frequently in adjacent habitat for spring roost sites, probably reflecting their seasonal occurrence in the landscape, and permanent wetlands occurred more frequently adjacent to roost sites in fall. Upland cropland was more common in spring than in fall, but the large number of fall records from Cheyenne Bottoms, Quivira, and Salt Plains, where habitat adjacent to roosts is more likely to be non-cropland habitat than on private lands, may be a factor in these seasonal differences.

The most common habitats adjacent to feeding sites were cropland and upland perennial cover; permanent and seasonal wetlands and woodland were less common nearby. Occurrences of seasonal wetlands and upland cover in adjacent habitat were higher in spring than in fall. The higher occurrence of woodland in spring likely relates to greater occurrence of feeding observations in spring on river systems, all of which occurred in Nebraska. Adjacent croplands were most likely to be green crops (winter wheat, alfalfa, winter rye, barley) or row-crop stubble.

Whooping cranes appear similar to sandhill cranes in their frequent use of cropland for feeding particularly corn and wheat stubble (Howe 1987, Johns et al. 1997). However, data from dual use sites indicated that wetlands may provide important feeding areas for some whooping cranes. Howe (1987) did not distinguish between feeding-only and dual-use sites for radio-marked whooping cranes. He noted that the importance of cropland for feeding-only sites was likely higher than the 42% he reported because many feeding sites were actually categorized as roost sites. That is consistent with the frequent use of permanent or seasonally-flooded wetlands for dual-use sites in the USGS evaluation (Austin and Richert 2001).

## 12. Whooping Crane ESA Status and Other Organizational Rankings

ESA Status	Conservation Status <sup>1</sup>
ESA—Endangered	G1, N1N; Nebraska - S1; South Dakota - SNA

<sup>1</sup> Definitions - <http://www.natureserve.org/explorer/nsranks.htm>

### Existing Conditions

In spring and fall, whooping cranes migrate through South Dakota between and conservation agencies monitor the migration. A contingency plan protects whooping cranes should they appear locally during fall migration.

### Direct, Indirect, and Cumulative Effects

There is a remote possibility that whooping cranes could be exposed to rodenticide bait (oats) if they stopped on a project area during migration. Feeding patterns of the cranes, the low concentration of zinc phosphide in the bait, the small amount of bait applied per unit area, widely scattered bait, and the short exposure of bait contribute to low primary and secondary hazards to the birds (Tietjen 1976).

The actual process of applying rodenticide might also deter or scare cranes from the immediate area. Changes in the structure and composition of vegetation after prairie dog poisoning occurs would not affect cranes because they are only present on the ground briefly during migration.

Cumulative effects include mortality and injuries from prairie dog shooters, but this would be unusual. The whooping crane is a well known protected species and there are very stiff penalties for killing or injuring an endangered species. Gunfire and other hunter activities might scare birds locally, but these potential effects are considered insignificant and discountable. Other cumulative effects to whooping cranes include collisions with fences and power lines.

## 13. Determination of Effect and Rationale for the Whooping Crane

### Entire Project Area

The biological determination for the whooping crane under the proposed action is: “*may affect, not likely to adversely affect.*”

### Rationale

Whooping cranes could ingest rodenticide bait in treated colonies. However, the likelihood of whooping cranes landing where rodenticide was recently applied is so remote that it is considered a “discountable effect.” The contingency plan (USFWS 2000a) and consultation with the USFWS would reduce risks to cranes. If whooping cranes are sighted in an area where rodenticide is being applied, poisoning will stop until the cranes leave the area or are hazed out of the area. In addition, if rodenticide has been applied to an area where cranes have been seen, the area will be watched and any cranes appearing will be hazed to eliminate exposure to treated grain. The application of prairie dog rodenticide or sport shooting would

not influence food availability for these migrants. Whooping cranes rarely use prairie dog colonies as habitat. Therefore, the range of prairie dog colony acres is not critical to the whooping crane.

Table 7. Summary of Determinations of Effects of Alternative 1 on black-footed ferret (BFF) and whooping crane (WHC) in the non-MA 3.63 Areas (Geographic Areas) of the Nebraska National Forest.

	Buffalo Gap National Grassland						Oglala National Grassland	Fort Pierre National Grassland
	Fall River West	Fall River Southeast (excludes MA 3.63)	Fall River Northeast	Wall North	Wall Southeast	Wall Southwest (excludes MA 3.63)		
BFF	NE	NE	MA-NLAA	NLJ	NLJ	NLJ	NE	NE
WHC	MA-NLAA	MA-NLAA	MA-NLAA	MA-NLAA	MA-NLAA	MA-NLAA	MA-NLAA	MA-NLAA

BFF = black-footed ferret

WHC = whooping crane

NE = No effect-- where no effect is expected.

MA-NLAA = May affect, not likely to adversely affect -- where effects are expected to be insignificant (immeasurable) or discountable (extremely unlikely to occur).

MA-LAA = May affect, likely to adversely affect -- where effects are expected to be adverse or detrimental.

NLJ = Not likely to jeopardize continued existence -- where effects are expected to be beneficial, insignificant (immeasurable), or discountable (extremely unlikely to occur).

LJ = Likely to jeopardize continued existence -- where effects are expected to reduce appreciably the reproduction, numbers, or distribution of the species

## 14. References for the Biological Assessment and Biological Effects Report

- Agnew, W. 1983. Flora and fauna associated with prairie dog ecosystems. M.S. Thesis. Colorado State University, Fort Collins, Colorado.
- Agnew, W., D.W. Uresk, and R.M. Hansen. 1986. Flora and fauna associated with prairie dog colonies and adjacent ungrazed mixed-grass prairie in western South Dakota. *Journal of Range Management* 39:135-139.
- Agnew, W., D.W. Uresk, and R.M. Hansen. 1988. Arthropod consumption by small mammals on prairie dog colonies and adjacent ungrazed mixed grass prairie in western South Dakota. Pages 81-87 In D. W. Uresk, G.L. Schenbeck, and R. Cefkin (technical editors) Eighth Great Plains wildlife damage control workshop proceedings. General technical Report RM-154, Rocky Mountain Forest and Range Experiment Station, USDA Forest Service, Fort Collins, Colorado.
- Allaby, M. Oxford dictionary of ecology. Oxford University Press, New York.
- Allendorf, F.W. and N. Ryman. 2002. The role of genetics in population viability analysis. Pages 50-85 In S.R. Beissinger and D.R. McCullough, editors, Population viability analysis, University of Chicago Press, Chicago, Illinois.
- Andelt, W. F.. 2006. Methods and economics of managing prairie dogs. Pages 129-138 in J. L. Hoogland editor. Conservation of the black-tailed prairie dog: saving North America's western grasslands. Island Press, Washington D.C.
- Anderson, E., S.C. Forrest, T.W. Clark, and L. Richardson. 1986. Paleobiology, biogeography, and systematics of the black-footed ferret, *Mustela nigripes* (Audubon and Bachman), 1851. *Great Basin Naturalist* 8:11-62.
- Anderson, R.C. 1982. An evolutionary model summarizing the roles of fire, climate, and grazing animals in the origin and maintenance of grasslands: an end paper. Pages 297-308 In J.R. Estes, R.J. Tyrl, and Brunken (editors), Grasses and grasslands; systematics and ecology. University of Oklahoma Press, Norman, Oklahoma.
- Apa, A.D. 1985. Efficiency of two black-tailed prairie dog rodenticides and their impacts on non-target bird species. MS Thesis, South Dakota State University, Brookings, South Dakota.
- Apa, A.D., D. W. Uresk, and R. L. Linder. 1991. Impacts of black-tailed prairie dog rodenticides on nontarget passerines. *Great Basin Naturalist* 51(4):301-309.
- Archer, S., M.G. Garrett, and J.K. Detling. 1987. Rates of vegetation change associated with prairie dog (*Cynomys ludovicianus*) grazing in North American mixed-grass prairie. *Vegetatio* 72:159-166.
- Ashton, D.E., and E.M. Dowd. 1991. Fragile legacy; endangered, threatened and rare animals of South Dakota. South Dakota Dept. of Game, Fish and Parks. Rep. 91-04. 150pp.
- Augustine, D.J., M.R. Matchett, T.P. Toombs, J.F. Cully, T.L. Johnson, and J.G. Sidle. 2008. Spatiotemporal dynamics of black-tailed prairie dog colonies affected by plague. *Landscape Ecology* 23:255-267.

- Austin, Jane E., and Amy L. Richert. 2001. A comprehensive review of observational and site evaluation data of migrant whooping cranes in the United States, 1943-1999. U.S. Geological Survey, Northern Prairie Wildlife Research Center, Jamestown, North Dakota.
- Baker, R. R. 1978. The evolutionary ecology of animal migration. Holmes and Meier, New York, New York.
- Barnes, A.M. 1993. A review of plague and its relevance to prairie dog populations and the black-footed ferret. In Proceedings of the Symposium on the Management of Prairie Dog Complexes for the Reintroduction of the Black-footed Ferret. USFWS Biological Report 13. Pp. 28-37.
- Beissinger, S.R. and D.R. McCullough (editors). 2002. Population viability analysis. University of Chicago Press, Chicago, Illinois.
- Belak, J. R. 2001. Modeling the effects of habitat quality on black-tailed prairie dog habitat occupancy using spatially correlated data. M.S. Thesis, Colorado State University, Fort Collins, Colorado.
- Biggins, D.E., B.J. Miller, L.R. Hanebury, B. Oakleaf, A.H. Farmer, R. Crete, and A. Dood. 1993. A technique for evaluating black-footed ferret habitat. Pages 73-38 in J.L. Oldemeyer, D.E. Biggins, and B.J. Miller, editors. Proceedings of the Symposium on the Management of Prairie Dog Complexes for the Reintroduction of the Black-footed Ferret. USFWS Biological Rep. No. 13.
- Biggins, D.E., B.J. Miller, T.W. Clark, and R.P. Reading. 1997. Management of an endangered species: the black-footed ferret. In Principles of Conservation Biology. Edited by G.K. Meffe and C.R. Carroll. Pp. 420-436.
- Biggins, D.E., J.L. Godbey, T.M. Livieri, M.R. Matchett, and B.D. Bibles. 2006. Post-release movements and survival of adult and young black-footed ferrets. In Recovery of the Black-footed Ferret: Progress and Continuing Challenges. Edited by J.E. Roelle, B.J. Miller, J.L. Godbey, and D.E. Biggins. US Geological Survey. Pp. 189-198.
- Biggins, D.E. 2000. Predation on black-footed ferrets (*Mustela nigripes*) and Siberian polecats (*M. eversmannii*): conservation and evolutionary implications. PhD Dissertation. Colorado State University, Fort Collins, Colorado.
- Biggins, D.E. 2001. Prairie dog research update #2. Unpublished report, U.S. Geological Survey, Midcontinent Science Center, Fort Collins, Colorado.
- Biggins, D.E. 2006a. The symposium in context. In Recovery of the Black-footed Ferret: Progress and Continuing Challenges. Edited by J.E. Roelle, B.J. Miller, J.L. Godbey, and D.E. Biggins. U.S. Geological Survey. Pp. 3-5.
- Biggins, D. E., J. L. Godbey, M. R. Matchett, L. Hanebury, and T. M. Livieri. 2006b. Monitoring black-footed ferrets during reestablishment of free-ranging 40 populations: discussion of alternative methods and prescription of minimum standards. Pages 155-174 in Recovery of the black-footed ferret – progress and continuing challenges (J.E. Rolle, B.J. Miller, J.L. Godbey, and D.E. Biggins, editors). U.S. Geological Survey Scientific Investigations Report 2005-5293. 288 pp.
- Biggins, D. E., J.L. Godbey, M. R. Matchett, and T. M. Livieri. 2006c. Habitat preferences and intraspecific competition in black-footed ferrets. Pages 129-142 in Recovery of the black-footed ferret – progress and continuing challenges (J.E. Rolle, B.J. Miller, J.L. Godbey, and D.E. Biggins, editors). U.S. Geological Survey Scientific Investigations Report 2005-5293. 288 pp.
- Bolen, E.G. and W.L. Robinson. 2003. Wildlife ecology and management. Prentice-Hall, Upper Saddle River, New Jersey.

- Bonham, C. D. and A. Lerwick. 1976. Vegetation changes induced by prairie dogs on shortgrass range. *Journal of Range Management* 29(3):221-225.
- Boyce, M.S. 1992. Population viability analysis. *Annual Review of Ecology and Systematics* 23: 481-506.
- Boyce, M.S. 1997. Population viability analysis: adaptive management for threatened and endangered species. Pages 226-236 In M.S. Boyce and A. Haney, editors, *Ecosystem management*. Yale University Press, New haven, Connecticut.
- Boyce, M.S. 2001. Population viability analysis: development, interpretation, and application. Pages 123-136 In T.M. Shenk and A.B. franklin, editors. *Modelling in natural resource management*. Island press, Washington, D.C.
- Boyce, M. S., B.M. Blanchard, R. R. Knight, and C. Servheen. 2001. Population viability for grizzly bears. *International Association for Bear Research and Management Monograph Series Number 1*.
- Boyce, M. S., E.M. Kirsch, and C. Servheen. 2002. Bet-hedging applications for conservation. *Journal of Biosciences* 27:385-392.
- Boyce, M. S. 1997. Population viability analysis: adaptive management for threatened and endangered species. Pages 226–236 in M. S. Boyce and A. Haney, eds. *Ecosystem management: applications for sustainable forest and wildlife resources*. Yale University Press, New Haven, Connecticut.
- Breck, S.W., D.E. Biggins, T.M. Livieri, M.R. Matchett, and V. Kopcsó. 2005. Does predator management enhance survival of reintroduced black-footed ferrets? *In* *Recovery of the Black-footed Ferret: Progress and Continuing Challenges*. Edited by J.E. Roelle, B.J. Miller, J.L. Godbey, and D.E. Biggins. U.S. Geological Survey. Pp. 203-209.
- Brown, B. 2003. Prairie dogs, catde, and conventional wisdom-part I. Pages 81-86 in M. Peale, editor. *The new ranch at work*. Quivera Coalition, Santa Fe, New Mexico.
- Brussard, P.F., and M.E. Gilpin. 1989. Demographic and genetic problems of smallpopulations. Pages 37-48 In U.S. Seal, E.T. Thorne, M.A. Bogan, and S.H. Anderson, editors, *Conservation biology and the black-footed ferret*. Yale University Press, New Haven, Connecticut.
- Cable, K. A., and R. N. Timm. 1988. Efficacy of deferred grazing in reducing prairie dog infestation rates. U.S. Forest Service General Technical Report RM-154:46-49.
- Campbell, T. M., III, and T. W. Clark. 1981. Colony characteristics and vertebrate associations of white-tailed and black-tailed prairie dogs in Wyoming. *American Midland Naturalist* 105:269-276.
- Carlson, D.C. and E.M. White. 1987. Effects of prairie dogs on mound soils. *Journal of the American Society of Soil Scientists* 51:389-393.
- Carr, J.F. 1973. A rancher's view towards prairie dogs. Pages 168-171 In R.L. Linder and C.N. Hillman, editors. *Proceedings of the black-footed ferret and prairie dog workshop, September 4-6, 1973, Rapid City, South Dakota*. South Dakota State University, Brookings, South Dakota.
- Caughley, G. 1994. Directions in conservation biology. *Journal of Animal Ecology* 63:215–244.
- Caughley, G. and A.R.E. Sinclair. 1994. *Wildlife ecology and management*. Blackwell Scientific Publications, Cambridge, Massachusetts.
- Chace, G.E. 1973. Prairie dogs, ferrets and cattle-conflict on the plains. *Animal Kingdom* 76:2-8.
- Cincotta, R. P. 1985. Habitat and dispersal of black-tailed prairie dogs in Badlands National Park. Ph.D Thesis, Colorado State University, Fort Collins, Colorado.

- Clark, T.W. 1986. Technical introduction. In Great Basin Naturalist Memoirs No. 8 The Black-footed Ferret. S.L. Wood Editor. Brigham Young University. Pp. 8-10.
- Clark, T.W., S.C. Forrest, L. Richardson, D.E. Casey, and T.M. Campbell III. 1986. Description and history of the Meeteetse black-footed ferret environment. Great Plains Naturalist 8:72-84.
- Clark, Tim W. 1989. Conservation Biology of the Black-footed Ferret (*Mustela nigripes*). Wildlife Preservation Trust, Special Scientific Report No. 3. Philadelphia, Pennsylvania.
- Clippinger, N.W. 1989. Habitat suitability index model: black-tailed prairie dog. USFWS, Biol. Rep. 82(10).
- Clobert, J., R. Anker Ims, and F. Rousset. 2004. Causes, mechanisms and consequences of dispersal. Pages 307-335 in I. Hanski and O.E. Gaggiotti, editors, Ecology, genetics, and evolution of metapopulations. Academic Press, New York.
- Collinge, S. K., W. C. Johnson, C. Ray, R. Matchett, J. Grensten, J. F. Cully, Jr., K. L. Gage, M. Y. Kosoy, J. E. Loye, and A. P. Martin. 2005. Landscape structure and plague occurrence in black-tailed prairie dogs on grasslands of the western USA. Landscape Ecology 20:941–955.
- Collins, A. R., J. P. Workman, and D. W. Uresk. 1984. An economic analysis of black-tailed prairie dog (*Cynomys ludovicianus*) control. Journal of Range Management 37:358-361.
- Conservation Breeding Specialist Group. 2004. Black-footed ferret population management planning workshop. Final Report. IUCN/SSC Conservation Breeding Specialist Group: Apple Valley, Minnesota.
- Cooper, J and L. Gabriel 2005. South Dakota black-tailed prairie dog conservation and management plan. [Online] South Dakota Game Fish and Parks. Available: <http://www.sdgifp.info/Wildlife/hunting/Prairiedogfinalplan.pdf> [January 30, 2007]
- Coppock, D.L., J.K. Detling, J.E. Ellis, and M.I. Dyer. 1983a. Plant- herbivore interactions in North American mixed- grass prairie I. Effects of black –tailed prairie dogs on intraseasonal aboveground plant biomass and nutrient and plant species diversity. Oecologia 56:1-9.
- Coppock, D.L., J.E. Ellis, J.K. Detling and M.I. Dyer. 1983b. Plant- herbivore interactions in a North American mixed-grass prairie II. Responses of bison to modification of vegetation by prairie dogs. Oecologia 56: 10-15.
- Cottam, C. and M. Caroline. 1965. The black-tailed prairie dog in Texas. Texas Journal of Science 17:294-302.
- Cully, J.F. 1993. Plague, prairie dogs, and black-footed ferrets. In Proceedings of the Symposium on the Management of Prairie Dog Complexes for the Reintroduction of the Black-footed Ferret. USFWS Biological Report 13. Pp. 38-49.
- Cully, J.F. and E.S. Williams. 2001. Interspecific comparisons of sylvatic plague in prairie dogs. Journal of Mammalogy 82:894-904.
- Cully, J. F., D. E. Biggins, and D. B. Seery. 2006. Conservation of prairie dogs in areas with plague. Pages 157-168 in J. L. Hoogland editor. Conservation of the black-tailed prairie dog: saving North America’s western grasslands. Island Press, Washington D.C.
- Curtin, C. 2003. Prairie dogs, cattle, and conventional wisdom-part II. Pages 87-92 in M. Peale, editor. The new ranch at work. Quivera Coalition, Santa Fe, New Mexico.
- Curtin, C. 2006. Initial results of experimental studies of prairie dogs in arid grasslands: implications for landscape conservation and importance of scale. U.S. Forest Service Proceedings RMRS-P-40:57-58.

- Czech, B. and P.R. Krausman. 2001. The endangered species act: history, conservation biology, and public policy. The John Hopkins University Press, Baltimore, Maryland.
- Day, T.A. 1988. Modification of individual plant and community water and nitrogen relations by grassland herbivores. PhD Thesis, Colorado State University, Fort Collins, Colorado.
- Deisch, M.S. 1986. The effects of three rodenticides on nontarget small mammals and invertebrates. MS Thesis, South Dakota State University, Brookings, South Dakota.
- Deisch, M.S., D.W. Uresk, and R.L. Linder. 1990. Effects of prairie dog rodenticides on deer mice in western South Dakota. *Great Basin Naturalist* 50:347-353.
- Derner, J.D., J.K. Detling, and M.F. Antolin. 2006. Are livestock weight gains affected by black-tailed prairie dogs? *Frontiers in Ecology and Environment* 4(9): 459–464.
- Detling, J.K., and E.L. Painter. 1983. Defoliation responses of western wheatgrass populations with diverse histories of prairie dog grazing. *Oecologia* 57:65-71
- Detling, J.K., E.L. Painter, and D.L. Coppock. 1986. Ecotypic differentiation resulting from grazing pressure: evidence for a likely phenomenon. Pages 431-433 in P.G. Joss, P. W. Lynch, and O.B. Williams, eds. *Rangelands: A Resource under Siege*. Australian Academy of Science, Canberra, Australia.
- Detling, J.K. 1988. Grassland and savannas: regulation of energy flow and nutrient cycling by herbivores. Pages 131-148 in L.R. Pomeroy and J.J. Alberts, eds. *Concepts of Ecosystem Ecology*. Springer-Verlag, New York.
- Detling, J. K. 2006. Do prairie dogs compete with livestock? Pages 65-88 in J. L. Hoogland editor. *Conservation of the black-tailed prairie dog: saving North America’s western grasslands*. Island Press, Washington D.C.
- Dowd Stukel, E., J. Sidle, and A. Nicholas. 2004. Results of South Dakota black-tailed prairie dog acreage survey, 2002-2003. Wildlife Division Report 2004-12. South Dakota Department of Game, Fish and Parks, Pierre, South Dakota.
- Ernst, A.E., A.L. Clark, and D.R. Gober. 2005. A habitat-based technique to allocate black-footed ferret recovery among jurisdictional entities. In *Recovery of the Black-footed Ferret: Progress and Continuing Challenges*. Edited by J.E. Roelle, B.J. Miller, J.L. Godbey, and D.E. Biggins. U.S. Geological Survey. Pp. 89-95.
- Ewer. R.F. 1973. *The carnivores*. Cornell University Press, Ithaca, New York.
- Extension Toxicology Network 1993. Zinc Phosphide. [Online] EXTTOXNET. Available: <http://pmep.cce.cornell.edu/profiles/exttoxnet/pyrethrins-ziram/zinc-phosphide-ext.html> [March 1, 2007]
- Fagerstone, K.A. and D.E. Biggins. 1986. Comparison of capture-recapture and visual count indices of prairie dog densities in black-footed ferret habitat. In *Great Basin Naturalist Memoirs No. 8 The Black-footed Ferret*. S.L. Wood Editor. Brigham Young University. Pp. 94-98.
- Fagerstone, K.A. and C.A. Ramey. 1996. Rodents and lagomorphs. Pages 83-132 In P.R. Krausman, editor, *Rangeland Wildlife*. The Society for Range Management, Denver, Colorado.
- Fagerstone, K.A., H.P. Tietjen, J.F. Glahn, G.L. Schenbeck, and J. Bourassa. 2005. Black-tailed prairie dog colony dynamics in South Dakota over a 10-year period. Pages 323-336 In D.L. Nolte and K.F. Fagerstone, eds, *Proceedings Eleventh Wildlife Damage Management Conference*.
- Fahnestock, J. T., and J. K. Detling. 2002. Bison-prairie dog-plant interactions in a North American mixed-grass prairie. *Oecologia* 132:8695.

- Fahrig, L. and G. Merriam. 1994. Conservation of fragmented populations. *Conservation Biology* 8: 50–59.
- Farrar, J. P. 2002. Effects of prairie dog mound-building and grazing activities on vegetation in the central grasslands. Thesis, Colorado State University, Fort Collins, Colorado.
- Forrest, S.C., T.W. Clark, L. Richardson, and T.M. Campbell III. 1985. Black-footed ferret habitat: some management and reintroduction considerations. Wyoming BLM Wildlife Technical Bulletin No. 2.
- Forrest, S.C., Biggins, D.E., Richardson, L., Clark, T.W. Campbell III, K.A. Fagerstone, and E.T. Thorne. 1988. Population attributes for the black-footed ferret (*Mustela nigripes*) at Meeteetse, Wyoming, 1981-1985. *Journal of Mammalogy* 69: 261-273.
- Forrest, S.C. and J.C. Luchsinger. 2006. Past and current chemical control of prairie dogs. Pages 115-128 In J.L. Hoogland, editor, *Conservation of the black-tailed prairie dog*. Island Press, Washington, DC.
- Foster, K.R., P. Vecchia, and M.H. Repacholi. 2000. Science and the Precautionary Principle. *Science* 288:979-981.
- Fox-Parrish, L. 2002. Attitudes and opinions of landowners and general citizens relative to the black-tailed prairie dog. MS thesis. Emporia State University, Emporia, Kansas.
- Frankham, R.J., J.D. Ballou, and D.A. Briscoe. 2002. *Introduction to conservation genetics*. Cambridge University Press, Cambridge, United Kingdom.
- Frankham, R.J., J.D. Ballou, and D.A. Briscoe. 2004. *A primer of conservation genetics*. Cambridge University Press, Cambridge, United Kingdom.
- Freilich, J. E., J. M. Emlen, J. J. Duda, D. C. Freeman, and P. J. Cafaro. 2003. Ecological effects of ranching: a six-point critique. *BioScience* 8: 759-765.
- Fritz, S. 2008. A long-term perspective on drought history in the Great Plains. *Water Current* 40(2):6-7, 14.
- Gage, K.L. and M.Y. Kosoy. 2006. Recent trends in plague ecology. *In* *Recovery of the Black-footed Ferret: Progress and Continuing Challenges*. Edited by J.E. Roelle, B.J. Miller, J.L. Godbey, and D.E. Biggins. U.S. Geological Survey. Pp. 213-231.
- Gaggiotti, O.E. and I. Hanski. 2004. Mechanisms of population extinction. Pages 337-366 In I. Hanski and O.E. Gaggiotti, editors, *Ecology, genetics, and evolution of metapopulations*. Academic Press, New York.
- Garelle, D., P. Marinari, and C. Lynch. 2006. Black-footed ferret species survival plan. American Zoo and Aquarium Association Population Management Center.
- Gigliotti, L.M. 2000. Prairie dog shooting in South Dakota. South Dakota Game, Fish and Parks, Pierre, South Dakota.
- Giles, R.H. 1982. Management knowledge through wildlife research: a perspective. *Environmental Management* 6:185-191.
- Gilpin, M.E. 1999. An inquiry into the population viability of the black-tailed prairie dog. [http://www.homepage.montana.edu/~mgilpin/prairie\\_dog.html](http://www.homepage.montana.edu/~mgilpin/prairie_dog.html).
- Gilpin, M.E., and M.E. Soulé. 1986. Minimum viable populations: processes of species extinction. Pages 19 – 34 in: Soulé, M.E. (ed.). *Conservation Biology: The Science of Scarcity and Diversity*. Sunderland, MA: Sinauer Associates.

- Godbey, J.L., D.E. Biggins, and D. Garrelle. 2006. Exposure of captive black-footed ferrets to plague and implications for species recovery. In *Recovery of the Black-footed Ferret: Progress and Continuing Challenges*. Edited by J.E. Roelle, B.J. Miller, J.L. Godbey, and D.E. Biggins. US Geological Survey. Pp. 233-237.
- Gold, I. K. 1976. Effects of black-tailed prairie dog mounds on shortgrass vegetation. MS Thesis. Colorado State University, Fort Collins, Colorado.
- Goodrich, J. M., and S. W. Buskirk. 1998. Spacing and ecology of North American badgers (*Taxidea taxus*) in a prairie-dog (*Cynomys leucurus*) complex. *Journal of Mammalogy* 79:171-179.
- Graupman, K., R. A. Peterson, and N. R. Whitney, compilers. 1991. Birds of the Buffalo Gap National Grassland, field checklist. USDA Forest Service, Chadron, Nebraska.
- Grenier, M. B., D.B. McDonald, and S.W. Buskirk. 2007. Rapid growth of a critically endangered carnivore. *Science* 317:779.
- Groves, C.R., and T.W. Clark. 1986. Determining minimum population size for recovery of the black-footed ferret. *Great Basin Naturalist Memoirs* 8:150-159.
- Guenther, D. A., and J. K. Detling. 2003. Observations of cattle use of prairie dog towns. *Journal of Range Management* 56:410-417.
- Gurevitch, J., S.M. Scheiner, and G.A. Fox. 2002. *The ecology of plants*. Sinauer Associates, Sunderland, Massachusetts.
- Hanebury, L.R. and D.E. Biggins. 2006. A history of searches for black-footed ferrets. Pages 47-65 In J.E. Roelle, B.J. Miller, J.L. Godbey, and D.E. Biggins, editors, *Recovery of the black-footed ferret: progress and continuing challenges*. U.S. Geological Survey Scientific Investigations Report 2005-5293.
- Hanski, I. 1994a. A practical model of metapopulation dynamics. *Journal of Animal Ecology* 63:151-162
- Hanski, I. 1994b. Patch-occupancy dynamics in fragmented landscapes. *Trends in Ecology and Evolution* 9:131-135.
- Hanski, I. 1999. *Metapopulation ecology*. Oxford University Press, New York.
- Hanski, I. 1997. Metapopulation dynamics: from concepts and observations to predictive models. Pages 69-91 in I. Hanski, and M. E. Gilpin, eds. *Metapopulation biology: ecology, genetics and evolution*. Academic Press, San Diego, California.
- Hanski, I. and D. Simberloff. 1997. The metapopulation approach: its history, conceptual domain, and application to conservation. Pages 5-26 in *Metapopulation biology: ecology, genetics, and evolution* (I. Hanski and M. E. Gilpin, eds.). Academic Press, San Diego, California.
- Hansen, R. M., and I. K. Gold. 1977. Black-tail prairie dogs, desert cottontails and cattle trophic relations on shortgrass range. *Journal of Range Management* 30:210-213.
- Hanson, D.A., H.B. Britten, M. Restani, and L.R. Washburn. 2007. High prevalence of *Yersinia pestis* in black-tailed prairie dog colonies during an apparent enzootic phase of sylvatic plague. *Conservation Genetics* 8:789-795.
- Harris, R.B., T.W. Clark, and M.L. Shaffer. 1989. Extinction probabilities for isolated black-footed ferret populations. Pages 69-82 In U.S. Seal, E.T. Thorne, M.A. Bogan, and S.H. Anderson, editors, *Conservation biology and the black-footed ferret*. Yale University Press, New Haven, Connecticut.

- Harrison, S. and D. Taylor. 1997. Empirical evidence for metapopulation dynamics. Pp. 27–42 in *Metapopulation biology: ecology, genetics, and evolution* (I. A. Hanski and M. E. Gilpin, eds.). Academic Press, San Diego, California.
- Hassien, F.D. 1976. A search for black-footed ferrets in the Oklahoma panhandle and adjacent area and an ecological study of black-tailed prairie dogs in Texas County, Oklahoma. MS Thesis, Oklahoma State University, Stillwater, Oklahoma
- Hellstedt, P. and H. Henttonen. 2006. Home range, habitat choice and activity of stoats (*Mustela erminea*) in a subarctic area. *Journal of Zoology* 269: 205-212.
- Higgins, K.F., E. Dowd Stukel, J.M. Goulet, and D.C. Backlund. 2000. Wild mammals of South Dakota. South Dakota Department of Game, Fish and Parks.
- Hillman, C.N. 1968. Field observations of black-footed ferrets in South Dakota. *Transactions of the North American Wildlife and Natural Resources Conference* 33:433-443.
- Hillman, C.N. and R.L. Linder. 1973. The black-footed ferret. Pages 10-23 In R.L. Linder and C.N. Linder (editors). *Proceedings of the black-footed ferret and prairie dog workshop*. South Dakota State University, Brookings, South Dakota.
- Hillman, C.N., R.L. Linder, and R.B. Dahlgren. 1979. Prairie dog distribution in areas inhabited by black-footed ferrets. *American Midland naturalist* 102:185-187.
- Hillman, C.N. and T.W. Clark. 1980. *Mustela nigripes*. In *Mammalian Species No. 126*. The American Society of Mammalogists.
- Holechek, J. L., T. T. Baker, and J. C. Boren. 2004. Impacts of controlled grazing versus grazing exclusion on rangeland ecosystems: what we have learned? Range Improvement Task Force Report No. 57. College of Agriculture and Home Economics, New Mexico State University, Las Cruces, New Mexico.
- Holechek, J.L., R.D. Pieper, and C.H. Herbel. 2004. *Range management: principles and practices*. Pearson Education, Upper Saddle River, New Jersey.
- Holmes, E.E. and B.X. Semmens. 2004. Viability analysis for endangered metapopulations: a diffusion approximation approach. Pages 565-597 In I. Hanski and O.E. Gaggiotti, editors, *Ecology, genetics, and evolution of metapopulations*. Elsevier Academic Press, Burlington, Massachusetts.
- Holsinger, K.E. 2000. Demography and extinction in small populations. Pages 55-74 In A. Young and G. Clark, editors, *Genetics, demography, and the viability of fragmented populations*. Cambridge University Press, Cambridge, United Kingdom.
- Hooge, P. N., B. Eichenlaub, and E. Solomon. 1999. The animal movement program. Alaska Biological Science Center, U.S. Geological Survey, Anchorage, Alaska.
- Hoogland, J.L., 1995. *The black-tailed prairie dog, social life of a burrowing mammal*. The University of Chicago Press, Chicago, Illinois.
- Hoogland, J.L. (editor). 2006. *Conservation of the black-tailed prairie dog*. Island Press, Washington, D.C.
- Howard, W.E. 1960. Innate and environmental dispersal of individual vertebrates. *American Midland Naturalist* 63:152-161.
- Howard, J, R.M. Santymire, P.E. Marinari, J.S. Kreeger, L. Williamson, and E.E. Wildt. 2006. Use of reproductive technology for black-footed ferret recovery. Pages 28-36 In J.E. Roelle, B.J. Miller, J.L. Godbey, and D.E. Biggins, editors, *Recovery of the Black-footed Ferret: Progress and Continuing Challenges*. U.S. Geological Survey Scientific Investigations Report 2005-5293.

- Huston, M.A. 1994. Biological diversity: the coexistence of species on changing landscapes. Cambridge University Press, New York.
- Hutchins, M., R.J. Wiese, and J. Bowdoin. 1996. Black-footed ferret recovery program analysis and action plan. American Zoo and Aquarium Association.
- Ingham, R. E., and J.K. Detling. 1984. Plant herbivore interactions in a North American mixed grass prairie III. Soil nematode populations and root biomass on *Cynomys ludovicianus* colonies and adjacent uncolonized areas. *Oecologia* 63:307-313.
- Intergovernmental Panel on Climate Change. 2007. The AR4 Synthesis Report, Working Group I Report "The Physical Science Basis", Working Group II Report "Impacts, Adaptation and Vulnerability, Working Group III Report "Mitigation of Climate Change".
- IUCN. 1996. 1996 IUCN red list of threatened animals. IUCN, Gland, Switzerland.
- Jachowski, D.S. 2007. Resource selection by black-footed ferrets in relation to the spatial distribution of prairie dogs. MS Thesis. University of Missouri, Columbia, Missouri.
- Jachowski, D.S., J.L. Millspaugh, D.E. Biggins, T.M. Livieri, and M.R. Matchett. 2008. Implications of black-tailed prairie dog spatial dynamics to black-footed ferrets. *Natural Areas Journal* 28:14-25.
- Jaramillo, V. J., and J. K. Detling. 1988. Grazing history, defoliation, and competition: Effects on shortgrass production and nitrogen accumulation. *Ecology* 69:1599-1608.
- Johnson, L. E., L. R. Albee, R. O. Smith, and A. L. Moxon. 1951. Cows, calves and brass: effects of grazing intensities on beef cow and calf production an on mixed prairie vegetation on western South Dakota ranges. *South Dakota Bulletin* 412. South Dakota College, Brookings, South Dakota.
- Johnson-Nistler, C.M., B.F. Sowell, H.W. Sherwood, and C.L. Wambolt. 2004. Black-tailed prairie dog effects on Montana's mixed-grass prairie. *Journal of Range Management* 57:641-648.
- Kayser, M. 1998. Have varmint rifle, will travel. *American Hunter* (June):44-47, 61-62.
- Kempema, S.L.F. 2007. South Dakota black-tailed prairie dog colony acreage and distribution, 2006. Wildlife Division Report 2007-07. South Dakota Department of Game, Fish and Parks, Pierre, South Dakota.
- Kilpatrick, C.W., S.C. Forrest, and T.W. Clark. 1986. Estimating genetic variation in the black-footed ferret – a first attempt. Pages 145-149 In S.L. Wood, editor, *The Black-footed Ferret*. Great Basin Naturalist Memoirs No. 8.
- King, J.A. 1955. Social behavior, social organization, and population dynamics in a black-tailed prairie dog town in the Black Hills of South Dakota. *Contributions of the Laboratory of Vertebrate Biology* 67:1-123. University of Michigan, Ann Arbor, Michigan.
- Klute, D. S., L. W. Ayers, M. T. Green, W. H. Howe, S. L. Jones, J. A. Shaffer, S. R. Sheffield, and T. S. Zimmerman. 2003. Status assessment and conservation plan for the western burrowing owl in the United States. USFWS, Denver, Colorado.
- Knapp, A. K., J. M. Blair, J. M. Briggs, S. L. Collins, D. C. Hartnett, L. C. Johnson, and E. G. Towne. 1999. The keystone role of bison in North American tallgrass prairie. *BioScience* 49:39-50.
- Knowles, C.J. 1982. Habitat affinity, populations and control of black-tailed prairie dogs on the Charles M. Russell National Wildlife Refuge. Ph.D. Dissertation. University of Montana, Missoula, Montana.
- Knowles, C.J. 1986. Some relationships of black- tailed prairie dogs to livestock grazing. *Great Basin Naturalist* 46:198-203.

- Knowles, C.J. 1986. Population recovery of black-tailed prairie dogs following control with zinc phosphide. *Journal of Range Management* 39:249-250.
- Knowles, C. J. 1987. An evaluation of shooting and habitat alteration for control of black-tailed prairie dogs. Pages 53-56 In D. W. Uresk, G.L. Schenbeck, and R. Cefkin (technical editors) Eighth Great Plains wildlife damage control workshop proceedings. General technical Report RM-154, Rocky Mountain Forest and Range Experiment Station, USDA Forest Service, Fort Collins, Colorado.
- Knowles, C. J. 2000. Black-tailed prairie dog population viability assessment for North Dakota. Dec. 14, 2000. Report to the North Dakota Game and Fish Department. 57 pp.
- Knowles, C.J., C.J. Stoner S. P. Gieb. 1982. Selective use of black-tailed prairie dog towns by mountain plovers. *Condor* 84:71-74.
- Kotliar, N.B., B.W. Baker, A.D. Whicker, and G. Plumb. 1999. A critical review of assumptions about the prairie dog as a keystone species. *Environmental Management* 24:177-192.
- Kotliar, N.B. 2000. Application of the new keystone-concept to prairie dogs: How well does it work? *Conservation Biology* 14:1715-1721.
- Kotliar, N.B., B.J. Miller, R.P. Reading, and T.W. Clark. 2006. The prairie dog as a keystone species. Pages 53-64 In J.L. Hoogland, editor, *Conservation of the black-tailed prairie dog*. Island Press, Washington, DC.
- Kuchler, A.W. 1964. *Potential Natural Vegetation of the Conterminous United States*, American Geographical Society, Special Publication No. 36.
- Lacey, J. R., and H. W. Van Poolen. 1981. Comparison of herbage production on moderately grazed and ungrazed western ranges. *Journal of Range Management* 34:210-212.
- Lacy, R.C. 2000. Structure of the VORTEX simulation model for population viability analysis. *Ecological Bulletins* 48:191-203.
- Lamb, B.L., K. Cline, A. Brinson, N. Sexton, and P.D. Ponds. 2001. Citizen knowledge of and attitudes toward black-tailed prairie dogs. Completion report, USGS Open-File Report 01-471. USGS, Midcontinent Ecological Science Center, Fort Collins, Colorado.
- Lamb, B.L. and K. Cline. 2003. Public knowledge and perceptions of black-tailed prairie dogs. *Human Dimensions of Wildlife* 8:127-143.
- Lamb, B.L., R. P. Reading, and W. F. Andelt. 2006. Attitudes and perceptions about prairie dogs. Pages 108-114 in J. L. Hoogland editor. *Conservation of the black-tailed prairie dog: saving North America's western grasslands*. Island Press, Washington DC.
- Lande, R. 1995. Mutation and conservation. *Conservation Biology* 9:782-791.
- Lande, R. 2002. Incorporating stochasticity in population viability analysis. Pages 18-40 In S.R. Beissinger and D.R. McCullough, editors, *Population viability analysis*, University of Chicago Press, Chicago, Illinois.
- Langer, T.J. 1998. Black-tailed prairie dogs as indicators of human-caused changes in landscape at Badlands National Park, South Dakota. M.S. Thesis, North Carolina State University, Raleigh, North Carolina.
- Lebreton, J.-D. and J. Clobert. 1991. Bird population dynamics, management, and conservation: the role of mathematical modelling. Pages 105-125 In C.M. Perrins, J.-D. Lebreton, and G.J.M. Hiron, editors. *Bird population studies: relevance to conservation and management*. Oxford University Press, Oxford, United Kingdom.

- Lee, C.D. and F.R. Henderson. 1989. Kansas attitudes on prairie dog control. Great Plains Wildlife Damage Control Workshop 9:162-165.
- Leitzell, T.L. 1986. Species protection and management decisions in an uncertain world. Pages 243-254 In B.G. Norton, editor, The preservation of species. Princeton University Press, Princeton, New Jersey.
- Lewis, J. K., G. M. Van Dyne, L. R. Albee, and F. W. Whetzal. 1956. Intensity of grazing: its effect on livestock and forage production. South Dakota Bulletin 459. South Dakota State College, Brookings, South Dakota.
- Lewis, J.C. 1995. Whooping crane (*Grus americana*) in The birds of North America, No. 153 (A. Poole and F. Gill, eds.). The Acad. of Nat. Sci., Philadelphia, and The American Ornithological Union, Washington D.C.
- Livieri, T.M. and W. Perry. 2005. Effects analysis of black-tailed prairie dog reduction on black-footed ferret populations in Conata Basin. Unpublished report, Wall, South Dakota.
- Livieri, T.M. 2006. Ten-Year History of the Conata Basin Black-Footed Ferret Population: 1996-2005. Report to U.S. Forest Service, Prairie Wildlife Research, Fort Collins, Colorado.
- Livieri, T.M. 2007a. Estimates of black-footed ferret population viability and spatial requirements. Final report to U.S. Forest Service, Prairie Wildlife Research, Fort Collins, Colorado.
- Livieri, T.M. and D. Biggins. 2007b. The value of black-footed ferrets as a harvestable “crop.” Unpublished report. On file in the administrative record. Nebraska National Forest. Chadron, Nebraska.
- Livieri, T.M. 2007c. Black-footed ferret spatial use of prairie dog colonies in South Dakota. MS Thesis, University of Wisconsin, Stevens Point, Wisconsin.
- Lockhart, J.M., E.T. Thorne, and D.R. Gober. 2006. A historical perspective on recovery of the black-footed ferret and the biological and political challenges affecting its future. In Recovery of the Black-footed Ferret: Progress and Continuing Challenges. Edited by J.E. Roelle, B.J. Miller, J.L. Godbey, and D.E. Biggins. U.S. Geological Survey. Pp. 6-19.
- Lomolino, M.V. and G.A. Smith. 2003. Terrestrial vertebrate communities at black-tailed prairie dog (*Cynomys ludovicianus*) towns. Biological Conservation 115:89-100.
- Lomolino, M. V., and G. A. Smith. 2003. Prairie dogs as islands: applications of island biogeography and landscape ecology for conserving non-volant terrestrial vertebrates. Global Ecology and Biogeography 12: 275-286.
- Luce, R. J. 2001. An umbrella, multi-state approach for the conservation and management of the black-tailed prairie dog, *Cynomys Ludovicianus*, in the United States – an addendum to the Black-tailed Prairie Dog Conservation Assessment and Strategy, November 3, 1999.
- Luce, R. J. 2002. A multi-state conservation plan for the black-tailed prairie dog, *Cynomys Ludovicianus*, in the United States – an addendum to the Black-tailed Prairie Dog Conservation Assessment and Strategy, November 3, 1999.
- Luce, R.J. 2006. Areas where habitat characteristics could be evaluated to identify potential black-footed ferret reintroduction sites and develop conservation partnerships. In Recovery of the Black-footed Ferret: Progress and Continuing Challenges. Edited by J.E. Roelle, B.J. Miller, J.L. Godbey, and D.E. Biggins. U.S. Geological Survey. Pp. 69-88.
- Luce, R.J., R. Manes, and W. Van pelt. 2006. A multi-state plan to conserve prairie dogs. Pages 210-217 In J.L. Hoogland, editor, Conservation of the black-tailed prairie dog. Island Press, Washington, DC.

- Ludwig, D. and C.J. Walters. 2002. Fitting population viability analysis into adaptive management. Pages 511-520 In S.R. Beissinger and D.R. McCullough, editors, Population viability analysis, University of Chicago Press, Chicago, Illinois.
- MacArthur, R.H. and E.O. Wilson. 1967. The theory of island biogeography. Princeton University Press, Princeton, New Jersey.
- Maehr, D.S., R.C. Lacy, E.D. Land, O.L. Bass, and T.S. Hoctor. 2002. Evolution of population viability assessments for the Florida panther: a multiperspective approach. Pages 284-311 In S.R. Beissinger and D.R. McCullough, editors, Population viability analysis, University of Chicago Press, Chicago, Illinois.
- Magle, S. 2003. Black-tailed prairie dog (*Cynomys ludovicianus*) response to human intrusion and urban development in the Colorado Front Range. M.S. Thesis, University of Wisconsin, Madison, Wisconsin.
- Manes, R. 2006. Does the prairie dog merit protection via the Endangered Species Act. Pages 169-183 In J.L. Hoogland, editor, Conservation of the black-tailed prairie dog. Island Press, Washington, DC.
- Manno, T.G., F.S. Dobson, J.L. Hoogland, and D.W. Foltz. 2007. Social group fission and gene dynamics among black-tailed prairie dogs (*Cynomys ludovicianus*). Journal of Mammalogy 88:448-456.
- Marinari, P.E. and J.S. Kreeger. 2006. An adaptive management approach for black-footed ferrets in captivity. In Recovery of the Black-footed Ferret: Progress and Continuing Challenges. Edited by J.E. Roelle, B.J. Miller, J.L. Godbey, and D.E. Biggins. U.S. Geological Survey. Pp 23-27.
- MacDonald D. 1983. The ecology of carnivore social behavior. Nature 301: 384–397.
- MacDonald D.W. 1981. Resource dispersion and the social organization of the red fox (*Vulpes vulpes*). In: Proceedings of the Worldwide Furbearer Conference (eds J.A. Chapman & D. Pursey) pp. 918–49. Frostburg, Maryland.
- McCain, L.A., R.P. Reading, and B.J. Miller. 2002. Prairie dog gone: myth, persecution, and preservation of a keystone species. Pages 230-235 In M. Matteson and G. Wuerthner, editors. Welfare ranching: the subsidized destruction of the American west. Island Press, Washington, D.C.
- McCullough, D.R. (editor). 1996. Metapopulations and wildlife conservation. Island Press, Washington, D.C.
- McNaughton, S. J. 1984. Grazing lawns: animals in herds, plant form and coevolution. American Naturalist 124:863-868.
- Meffe, G.K. and C.R. Carroll. 1997. Principles of conservation biology. 2<sup>nd</sup> Edition. Sinauer Associates, Inc., Sunderland, Massachusetts.
- Merriam, C.H. 1902. The prairie dog of the Great Plains. Pages 257-270 in USDA Yearbook.
- Miller, B., R.P. Reading, and S. Forrest. 1996. Prairie night. Smithsonian Institution Press, Washington, DC.
- Miller, B.J., R.P. Reading, D.E. Biggins, J.K. Detling, S.C. Forrest, J.L. Hoogland, J. Javersak, S.D. Miller, J. Proctor, J. Truett, and D.W. Uresk. 2007. Prairie dogs: an ecological review and current biopolitics. Journal of Wildlife Management 71:2801-2810.

- Miller, P.S, and R.C. Lacy. 2005. VORTEX, A Stochastic Simulation of the Extinction Process. Version 9.50 User's Manual. Conservation Breeding Specialist Group (SSC/IUCN), Apple Valley, Minnesota.
- Mills, L.S., J.M. Scott, K.M. Strickler, and S.A. Temple. 2005. Ecology and management of small populations. Pages 691-713 In C.E. Braun, editor, Techniques for Wildlife Investigations and Management. The Wildlife Society, Bethesda, Maryland.
- Mills, L.S. 2007. Conservation of wildlife populations: demography, genetics, and management. Blackwell Publishing, Malden, Massachusetts.
- Mills L.S. and Knowlton F.F. 1991. Coyote space use in relation to prey abundance. Canadian Journal of Zoology 69:1516–1521.
- Molles, M.C. 2005. Ecology: concepts and applications. McGraw-Hill, New York.
- Morgan, N.K., P. Newman, and G.N. Wallace. 2007. Conflicts associated with recreational shooting on the Pawnee national Grassland. Human Dimensions of Wildlife 12:145-156.
- Morrison, M.L., B.G. Marcot, and R.W. Mannan. 2006. Wildlife-habitat relationships. Island Press, Washington, DC.
- Moul, F. 2006. The national grasslands: a guide to America's undiscovered treasures. University of Nebraska Press, Lincoln, Nebraska.
- National Park Service. 2006a. Black-footed ferret reintroduction plan, draft environmental assessment. National Park Service, Wind Cave National Park, Hot Springs, South Dakota.
- National Park Service. 2006b. Black-tailed prairie dog management plan, draft environmental assessment. National Park Service, Wind Cave National Park, Hot Springs, South Dakota.
- National Park Service. 2007. Black-footed ferret reintroduction plan: finding of no significant impact. National Park Service, Wind Cave National Park, South Dakota.
- National Research Council. 1995. Science and the endangered species act. National Academy Press, Washington, DC.
- NatureServe. 2006. NatureServe Explorer: An online encyclopedia of life [web application]. . Arlington, VA <http://www.natureserve.org/explorer>. [March 2007].
- O'Meilia, M.E., F.L. Knopf, and J.C. Lewis. 1982. Some consequences of competition between prairie dogs and beef cattle. Journal of Range Management 35:580-585.
- Osborn, B., and P.F. Allan. 1949. Vegetation of an abandoned prairie dog town in tall-grass prairie. Ecology 30:322-332.
- Ovaskainen, O. and I. Hanski. 2004. Metapopulation dynamics in highly fragmented landscapes. Pages 73-103 In I. Hanski and O.E. Gaggiotti, editors, Ecology, genetics, and evolution of metapopulations. Academic Press, New York.
- Paine, R.T. 1969. A note on trophic complexity and community stability. American Naturalist 103: 91-93.
- Pauli, J.N. 2005. Ecological studies of the black-tailed prairie dog (*Cynomys ludovicianus*): implications for biology and conservation. M.S. Thesis, University of Wyoming, Laramie, Wyoming.
- Pauli, J.N. and S.W. Buskirk. 2007a. Recreational shooting of prairie dogs: a portal for lead entering wildlife food chains. Journal of Wildlife Management 71:103-108.

- Pauli, J.N. and S.W. Buskirk. 2007b. Risk-disturbance overrides density dependence in a hunted colonial rodent, the black-tailed prairie dog *Cynomys ludovicianus*. *Journal of Applied Ecology* 44:1219-1230.
- Peterson, R. A., R. Riis, and R. V. Summerside, compilers. 1991. Birds of central South Dakota, including the Fort Pierre National Grassland. Field Checklist. USDA Forest Service.
- Potter, R. L. 1980. Secondary successional patterns following prairie dog removal on shortgrass range. M.S. Thesis. Colorado State University, Fort Collins, Colorado.
- Powell, R. A. 1979. Mustelid spacing patterns: variations on a theme by *Mustela*. *Zeitschrift für Tierpsychologie* 50:153-165.
- Proctor, J., B. Haskins, and S.C. Forrest. 2006a. Focal areas for conservation of prairie dogs. Pages 232-247 In J.L. Hoogland, editor, Conservation of the black-tailed prairie dog. Island Press, Washington, D.C.
- Proctor, J., B. Haskins, and S.C. Forrest. 2006b. Identifying focal areas for conservation of black-footed ferrets and prairie dog associates. Pages 271-274 In J.E. Roelle, B.J. Miller, J.L. Godbey, and D.E. Biggins, editors, Recovery of the black-footed ferret: progress and continuing challenges. U.S. Geological Survey Scientific Investigations Report 2005-5293.
- Purves, W.K. and G.H. Orians. 1983. Life: the science of biology. Sinauer Associates, Sunderland, Massachusetts.
- Ralls, K., S.R. Beissinger, and J. Fitts Cochrane. 2002. Guidelines for using population viability analysis in endangered species management. Pages 521-550 In S.R. Beissinger and D.R. McCullough, editors, Population viability analysis, University of Chicago Press, Chicago, Illinois.
- Reading, R.P. 1993. Toward an endangered species reintroduction paradigm: a case study of the black-footed ferret. PhD dissertation, Yale University, New Haven, Connecticut.
- Reading, R.P. and S.R. Kellert. 1993. Attitudes toward a proposed reintroduction of black-footed ferrets (*Mustela nigripes*). *Conservation Biology* 7:569-580.
- Reading, R.P., B.J. Miller, and S.R. Kellert. 1999. Values and attitudes toward prairie dogs. *Anthroöös* 12:43-52.
- Reading, R.P., T.W. Clark, L. McCain, and B.J. Miller. 2002. Black-tailed prairie dog conservation: a new approach for a 21st century challenge. *Endangered Species Update* 19:162-170.
- Reading, R.P., L. McCain, T.W. Clark, and B.J. Miller. 2005. Understanding and resolving the black-tailed prairie dog conservation challenge. Pages 209-223 In R. Woodroffe, S.J. Thirgood, and A. Rabinowitz, editors. People and wildlife: conflict or co-existence. Cambridge University Press, Cambridge, United Kingdom.
- Reed, D.H., J.J. O'Grady, B.W. Brook, J.D. Ballou, and R. Frankham. 2003. Estimates of minimum viable population sizes for vertebrates and factors influencing those estimates. *Biological Conservation* 113:23-34.
- Reid, N.J. 1954. The distribution of the black-tailed prairie dog in the Badlands of southwestern North Dakota. MS Thesis, Iowa State University, Ames, Iowa.
- Richardson, L.T., W. Clark, S.C. Forrest, and T.M. Campbell III. 1987. Winter ecology of black-footed ferrets (*Mustela nigripes*) at Meeteetse, Wyoming. *American Midland Naturalist* 117: 225-239.
- Ricklefs, R.E. and G.L. Miller. 2000. Ecology. W.H. Freeman and Company, New York.
- Roach, J.L., P. Stapp, B. Van Horne, and M.F. Antolin. 2001. Genetic structure of a metapopulation of black-tailed prairie dogs. *Journal of mammalogy* 82:946-959.

- Rocke, T.E., P. Nol, P.E. Marinari, J.S. Kreeger, S.R. Smith, G.P. Andrews, and A.W. Friedlander. 2006. Vaccination as a potential means to prevent plague in black-footed ferrets. *In* Recovery of the Black-footed Ferret: Progress and Continuing Challenges. Edited by J.E. Roelle, B.J. Miller, J.L. Godbey, and D.E. Biggins. U.S. Geological Survey. Pp 243-247.
- Roelle J. E., B.J. Miller, J.L. Godbey, and D.E. Biggins 2006. Recovery of the Black-footed Ferret: Progress and Continuing Challenges. U.S. Geological Survey. Reston, Virginia.
- Saether, B. and S. Engen. 2002. Including uncertainties in population viability analysis using population prediction intervals. Pages 191-212 In S.R. Beissinger and D.R. McCullough, editors, Population viability analysis, University of Chicago Press, Chicago, Illinois.
- Sandell, M. 1989. The mating tactics and spacing patterns of solitary carnivores. Pages 164-182 *in* J. L. Gittleman, editor. Carnivore behavior, ecology, and evolution Volume 1. Cornell University Press, Ithaca, New York.
- Sauer, C.O. 1950. Grassland, climax, fire, and man. *Journal of Range Management* 3:16-22.
- Schenbeck, G.L., and G. Mason. 1994. Summary of the 1990 and 1992 prairie dog hunter surveys on the Buffalo Gap National Grassland in the Conata Basin, South Dakota. Unpublished report. Nebraska National Forest, Chadron, Nebraska.
- Schenbeck, G.L., and G. Mason, and C Loop. 1994. Aerial photograph inventory of black-tailed prairie dog colonies in the Conata Basin/Badlands area of South Dakota, Summer, 1993. Unpublished report. Nebraska National Forest, Chadron, Nebraska.
- Schitoskey F. 1975. Primary and secondary hazards of three rodenticides to kit fox. *Journal of Wildlife Management* 39:416-418.
- Seal, U.S., E.T. Thorne, M.A. Bogan, and S.H. Anderson (editors). 1989. Conservation biology and the black-footed ferret. Yale University Press, New Haven, Connecticut.
- Seery, D.B., D.E. Biggins, J.A. Monteneri, R.E. Ensore, D.T. Tanda, and K.L. Gage. 2003. Treatment of black-tailed prairie dog burrows with deltamethrin to control fleas (Insecta: Siphonaptera) and plague. *Journal of Medical Entomology* 40:718-722.
- Seery, D.B. 2006. Use of pesticides to mitigate the effects of plague Pages 238-242 In J.E. Roelle, B.J. Miller, J.L. Godbey, and D.E. Biggins (editors) Recovery of the Black-footed Ferret: Progress and Continuing Challenges. U.S. Geological Survey Scientific Investigations Report 2005-5293.
- Severe, D.S. 1977. Revegetation of blacktail prairie dog mounds on shortgrass prairie in Colorado. MS Thesis, Colorado State University, Fort Collins, Colorado.
- Sexton, N.R., A. Brinson, P.D. Ponds, K. Cline, and B.L. Lamb. 2001. Citizen knowledge and perception of black-tailed prairie dog management: Report to respondents. USGS Open-File Report 01-467, USGS Midcontinent Ecological Science Center, Fort Collins, Colorado.
- Shaffer, M.L. 1987. Minimum viable populations: coping with uncertainty. Pages 69-86 In M.E. Soulé (editor). Viable populations for conservation. Cambridge University Press, New York, New York.
- Shaffer, M., L Hood Watchman, W.J. Snape III, and I.K. Latchis. 2002. Population viability analysis and conservation policy. Pages 123-142 In S.R. Beissinger and D.R. McCullough, editors, Population viability analysis, University of Chicago Press, Chicago, Illinois.
- Sharps, J. C., and D. W. Uresk. 1990. Ecological review of black-tailed prairie dogs and associated species in western South Dakota. *Great Basin Naturalist* 50:339-345.

- ShIPLEY, B. K. and R. P. Reading. 2006. A comparison of herpetofauna and small mammal diversity on black-tailed prairie dog (*Cynomys ludovicianus*) colonies and non-colonized grassland in Colorado. *Journal of Arid Environments* 66:27-41.
- Shipworth, D. and R. Kenley. 1996. Fitness landscapes and the Precautionary Principle: the geometry of environmental risk. *Environmental management* 24:121-131.
- Sidle, J.G., G.L. Schenbeck, E.A. Lawton, and D.S. Licht. 2006. Role of federal lands in the conservation of the black-tailed prairie dog. Pages 218-231 In J.L. Hoogland, editor, *Conservation of the black-tailed prairie dog*. Island Press, Washington, DC.
- Slobodkin, L. B. 1986. On the susceptibility of different species to extinction: elementary instructions for owners of a world. Pages 226-242 In B.G. Norton, editor, *The preservation of species*. Princeton University Press, Princeton, New Jersey.
- Smith, R.E. 1958. *Natural history of the prairie dog in Kansas*. University of Kansas Museum of Natural History and State Biological Survey of Kansas.
- Smith, A. 2007. South Dakota prairie dog control. Report submitted to the Feb 22, 2007 Interagency Animal damage Control meeting, Wall, South Dakota.
- Snell, G.P., and B.D. Hlavachick. 1980. Control of prairie dogs--the easy way. *Rangelands* 2:239-240.
- Snell, G. P. 1985. Results of control of prairie dogs. *Rangelands* 7:30
- Soulé, M. E., editor. 1987. *Viable populations for conservation*. Cambridge University Press, Cambridge, United Kingdom.
- Soulé, M. E. and J. Terborgh (editors). 1999. *Continental conservation: scientific foundation of regional reserve networks*. Island Press, Washington, DC.
- Soulé, M.E., J.A. Estes, J. Berger, and C. Martinez Del Rio. 2003. Ecological effectiveness: conservation goals for interactive species. *Conservation Biology* 17:1238-1250.
- Soulé, M.E., J.A. Estes, B. Miller, and D.L. Honnold. 2005. Strongly interacting species: conservation policy, management, and ethics. *BioScience* 55:168-176.
- South Dakota Department of Agriculture. 2007. Pesticide applicators newsletter. Winter 2007 Issue 35. Division of Agricultural Services. Office of Agronomy Services 523 East Capitol – Foss Building Pierre, South Dakota 57501-3182.
- South Dakota Department of Agriculture, South Dakota Department of Game, Fish and Parks, United States Department of Agriculture – ADC, United States Fish and Wildlife Service; Cooperative Extension Service; and United States Department of Agriculture – SCS. 1994. *Prairie Dog Management in South Dakota* (1.1mb) South Dakota Department of Agriculture 1994. Brochure entitled *Prairie Dog Management In South Dakota*.
- South Dakota Department of Game, Fish and Parks. 2001. Executive Summary, *Prairie Dog Shooting in South Dakota*. HD-8-02.AMS.
- South Dakota Department of Game, Fish and Parks. 2005. News Release. January 24, 2005.
- Stapp, P., M.F. Antolin, and M. Ball. 2004. Patterns of extinction in prairie-dog metapopulations: plague outbreaks follow El Niño events. *Frontiers of Ecology and Environment* 2:235-240.
- Stebbins, G.L. 1981. Coevolution of grasses and herbivores. *Annals of the Missouri Botanical Garden* 68:75-86.
- Stokstad, E. 2005. What's wrong with the Endangered Species Act? *Science* 309:2150-2152.

- Stoltenberg, M.B. 2004. Effects of prairie dogs on plant community composition and vegetation disappearance in mixed-grass prairie. MS Thesis, South Dakota State University, Brookings, South Dakota.
- Summers and Linder. 1978. Food habits of the black-tailed prairie dog in western South Dakota. *Journal of Range Management* 31:134-136.
- Svingen, D. 2006. Black-tailed prairie dog conservation assessment and strategy for the Medora Ranger District, Little Missouri National Grassland, North Dakota. Dakota Prairie Grasslands, internal report.
- Svingen, D. 2008. Black-tailed prairie dog conservation assessment and strategy: Grand River Ranger District. Dakota Prairie Grasslands, internal report.
- Temple, S.A. and J.R. Cary. 2002. Reserve design. Pages 281-292 In S.E. Gergel and M.G. Turner, editors, *Learning landscape ecology: a practical guide to concepts and techniques*, Springer Science and Business Media, New York, New York.
- Terrall, D.F. 2006. Use of natural vegetative barriers to limit black-tailed prairie dog town expansion in western South Dakota. MS Thesis, South Dakota State University, Brookings, South Dakota.
- Thomas, C.D. 1990. What do real population dynamics tell us about minimum viable population sizes? *Conservation Biology* 4: 324-327.
- Tietjen H. P. 1976. Zinc phosphide-Its development as a control agent for black-tailed prairie dogs. Special Scientific Report-Wildlife No. 195. USFWS. Washington DC.
- Tileston, J. V., and R. R. Lechleitner. 1966. Some comparisons of the black-tailed and white-tailed prairie dogs in north-central Colorado. *American Midland Naturalist* 75:292-316.
- Timm, R. M. 1983. Description of active ingredients. pp.G31-G131. In R.M. Timm ed. *Prevention and Control of Wildlife Damage*. Great Plains Agricultural Council Wildlife Resource Committee and Cooperative Extension Service University of Nebraska, Lincoln, Nebraska.
- Turner, M.G., R.H. Gardner, and R.V. O'Neill. 2001. *Landscape ecology in theory and practice*. Springer-Verlag, New York, New York.
- Tyler, J. D. 1968. Distribution and vertebrate associates of the black-tailed prairie dog in Oklahoma. Ph.D. Dissertation, University of Oklahoma, Norman, Oklahoma.
- U.S. Fish and Wildlife Service. 1988. Black-footed ferret recovery plan. U.S. Fish and Wildlife Service, Denver, Colorado.
- U.S. Fish and Wildlife Service. 1993. Endangered and threatened wildlife and plants; Proposed establishment of a nonessential experimental population of black-footed ferrets in southwestern South Dakota. *Federal Register* 58(95): 29176-29186.
- U.S. Fish and Wildlife Service. 1994. Endangered and threatened wildlife and plants; Establishment of a nonessential experimental population of black-footed ferrets in southwestern South Dakota. *Federal Register* 59(159):42682-42694.
- U.S. Fish and Wildlife Service, U.S. Forest Service, and U.S. National Park Service. 1994. Final environmental impact statement, black-footed ferret reintroduction, Conata Basin/Badlands, South Dakota.
- U.S. Fish and Wildlife Service and National Marine Fisheries Service. 1998. *Consultation Handbook: Procedures for conducting consultation and conference activities under section 7 of the Endangered Species Act*. U. S. Government Printing Office. Superintendent of Documents. Washington, DC.

- U.S. Fish and Wildlife Service. 2000a. 2000-2001 Contingency plan. Federal-state cooperative protection of whooping cranes. Aransas/Matagorda Island National Wildlife Refuge, Austwell, Texas.
- U.S. Fish and Wildlife Service. 2000b. Endangered and Threatened Wildlife and Plants; 12-month finding for a petition to list the black-tailed prairie dog as threatened. Federal Register 65(24).
- U.S. Fish and Wildlife Service. 2004a. Endangered and threatened wildlife plants; finding for the resubmitted petition to list the black-tailed prairie dog as threatened. Federal Register 69 (159):51217-51226.
- U.S. Fish and Wildlife Service. 2004b. Species assessment and listing priority assignment for the black-tailed prairie dog. Current States/Counties/Territories/Countries of Occurrence.
- U.S. Fish and Wildlife Service. 2006a. Draft recovery plan for the black-footed ferret (*Mustela nigripes*). U.S. Fish and Wildlife Service, Denver, Colorado.
- U.S. Fish and Wildlife Service. 2006b. Draft Five-Year Status Review of the black-footed ferret (*Mustela nigripes*). U.S. Fish and Wildlife Service, Denver, Colorado.
- U.S. Government Accountability Office. 2007. Agricultural conversion: farm program payments are an important factor in landowners' decisions to convert grassland to cropland. GAO-07-1054, Washington, DC.
- USDA Animal and Plant Health Inspection Service. 1994. Animal Damage Control Program Final Environmental Impact Statement. Appendix P Risk Assessment of Wildlife Damage Control Methods used by USDA Animal Damage Control Program. USDA APHIS Washington DC.
- USDA Forest Service. 1995a. Biological assessment for black-footed ferret. Rocky Mountain Region, USDA Forest Service. Denver, Colorado.
- USDA Forest Service. 1995b. Biological evaluation for whooping crane. Rocky Mountain Region, USDA Forest Service, Denver, Colorado.
- USDA Forest Service 2001a. Unpublished Report on the viability and recovery of black-footed ferrets, Chadron, Nebraska.
- USDA Forest Service 2002b. Final environmental impact statement for the northern Great Plains management plans revision. USDA Forest Service.
- USDA Forest Service 2002c. Land and resource management plan, Nebraska National Forest and associated units, Rocky Mountain Region.
- USDA Forest Service 2005a. Conata Basin black-footed ferret reintroduction program summary. Unpublished report, Wall Ranger District, Wall, South Dakota.
- USDA Forest Service 2005c. Black-tailed Prairie Dog Conservation & Management on the Nebraska National Forest & Associated Units - Final Environmental Impact Statement . Rocky Mountain Region, Nebraska National Forest. Chadron Nebraska.
- USDA Forest Service 2007. Forest Service Manual. USDA Forest Service, Washington D. C. <http://fsweb.wo.fs.fed.us/directives/html/fsm.shtml> [Accessed March, 2007]
- Uresk, D.W., R. M. King, A. D. Apa, M. S. Deisch, and R. L. Linder. 1985a. Rodenticidal effects of zinc phosphide and strychnine on nontarget species. Final Report Rocky Mountain Range and Forest Experiment Station. Forest Research Laboratory, South Dakota School of Mines and Technology, Rapid City.

- Uresk, D.W., R. L. Linder, and A. D. Apa 1985b. Efficiency of two black-tailed prairie dog rodenticides and their impacts on non-target bird species. Final Report, Rocky Mountain Range and Forest Experiment Station. Forest Research Laboratory, South Dakota School of Mines and Technology, Rapid City.
- Uresk, D.W., R. L. Linder, and M. S. Deisch 1986. Evaluation of three rodenticides on nontarget small mammals and invertebrates. Final Report, Rocky Mountain Range and Forest Experiment Station. Forest Research Laboratory, South Dakota School of Mines and Technology, Rapid City, South Dakota.
- Uresk, D.W., and G.L. Schenbeck. 1987. Effect of zinc phosphide rodenticide on prairie dog colony expansion as determined from aerial photography. *Prairie Naturalist* 19:57-61.
- Uresk, D.W., R. M. King, A. D. Apa, M. S. Deisch, and R. L. Linder. 1988. Rodenticidal effects of zinc phosphide and strychnine on nontarget species. Pages 57-63 In D. W. Uresk, G.L. Schenbeck, and R. Cefkin (technical editors) Eighth Great Plains wildlife damage control workshop proceedings. General technical Report RM-154, Rocky Mountain Forest and Range Experiment Station, USDA Forest Service, Fort Collins, Colorado.
- Vanderhuy, A.V.R. 1985. Interspecific nutritional facilitation: do bison benefit from feeding on prairie dog towns? MS Thesis, Colorado State University, Fort Collins, Colorado.
- Van Pelt, W.E. 1999. The black-tailed prairie dog conservation assessment and strategy-final draft. Nongame and Endangered Wildlife Program. Arizona Game and Fish Department, Phoenix, Arizona.
- Vermeire, L.T., R.K. Heitschmidt, P.S. Johnson, and B.F. Sowell. 2004. The prairie dog story: do we have it right? *BioScience* 54:689-695.
- Virchow, D. R, and S. E. Hygnstrom. 2002. Distribution and abundance of black-tailed prairie dogs in the Great Plains: a historical perspective. *Great Plain Research* 12:197-218.
- Von Loh, J., D. Cogan, D. Faber-Langendoen, D. Crawford, and M. Pucherelli. 1999. USGS-NPS Vegetation Mapping Program, Badlands National Park, South Dakota (Final Report). Technical Memorandum No. 8260-00-02, U.S. Bureau of Reclamation Technical Service Center. Denver, Colorado.
- Vosburgh, T.C. and L.R. Irby. 1998. Effects of recreational shooting on prairie dog colonies. *Journal of Wildlife Management* 62:363-372.
- Walters, C. J. 1986. Adaptive management of renewable resources. Macmillan Co., New York, New York.
- Waples, R.S. 2002. Definition and estimation of effective population size in the conservation of endangered species. Pages 147-168 In S.R. Beissinger and D.R. McCullough, editors, Population viability analysis, University of Chicago Press, Chicago, Illinois.
- Weaver, J.E. and F.W. Albertson. 1956. Grass country of the Great Plains: their nature and use. Johnson Publishing Company. Lincoln, Nebraska.
- White, E.M. 1986. Antiquity, original size and location of prairie dog towns in Wind Cave National Park. Final report for contract CX-1200-4-A040, Wind Cave National Park, Hot Springs, South Dakota.
- Williams, B.K., J.D. Nichols, and M.J. Conroy. 2002. Analysis and management of animal populations. Academic Press, New York.
- Wilson, D.E. and S. Ruff. 1999. The Smithsonian book of North American mammals. Smithsonian Institution Press, Washington, DC.

- Wimsatt, J., D.E. Biggins, E.S. Williams, and V.M. Becerra. 2006. The quest for a safe and effective canine distemper virus vaccine for black-footed ferrets. In Recovery of the Black-footed Ferret: Progress and Continuing Challenges. Edited by J.E. Roelle, B.J. Miller, J.L. Godbey, and D.E. Biggins. U.S. Geological Survey. Pp. 248-266.
- Winter, S. L., J. F. Cully, Jr., and J. S. Pontius. 2002. Vegetation of prairie dog colonies and non-colonized shortgrass prairie. *Journal of Range Management* 55:502-508.
- Wisely, S.M. 2005. The genetic legacy of the black-footed ferret: past, present, and future. Pages 37-43 In J.E. Roelle, B.J. Miller, J.L. Godbey, and D.E. Biggins, editors, Recovery of the Black-footed Ferret: Progress and Continuing Challenges. U.S. Geological Survey.
- With, K.A. 2004. Metapopulation dynamics: perspectives from landscape ecology. Pages 23-44 In I. Hanski and O.E. Gaggiotti, editors, Ecology, genetics, and evolution of metapopulations. Academic Press, New York.
- Wood, J.E. 1965. Response of rodent populations to controls. *Journal of Wildlife Management* 29:425-427.
- Wood, S.L. Editor. 1986. The black-footed ferret. Great Basin Naturalist Memoirs Number 8.
- Wydeven, A. P., and R B. Dahlgren. 1985. Ungulate habitat relationships in Wind Cave National Park. *Journal of Wildlife Management* 49:805-813.
- Wyoming Agricultural Statistics Service. 2001. Black-tailed prairie dog management survey: report of results to Wyoming Game and Fish Department. Wyoming Department of Agriculture, Cheyenne, Wyoming.
- Zinn, H.C. and W.F. Andelt. 1999. Attitudes of Fort Collins, Colorado residents toward prairie dogs. *Wildlife Society Bulletin* 27:1098-1106.